# Section 2 ASSESSMENT OF THE PACIFIC COD STOCK IN THE GULF OF ALASKA 

Grant G. Thompson, Harold H. Zenger, and Martin W. Dorn

U.S. Department of Commerce

National Oceanic and Atmospheric Administration
National Marine Fisheries Service
Alaska Fisheries Science Center
7600 Sand Point Way NE., Seattle, WA 98115-0070

## EXECUTIVE SUMMARY

Summary of Major Changes

Relative to the November edition of last year's GOA SAFE report, the following substantive changes have been made in the Pacific cod stock assessment.

## Changes in the Input Data

1) Size composition data from the 2001 and January-August 2002 commercial fisheries were incorporated into the model.
2) Catches from 2001 were updated and catches through August, 2002 were incorporated.

## Changes in the Assessment Model

Separate selectivity schedules were estimated for the intervals 1978-1986, 1987-1999, and 2000present. In previous assessments, only two intervals were specified: 1978-1986 and 1987-present.

## Changes in Assessment Results

1) The estimated 2003 spawning biomass for the GOA stock is $88,300 \mathrm{t}$, up about $8 \%$ from last year's estimate for 2002 and up about $15 \%$ from last year's $F_{A B C}$ projection for 2003.
2) The estimated 2003 total age 3+ biomass for the GOA stock is $452,000 \mathrm{t}$, up about $6 \%$ from last year's estimate for 2002 and up about $4 \%$ from last year's $F_{40 \%}$ projection for 2003.
3) The recommended 2003 ABC for the GOA stock is $52,800 \mathrm{t}$, down about $8 \%$ from last year's recommendation for 2002 and up about $7 \%$ from last year's $F_{A B C}$ projection for 2003.
4) The estimated 2003 OFL for the GOA stock is $70,100 \mathrm{t}$, down about $9 \%$ from last year's estimate for 2002.

Responses to Comments of the Scientific and Statistical Committee (SSC)

## SSC Comments Specific to the Pacific Cod Assessments

From the December, 2001 minutes: "Current model configurations estimate fishery selectivity in two time stanzas. Given the regulatory changes of the last two years, the SSC recommends that the stock assessment authors evaluate selectivity to determine if additional divisions are appropriate. We also reiterate our call to attempt to calculate a statistically valid spawner-recruit relationship for this stock." An alternative model configuration with a separate set of selectivity schedules for years 2000 and beyond is described in the "Model Structure" subsection of the "Analytic Approach" section and evaluated in the "Model Evaluation" section. As in last year's assessment, a provisional stock-recruitment relationship is described in the "Recruitment" subsection of the "Results" section. Additional research, not described in this assessment, has been conducted in support of a new assessment model capable of calculating a statistically valid spawner-recruit relationship for this stock.

## SSC Comments on Assessments in General

From the December, 2001 minutes: "The SSC encourages the use of retrospective analysis of stock abundance trends, (i.e. the sequential deletion of annual input data to check for changes in output trends.) The presence of a sustained retrospective pattern can be a diagnostic of model adequacy." A retrospective analysis is provided in the "Biomass" subsection of the "Results" section.

## INTRODUCTION

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $63^{\circ} \mathrm{N}$ latitude. Pacific cod is distributed widely over Gulf of Alaska (GOA), as well as the eastern Bering Sea (EBS) and the Aleutian Islands (AI) area. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and GOA, and genetic studies (e.g., Grant et al. 1987) have failed to show significant evidence of stock structure within these areas. Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the GOA.

## FISHERY

During the two decades prior to passage of the Magnuson Fishery Conservation and Management Act (MFCMA) in 1976, the fishery for Pacific cod in the GOA was small, averaging around 3,000 t per year. Most of the catch during this period was taken by the foreign fleet, whose catches of Pacific cod were usually incidental to directed fisheries for other species. By 1976, catches had increased to $6,800 \mathrm{t}$. Catches of Pacific cod since 1978 are shown in Table 2.1, broken down by year, fleet sector, and gear type. The foreign fishery peaked in 1981 at a catch of nearly $35,000 \mathrm{t}$. A small joint venture fishery
existed through 1988, averaging a catch of about $1,400 \mathrm{t}$ per year. The domestic fishery increased steadily through 1986, then increased more than three-fold in 1987 to a catch of nearly $31,000 \mathrm{t}$ as the foreign fishery was eliminated. Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components. Trawl gear has traditionally accounted for the bulk of the catch (nearly two-thirds on average since 1987).

The history of acceptable biological catch (ABC) and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate commercial catches in Table 2.2. For the first year of management under the MFCMA (1977), the catch limit for GOA Pacific cod was established at slightly less than the 1976 total reported landings. During the period 1978-1981, catch limits varied between 34,800 and $70,000 \mathrm{t}$, settling at $60,000 \mathrm{t}$ in 1982. Prior to 1981 these limits were assigned for "fishing years" rather than calendar years. In 1981 the catch limit was raised temporarily to 70,000 t and the fishing year was extended until December 31 to allow for a smooth transition to management based on calendar years, after which the catch limit returned to $60,000 \mathrm{t}$ until 1986, when ABC began to be set on an annual basis. From 1986 (the first year in which an ABC was set) through 2002, TAC averaged about $82 \%$ of ABC and catch averaged about $87 \%$ of TAC. In 8 of these 17 years ( $47 \%$ ), TAC equaled ABC exactly. In 5 of these 17 years ( $29 \%$ ), catch exceeded TAC. However, it should be noted that three of these apparent overages occurred in the most recent five years, when a substantial fishery for Pacific cod was conducted inside State of Alaska waters. To accommodate the State-managed fishery, TAC was set well below ABC in each of those years ( $15 \%$ in 1997 and $1998 ; 20 \%$ in 1999 ; and $23 \%$ in 2000, 2001, and 2002). Thus, the apparent overages in 1999, 2000, and 2002 is basically an artifact of the bijurisdictional nature of the fishery. Catch has exceeded ABC only twice (in 1992 and 1996). Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3) changes in the stock assessment model. For example, from 1986 through 2002, three different assessment models were used (Table 2.2), though the present model has remained unchanged since 1997.

Historically, the majority of the GOA catch has come from the Central regulatory area. To some extent the distribution of effort within the GOA is driven by regulation, as catch limits within this region have been apportioned by area throughout the history of management under the MFCMA. Changes in area-specific allocation between years have usually been traceable to changes in biomass distributions estimated by Alaska Fisheries Science Center trawl surveys or management responses to local concerns. Currently, the ABC allocation follows the average biomass distribution estimated by the three most recent $(1996,1999$, and 2001) trawl surveys, and the TAC allocation is within one percent of this distribution on an area-by-area basis. The complete history of allocation (in percentage terms) by regulatory area within the GOA is shown below:

| Rear(s) |  |  | Regutory Area |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | $\frac{\text { Western }}{28}$ | $\underline{\text { Central }}$ | $\frac{\text { Eastern }}{16}$ |  |  |
| $1977-1985$ | 28 | 16 |  |  |  |
| 1986 | 40 | 44 | 16 |  |  |
| 1987 | 27 | 56 | 17 |  |  |
| $1988-1989$ | 19 | 73 | 8 |  |  |
| 1990 | 33 | 66 | 1 |  |  |
| 1991 | 33 | 62 | 5 |  |  |
| 1992 | 37 | 61 | 2 |  |  |
| $1993-1994$ | 33 | 62 | 5 |  |  |
| $1995-1996$ | 29 | 66 | 5 |  |  |
| $1997-1999$ | 35 | 63 | 2 |  |  |
| $2000-2001$ | 36 | 57 | 7 |  |  |
| 2002 (ABC) | 39 | 55 | 6 |  |  |
| 2002 (TAC) | 38 | 56 | 6 |  |  |

The catches shown in Tables 2.1 and 2.2 include estimated discards. Recent (2000-2001) discard rates of Pacific cod in the various GOA target fisheries are summarized in Table 2.3.

For the 2001 and 2002 fisheries, several new regulations were adopted in an attempt to mitigate possible fishery impacts on the endangered western population of Steller sea lion (Eumetopias jubatus). Some of these regulations were designed to spread the catch of Pacific cod more evenly throughout the year. The table below compares the distribution of catch during the periods January-May ("Per. 1"), June-August ("Per. 2"), and September-December ("Per. 3") for the 2001 fishery with the average for the preceding three years (for each gear type, the numbers in a given row sum to 1.0 ):

|  | Trawl |  |  | Longline |  |  | Pot |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | $\underline{\text { Year(s) }}$ | $\underline{\text { Per. 1 }}$ | Per. 2 | $\underline{\text { Per. 3 }}$ | $\underline{\text { Per. 1 }}$ | $\underline{\text { Per. 2 }}$ | $\underline{\text { Per. 3 }}$ | $\underline{\text { Per. 1 }}$ | Per. 2 | Per. 3 |
| GOA | 2001 | 0.63 | 0.08 | 0.29 | 0.98 | 0.01 | 0.01 | 0.84 | 0.04 | 0.12 |
| GOA | $1998-2000$ | 0.84 | 0.07 | 0.09 | 0.97 | 0.02 | 0.01 | 0.90 | 0.07 | 0.03 |

This section describes data used in the current assessment. It does not attempt to summarize all available data pertaining to Pacific cod in the GOA.

Commercial Catch Data

## Catch Biomass

Catches (including estimated discards) taken in the GOA since 1978 are shown in Table 2.4, broken down by the three main gear types and the following within-year time intervals, or "periods": January-May, June-August, and September-December. This particular division, which was suggested by participants in the BSAI fishery, is intended to reflect actual intra-annual differences in fleet operation (e.g., fishing operations during the spawning period may be different than at other times of year). In years for which estimates of the distribution by gear or period were not available, proxies based on other years' distributions were used.

## Catch Size Composition

Fishery size compositions are presently available, by gear, for the years 1978 through the first part of 2002. As in all assessments since 1997, size composition data from trawl catches sampled on shore were not included in the set of input data, because a comparison of cruises for which both at-sea and shoreside size composition samples were available showed that, in the case of trawl catches, the shoreside data typically contained a smaller proportion of small fish than the at-sea data, indicating that these data may reflect post-discard landings rather than the entire catch. For ease of representation and analysis, length frequency data for Pacific cod can usefully be grouped according to the following set of 25 intervals or "bins," with the upper and lower boundaries shown in cm :

Bin Number: |  | 2 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Lower Bound: 912151821242730333639424550556065707580859095100105
Upper Bound: 1114172023262932353841444954596469747984899499104115
Total length sample sizes for each year, gear, and period are shown in Table 2.5. The collections of relative length frequencies are shown by year, period, and size bin for the pre-1987 trawl fishery in Table 2.6, the pre-1987 longline fishery in Table 2.7, the post-1986 trawl fishery in Table 2.8, the post1986 longline fishery in Table 2.9, and the pot fishery in Table 2.10.

## Survey Data

## Survey Size Composition and Abundance Estimates

The relative size compositions from trawl surveys of the GOA conducted triennially by the Alaska Fisheries Science Center since 1984 are shown in Table 2.11, using the same length bins defined above for the commercial catch size compositions. Total sample sizes are shown below:

| Year: | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Sample size: | 17413 | 19589 | 11440 | 17152 | 12190 | 8645 | 6772 |

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl
surveys are shown in Table 2.12, together with the standard errors and upper and lower $95 \%$ confidence intervals (CI) for the biomass estimates. One potentially problematic aspect of the survey time series is that the 2001 survey did not cover the Eastern regulatory area. The 2001 survey produced a biomass estimate of $258,025 \mathrm{t}$ with a standard error of $52,113 \mathrm{t}$ and a numerical abundance estimate of $157,386,813$ fish. To obtain an estimate of what the 2001 survey would have found had the Eastern regulatory area been surveyed, the biomass trends for the Eastern and combined Western/Central regulatory areas estimated by previous surveys were first compared to determine whether a consistent relationship existed. Finding no such relationship, the 1999 survey results for the Eastern regulatory area were assumed to represent the best point estimates of what the 2001 survey would have observed in that area. The 1999 survey estimates of biomass, biomass variance, and numbers for the Eastern regulatory area were therefore added to the respective 2001survey values for the combined Western/Central regulatory areas. This procedure resulted in a Gulf-wide biomass estimate of $277,743 \mathrm{t}$ with a standard error of $52,355 \mathrm{t}$ and a Gulf-wide numerical abundance estimate of $167,386,950$ fish.

The highest biomass ever observed by the survey was the 1984 estimate of $571,188 \mathrm{t}$, and the low point is the 2001 estimate of $277,743 \mathrm{t}$. In terms of numbers (as opposed to biomass), the record high was observed in 1996, when the population was estimated to include over 315 million fish. This estimate was more than $90 \%$ higher than the previous survey's estimate of 165 million fish, which was the low point in the time series. The 2001 estimate is only about $2 \%$ above the all-time low.

The 1999 trawl survey biomass estimate was distributed by regulatory area as follows: Western $-36 \%$, Central- $57 \%$, and Eastern-7\%. The 2001 trawl survey of the Western and Central regulatory areas estimated $51 \%$ of the biomass to be in the Western regulatory area and $49 \%$ to be in the Central regulatory area. If the procedure described above for extrapolating the actual 2001 survey biomass estimate into a Gulf-wide equivalent is accepted, the implied distribution of biomass by regulatory area is as follows: Western- $47 \%$, Central- $45 \%$, and Eastern- $8 \%$.

## Survey Removals

The amount of Pacific cod removed from the population as a result of NMFS hydroacoustic, longline, and bottom trawl survey operations is summarized for the GOA in Table 2.13. In all years, the magnitude of these removals has been negligible in comparison to the commercial catch (the average ratio of survey removals to commercial removals in the GOA over the period 1978-2002 was approximately 0.001 ).

Length at Age, Weight at Length, and Maturity at Length

The set of reliable length at age data for GOA Pacific cod has been small for the past several years and such data are used only sparingly in this assessment. The otoliths which have been read provide the following data regarding the relationship between age and length and the amount of spread around that relationship (lengths are in cm and ages are back-dated to January 1):

| Age group: | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Average length: | 45 | 52 | 60 | 66 | 74 | 81 | 85 | 90 | 94 | 95 |
| St. dev. of length: | 2.6 | 3.5 | 3.8 | 4.0 | 3.9 | 5.0 | 6.2 | 6.9 | 5.5 | 7.0 |

Although the supply of reliable length at age data has been severely limited in the past, it now appears likely that such data will become much more available in the future. Studies at the Alaska Fisheries

Science Center have resulted in an ageing methodology for Pacific cod that gives reliable age determinations (Roberson 2001), and production ageing of this species has recently begun (Delsa Anderl, pers. commun.).

Weight measurements taken during summer bottom trawl surveys since 1987 yield the following data regarding average weights (in kg ) at length, grouped according to size composition bin (as defined under "Catch Size Composition" above):

Bin Number: |  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |


In 1993, a sampling program was initiated to collect Pacific cod maturity information, using commercial fishery observers. So far, data have been analyzed for 1994 only. These data consist of observers' visual determinations regarding the spawning condition of 2312 females taken in the EBS fishery, which are used as proxy data for the GOA stock. Of these 2312 females, 231 were smaller than 42 cm (the lower boundary of length bin 12). None of these sub- 42 cm fish were mature. The observed proportions of mature fish in the remaining length bins, together with the numbers of fish sampled in those length bins, are shown below (bins are defined under "Catch Size Composition" above):

| Bin number: | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Prop. mature: | 0.03 | 0.05 | 0.14 | 0.19 | 0.28 | 0.53 | 0.69 | 0.82 | 0.89 | 0.94 | 0.94 | 0.91 | 0.89 | 1.00 |
| Sample size: | 39 | 122 | 226 | 313 | 295 | 300 | 320 | 177 | 103 | 70 | 50 | 35 | 19 | 12 |

## ANALYTIC APPROACH

Model Structure

This year's base model structure is identical to the base model structure used in all assessments of the GOA Pacific cod stock since 1997 (Thompson et al. 1997). Beginning with the 1994 SAFE report (Thompson and Zenger 1994), a length-structured Synthesis model (Methot 1986, 1989, 1990, 1998) has formed the primary analytical tool used to assess the GOA Pacific cod stock. Synthesis is a program that uses the parameters of a set of equations governing the assumed dynamics of the stock (the "model parameters") as surrogates for the parameters of statistical distributions from which the data are assumed to be drawn (the "distribution parameters"), and varies the model parameters systematically in the direction of increasing likelihood until a maximum is reached. The overall likelihood is the product of the likelihoods for each of the model components. Each likelihood component is associated with a set of data assumed to be drawn from statistical distributions of the same general form (e.g., multinomial, lognormal, etc.). Typically, likelihood components are associated with data sets such as catch size (or age) composition, survey size (or age) composition, and survey biomass.

The Synthesis program permits each data time series to be divided into multiple segments, or "eras," resulting in a separate set of parameter estimates for each era. To account for possible differences in selectivity between the mostly foreign (also joint venture) and mostly domestic fisheries, the fishery size composition time series in the base model has traditionally been split into pre-1987 and post-1986 eras. A minor modification of the base model was suggested by the SSC at its December, 2001 meeting,
namely, that consideration be given to dividing the domestic era into pre-2000 and post-1999 segments. Two models are therefore considered in this year's assessment: Model 1 is the base model, unchanged since 1997. Model 2 is identical to Model 1, except that an additional set of fishery selectivity schedules was estimated for the post-1999 era.

Symbols used in the stock assessment model are listed in Table 2.14 (note that this list applies to the stock assessment model only, and does not include all symbols used elsewhere in this document). Synthesis uses a total of 16 dimensional constants, special values of indices, and special values of continuous variables, all of which are listed on the first page of Table 2.14. The values of these quantities are not estimated statistically, in the strict sense, but are typically set by assumption or as a matter of structural specification. The values of these constants, indices, and variables are listed in Table 2.15, with a brief rationale given for each value used. In contrast to the quantities whose values are specified in Table 2.15, Synthesis uses a large number of parameters that are estimated statistically (though the estimation itself may not necessarily take place within Synthesis). For ease of reference, capital Roman letters are used to designate such "Synthesis parameters," which are listed on the second page of Table 2.14.

Functional representations of population dynamics are given in Appendix 2A, using the symbols defined in Table 2.14. It should be noted that, while the equations given in Appendix 2A are generally similar to those used in Synthesis, they may differ in detail. Also, only a subset of the equations actually used by Synthesis is shown. Basically, enough equations are shown to illustrate at least one use for each of the symbols shown in Table 2.14.

The assessments conducted during the period 1997-1999 (Thompson et al. 1997, Thompson et al. 1998, Thompson et al. 1999) used approximate Bayesian methods to address uncertainty surrounding the true values of two key model parameters, the natural mortality rate $M$ and the survey catchability coefficient $Q$. Due to limitations of the Synthesis software, a type of meta-analysis was used to implement the Bayesian portion of those assessments. This meta-analysis involved fitting a pair of bivariate distributions to the log-likelihood maxima and projected $F_{40 \%}$ catches returned from a very large number of individual model runs, each of which held $M$ and $Q$ constant at a unique pair of values. The pairs of $M$ and $Q$ values corresponded to points placed at regularly spaced intervals within a grid spanning the $95 \%$ confidence ellipse of the fitted bivariate log-likelihood surface. The purpose of the Bayesian meta-analysis was to recommend an ABC that accounted for parameter uncertainty in an appropriately risk-averse manner. This was accomplished by setting the recommended ABC equal to the geometric mean of the catch distribution corresponding to the product of the catch profile and the posterior distribution. However, the Bayesian meta-analysis was always extremely labor intensive. In the course of conducting the 2000 stock assessment (Thompson et al. 2000), it therefore seemed prudent to seek an efficient shortcut. Looking back at the results of the 1997-1999 stock assessments, it appeared that the ratio between the recommended $F_{A B C}$ emerging from the Bayesian meta-analysis and the $F_{40 \%}$ estimate emerging from the base model was converging over time. The average value of this ratio over the 19971999 period was 0.86 , with a 1999 value of 0.87 . Interestingly, identical three-year average and 1999 values were obtained in the 1997-1999 assessments of the BSAI Pacific cod stock (Thompson and Dorn 1997, Thompson and Dorn 1998, Thompson and Dorn 1999). Because the 1999 value represented the most recent estimate and was approximately equal to the 1997-1999 average, the 2000 stock assessment multiplied this value ( 0.87 ) by the maximum permissible $F_{A B C}$ to obtain the recommended $F_{A B C}$. The resulting ABC recommendation was accepted by the SSC and the Council. The same procedure was used in the 2001 assessment (Thompson et al. 2001) and is retained in the present assessment as well, thereby eliminating the need to re-perform the Bayesian meta-analysis.

Parameters Estimated Independently

Table 2.16 divides the set of Synthesis parameters into two parts, the first of which lists those parameters that were estimated independently (i.e., outside of Synthesis), and the second of which lists those parameters that were estimated conditionally (i.e., inside of Synthesis). This section describes the estimation of parameters in the first part of Table 2.16.

## Natural Mortality

The natural mortality rate was estimated independently of other parameters at a value of 0.37 . This value was used in the present assessment for the following reasons: 1) it was derived as the maximum likelihood estimate of $M$ in the 1993 BSAI Pacific cod assessment, 2) it has been used to represent $M$ in all BSAI Pacific cod assessments since 1993 and in all GOA Pacific cod assessments except one since 1994,3 ) it was explicitly accepted by the SSC for use as an estimate of $M$ in the GOA Pacific cod assessment (SSC minutes, December, 1994), and 4) it lies well within the range of previously published estimates of $M$ shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

## Trawl Survey Catchability

The trawl survey catchability coefficient was estimated independently of other parameters at a value of 1.0. This value was used in the present assessment mostly because it had been used in all previous assessments. Also, preliminary results of recent experimental work conducted in the EBS by the Alaska Fisheries Science Center's Resource Assessment and Conservation Engineering Division tend to confirm that this is a reasonable value (David Somerton, pers. commun.).

## Weight at Length

Parameters (Table 2.14) governing the relationship between weight and length (Appendix 2A) were estimated by regression from the available data (see "Data" above), giving the following values (weights are in kg , lengths in cm ): $W_{1}=5.80 \times 10^{-6}, W_{2}=3.159$.

## Length at First Age of Survey Observation

Assuming that the first age at which Pacific cod are seen in the trawl survey $\left(\alpha_{1}\right.$, Table 2.14) is
approximately 1.5 years, the length at this age ( $L_{1}$, Table 2.14 ) was estimated to be 23.2 cm by averaging the lengths corresponding to the first mode greater than or equal to 14 cm (bin 2) from each of the five most recent survey size compositions.

## Variability in Length at Age

Parameters (Table 2.14) governing the amount of variability surrounding the length-at-age relationship (Appendix 2A) were estimated by linear regression from the observed standard deviations in the available length-at-age data (see "Data" above), giving the following values (in cm ): $X_{1}=1.8, X_{2}=$ 6.9. Estimation of these two parameters constituted the only use of age data in the present assessment.

## Maturity at Length

Maximum likelihood estimates of the parameters (Table 2.14) governing the female maturity-atlength schedule (Appendix 2A) were obtained using the method described by Prentice (1976), giving the following values: $P_{1}=0.142, P_{2}=67.1 \mathrm{~cm}$. The variance-covariance matrix of the parameter estimates gave a standard deviation of 0.006 for the estimate of $P_{1}$, a standard deviation of 0.39 cm for the estimate of $P_{2}$, and a correlation of -0.154 between the estimates of the two parameters.

## Parameters Estimated Conditionally

Those Synthesis parameters that are estimated internally are listed in the second part of Table 2.16. The estimates of these parameters are conditional on each other, as well as on those listed in the first part of the table and discussed in the preceding section (i.e., those Synthesis parameters that are estimated independently).

## Likelihood Components

As noted in the "Model Structure" section, Synthesis is a likelihood-based framework for parameter estimation which allows several data components to be considered simultaneously. In this assessment, four fishery size composition likelihood components were included: the January-May ("early") trawl fishery, the June-September ("late") trawl fishery, the longline fishery, and the pot fishery. In addition to the fishery size composition components, likelihood components for the size composition and biomass trend from the bottom trawl survey were included in the model.

The Synthesis program allows the modeler to specify "emphasis" factors that determine which components receive the greatest attention during the parameter estimation process. As in previous assessments, all components were given an emphasis of 1.0 in the present assessment.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery, and time period within the year. In the parameter estimation process, Synthesis weights a given size composition observation (i.e., the size frequency distribution observed in a given year, gear/fishery, and period) according to the emphasis associated with the respective likelihood component and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which Synthesis was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to
compensate for contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. As in previous assessments, the present assessment uses a multinomial sample size equal to the square root of the true sample size, rather than the true sample size itself. Given the true sample sizes observed in the present assessment, this procedure tends to give values somewhat below 400 while still providing the Synthesis program with usable information regarding the appropriate effort to devote to fitting individual samples. Multinomial sample sizes derived by this procedure for the commercial fishery size compositions are shown in Table 2.17. In the case of survey size composition data, the square root (SR) assumption was also used, giving the multinomial sample sizes shown below:

| Year: | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| SR(sample size): | 132 | 140 | 107 | 131 | 110 | 93 | 82 |

## Use of Survey Biomass Data in Parameter Estimation

Each year's survey biomass datum is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey biomass in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey biomass datum's standard error to the survey biomass datum itself serves as the distribution's coefficient of variation.

## MODEL EVALUATION

As discussed under "Model Structure" above, two models are focused upon in this assessment: Model 1 is the base model, unchanged since 1997. Model 2 is identical to Model 1, except that an additional set of fishery selectivity schedules was estimated for the post-1999 era, as requested by the SSC.

## Evaluation Criteria

Four criteria will be used to evaluate the models developed in the present assessment: 1) the effective sample sizes of the size composition data, 2) the root mean squared error (RMSE) of the fit to the survey biomass data, 3) Akaike's Information Criterion (AIC), and 4) the overall reasonableness of the results.

## Effective Sample Size

Once maximum likelihood estimates of the model parameters have been obtained, Synthesis computes an "effective" sample size for the size composition data specific to a particular year, gear/fishery, and time period within the year. Roughly, the effective sample size can be interpreted as the multinomial sample size that would typically be required in order to produce the given fit. More precisely, it is the sample size that sets the sum of the marginal variances of the proportions implied by the multinomial distribution equal to the sum of the squared differences between the sample proportions and the estimated proportions (McAllister and Ianelli 1997). As a function of a multinomial random variable, the effective sample size has its own distribution. The harmonic mean of the distribution is
equal to the true sample size in the multinomial distribution. Thus, if the effective sample size is less than the true sample size in the multinomial distribution, it is reasonable to conclude that the fit is not as good as expected. The following table shows the average of the input sample sizes and the average effective sample sizes for each of the size composition components (in each column, the average is computed with respect to all years and periods present in the respective time series):

| Size composition <br> likelihood component | Ave. effective sample size |  | Ave. input sample size | Effective size / input size |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Model 1 | Model 2 |  | Model 1 | Model 2 |
| Early-season trawl fishery | 347 | 353 | 134 | 2.59 | 2.63 |
| Late-season trawl fishery | 75 | 81 | 41 | 1.83 | 1.99 |
| Longline fishery | 228 | 273 | 91 | 2.51 | 3.00 |
| Pot fishery | 315 | 302 | 94 | 3.35 | 3.22 |
| Trawl survey | 128 | 125 | 114 | 1.12 | 1.10 |
| All | 229 | 242 | 88 | 2.59 | 2.74 |

Both models produce average effective samples larger (usually much larger) than the average input values for all size composition components. Model 1 produces a higher ratio than Model 2 for two components (pot fishery and trawl survey), whereas Model 2 produces a higher ratio than Model 1 for the other three components (early-season trawl fishery, late-season trawl fishery, and longline fishery). Model 2 also produces a higher overall ratio (2.74) than Model 1 (2.59).

Observed and estimated size compositions in the January-May fisheries in 2000, 2001, and 2002 are compared for Model 1 in Figures 2.1, 2.2, and 2.3; and for Model 2 in Figures 2.4, 2.5, and 2.6. Observed and estimated size compositions from the three most recent bottom trawl surveys are compared for Model 1 in Figure 2.7 and Model 2 in Figure 2.8.

## Fit to Survey Biomass Data

The root-mean-squared value of the lognormal "sigma" parameter in the survey biomass data is 0.158 . The log-scale RMSEs from both models are slightly greater than this value. The log-scale RMSE from Model 1 is 0.184 , slightly higher than Model 2's value of 0.173 .

## Akaike's Information Criterion

AIC is an information-theoretic measure which can be used as a practical guide to model selection (Burnham and Anderson 1998). For model $i$ with $k$ parameters, the AIC is defined as follows:
$\mathrm{AIC}_{i}=-2 \ln \left(\max \left(\right.\right.$ likelihood $\left.\left._{i}\right)\right)+2 k$.
When comparing two nested models (i.e., where one model contains the other as a special case), the model with the higher AIC is better supported by the data. Thus, when used to compare models, only the difference between AIC values is important. Model 2 has 28 more parameters than Model 1 (four additional sets of fishery-specific selectivity parameters, with seven parameters each), so Model 2 is better supported by the data if, following a bit of algebra, the following relationship holds:
$\ln \left(\max \left(\right.\right.$ likelihood $\left.\left._{2}\right)\right)-\ln \left(\max \left(\right.\right.$ likelihood $\left._{1}\right)>28$,
whereas Model 1 is better supported by the data if the direction of the above inequality is reversed.

As it turns out, the difference between the maximum log likelihood for Model 2 and the maximum $\log$ likelihood for Model 1 in this assessment is 30.55 . Therefore, the comparative AIC values indicate that Model 2 is better supported by the data. However, it should be remembered that previous explorations of alternative models for this stock (e.g., Thompson et al. 1999) have suggested the existence of subtle difficulties in the way the likelihood is specified for these models, in which case AIC may not be completely reliable as an indicator of support.

## Overall Reasonableness of Results

The two models gave virtually identical estimates of length-at-age parameters $K$ and $L_{2}\left(L_{1}\right.$ was estimated independently, and thus did not vary with choice of model), as shown below:

| Parameter | Model 1 | Model 2 |
| :--- | ---: | ---: |
| $K$ | 0.143 | 0.143 |
| $L_{2}$ | 84.8 | 85.0 |

Model-specific estimates of fishing mortality rates $F_{g, y, i}$, recruitments $R_{y}$ and initial numbers at age $N_{a}$, and selectivity parameters $S_{1-7, g, e(y \mid g)}$ are shown in Tables 2.18, 2.19, and 2.20, respectively. Again, estimated parameter values were fairly similar between the two models, although Model 2 tends to estimate slightly lower fishing mortality rates and recruitments toward the end of the respective time series and considerably lower terminal selectivities for the longline and early-season trawl fisheries during the 2000-2002 era.

Model-specific estimates of age $3+$ biomass, spawning biomass, and survey biomass are shown in Table 2.21 and Figure 2.9. Here, too, results between the two models were fairly similar, although Model 2 tends to estimate slightly higher biomasses toward the end of the time series.

Given the degree of similarity between the two models with respect to all of the above measures, it seems fair to conclude that the results from the two models are equally reasonable. It is important to remember that the structural differences between the two models are relatively minor, so it is not surprising that they should give similar results.

## Selection of Final Model

One of the main purposes of stock assessments such as the present one is to provide reference estimates of historic biomass trends, target and limit harvest rates, and biomass projections. It is therefore convenient to choose a single model which can be used to generate a set of such reference estimates. Based on the evaluation criteria described above, the evidence is not completely one-sided. The overall results from the two models appear equally reasonable. Model 2 tended to give better effective sample sizes for most (but not all) of the size composition components (as would be expected given the addition of 28 parameters designed to do exactly that), Model 2 gave a slightly better fit to the survey biomass data, and Model 2 had a higher AIC value than Model 1. On balance, therefore, the evidence supported selection of Model 2 as the final model.

The model estimated length-at-age parameter values of $K=0.143$ and $L_{2}=85.0$. Estimates of fishing mortality rates $F_{g, y, i}$, recruitments $R_{y}$ and initial numbers at age $N_{a}$, and selectivity parameters $S_{1-7, g, e(y \mid g)}$ are shown in Tables $2.18,2.19$, and 2.20 , respectively. In addition, the parameter estimates listed in the section entitled "Parameters Estimated Independently" also pertain.

## Schedules Defined by Final Parameter Estimates

Lengths at age defined by the final parameter estimates are shown below (lengths are in cm and are evaluated at the mid-point of each age group):

| Age group: | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Average length: | 24 | 34 | 43 | 51 | 58 | 64 | 69 | 74 | 78 | 81 | 84 | 89 |

The distribution of lengths at age (measured in mid-year) defined by the final parameter estimates is shown in Table 2.22.

Weights at length and maturity proportions at length defined by the final parameters are shown in Table 2.23, and selectivities at length defined by the final parameter estimates are shown in Table 2.24.

## RESULTS

## Definitions

The biomass estimates presented here will be defined in three ways: 1) age $3+$ biomass, consisting of the biomass of all fish aged three years or greater in January of a given year (vector $b$ in Appendix 2A); 2) spawning biomass, consisting of the biomass of all spawning females in March of a given year (vector $c$ in Appendix 2A); and 3) survey biomass, consisting of the biomass of all fish that the Model estimates should have been observed by the survey in July of a given year (vector $d$ in Appendix 2A). The recruitment estimates presented here will be defined in two ways: 1) as numbers of age 3 fish in January of a given year and 2) as the recruitment parameter $R_{y}$, which represents numbers at age 1 in January of year $y$. The fishing mortality rates presented here will be defined as full-selection, instantaneous fishing mortality rates expressed on a per annum scale.

## Biomass

The model's description of the recent history of the stock is shown in Table 2.25, together with estimates provided in last year's final SAFE report (Thompson et al. 2001). The biomass trends estimated
in the present assessment are also shown in Figure 2.9. The age 3+ biomass trend shows an increase during the early 1980s followed by a period of sustained high abundance throughout the rest of that decade, followed by a steady decline through the present. Roughly paralleling the estimated age $3+$ biomass trend, the model's estimated spawning and survey biomass trends show declines throughout the past decade. The model's estimates of 2002 age $3+$ and spawning biomass are the lowest in their respective time series.

Figure 2.10 compares this year's (Model 2) estimate of the survey biomass time series with those from all other assessments since 1997 (the year in which the base model was standardized). These annual estimates have been remarkably consistent. If each assessment's estimate of the survey biomass time series had been used to predict the next assessment's estimate of the same time series, the $R^{2}$ would have ranged from a low of 0.939 (using the 2001 estimates to predict the 2002 estimates) to a high of 0.997 (using the 1998 estimates to predict the 1999 estimates). There is no obvious time trend in the survey biomass estimates between assessments.

Figure 2.11 compares this year's (Model 2) estimate of the age $3+$ biomass time series with those from all other assessments since 1997. These annual estimates have not been quite as consistent as the model's estimates of survey biomass. If each assessment's estimate of the age $3+$ biomass time series had been used to predict the next assessment's estimate of the same time series, the $R^{2}$ would have ranged as high as 0.977 (using the 1998 estimates to predict the 1999 estimates) and would have exceeded 0.854 in all years but one. The outlier arises when the 2000 estimates are used to predict the 2001 estimates, yielding an $R^{2}$ of -0.102 . As Figure 2.11 shows, the 2001 assessment resulted in an estimated age 3+ biomass time series that was lower than any previous assessment. This year's assessment bridges the gap somewhat, by bringing the age $3+$ biomass estimate back in line with those of the pre-2000 assessments by the end of the time series. To attempt to measure whether there is a consistent trend (i.e., retrospective bias) between assessments, the relative change in each year's age 3+ biomass estimate as assessed between each pair of successive assessments was computed (e.g., the relative change in the estimated value of age $3+$ biomass for 1985 as assessed in, say, the 2000 and 2001 assessments), then the relative changes were averaged for each pair of successive assessments, resulting in the values shown below:

| First assessment year | 1997 | 1998 | 1999 | 2000 | 2001 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Second assessment year | 1998 | 1999 | 2000 | 2001 | 2002 |
| Average relative change in age 3+ biomass | 0.022 | -0.010 | 0.010 | -0.138 | 0.024 |

Given that the sign of the average relative change alternates every year, it is reasonable to conclude that these assessments do not show a retrospective bias.

## Recruitment

## Numbers at Age 3

Traditionally, recruitment strengths for Pacific cod have been assessed at age 3, because this is the approximate age of first significant recruitment to the fishery and because model estimates of relative year class strength tend to stabilize by this age. The model's estimated time series of age 3 recruitments is shown in Table 2.26, together with the estimates provided in last year's final SAFE report (Thompson et al. 2001). The model's recruitment estimates are also plotted in Figure 2.12. The current time series has a mean value of 126 million fish, a coefficient of variation of $36 \%$, and an autocorrelation coefficient of 0.112 .

One possible means of assigning a qualitative ranking to each year class within this time series is as follows: an "above average" year class can be defined as one in which numbers at age 3 are at least $120 \%$ of the mean, an "average" year class can be defined as one in which numbers at age 3 are less than $120 \%$ of the mean but at least $80 \%$ of the mean, and a "below average" year class can be defined as one in which numbers at age 3 are less than $80 \%$ of the mean. These criteria give the following classification of year class strengths:

| Above average: | 1977 | 1979 | 1980 | 1984 | 1987 | 1989 |  |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Average: | 1976 | 1978 | 1981 | 1982 | 1983 | 1985 | 1986 | 1988 | 1990 | 1991 | 1993 | 1994 | 1995 |
| Below average: | 1975 | 1992 | 1996 | 1997 | 1998 | 1999 |  |  |  |  |  |  |  |

With respect to last year's assessment (Thompson et al. 2001), the changes in the above table consist of a downgrade in the relative strength of the 1976 year class from "above average" to "average" (returning it to the category in which it resided prior to last year's assessment), an upgrade in the relative strength of the 1993 and 1994 year classes from "below average" to "average," and the addition of the 1999 year class to the "below average" category. It may be noted that all four cohorts following the 1995 year class are in the "below average" category.

## Numbers at Age 1

The model's estimated time series of age 1 recruitments is shown in Table 2.19. This time series has a mean value of 268 million fish, a coefficient of variation of $34 \%$, and an autocorrelation coefficient of 0.153 . The qualitative rankings of year class strengths at age 1 naturally parallel the rankings at age 3 , except that estimates for the 1975 and 1976 year classes do not exist at age 1, the 2000 year class is added to the time series at age 1, and the ranking of the 1978 year classes changes from "below average" at age 1 to "average" at age 3 (it is very close to the boundary in both cases). The 2000 year class appears to be in the "average" category, but this estimate is based on relatively few data.

The present assessment model is not configured to estimate a stock-recruitment relationship. Estimation of stock-recruitment relationships is a notoriously difficult exercise in the field of stock assessment, because both the stock data and the recruitment data are measured with error and because the errors in the stock-recruitment data are autocorrelated (Walters and Ludwig 1981). Also, if the stock and recruitment data are generated by a model which assumes that no stock-recruitment relationship exists, these data will be biased. Nevertheless, the stock-recruitment relationship is potentially such an important component of stock dynamics that it seems prudent to provide some kind of investigation, albeit provisional, as to its possible shape. In addition, the SSC has requested that the assessment include a stock-recruitment relationship (SSC minutes, December, 2000 and December, 2001). To this end, the following analysis was conducted (use of symbols in this description does not necessarily follow Table 2.14, which pertains to the Synthesis assessment model only):

1) Age 1 recruitment $R$ in year $y+1$ was assumed to be related to spawning biomass $S$ in year $y$ by the Ricker (1954) stock-recruitment relationship subject to lognormal error:

$$
R_{y+1}=S_{y} \exp \left(-\alpha-\beta S_{y}+\varepsilon_{y}\right)
$$

where $\alpha$ and $\beta$ are parameters and the $\varepsilon_{y}$ are drawn from a normal distribution with mean 0 and variance $\sigma^{2}$.
2) The estimates of spawning biomass generated by Synthesis were treated as known constants (i.e., it was assumed that they are measured without error).
3) Parameters were estimated by the method of maximum likelihood.
4) The covariance of the parameter estimates was assumed to equal the inverse of the Hessian matrix.

The point estimates of the parameters were $\alpha=-1.021, \beta=0.003398$, and $\sigma=0.235$. The $95 \%$ confidence interval of the stock-recruitment parameters is shown in the upper panel of Figure 2.13. One of the attractive features of the method described above is that it implies that the stock-recruitment relationship $r(S)=S \exp (-\alpha-\beta S)$ is itself a lognormal random variable with parameters that are functions of stock size. The coefficient of variation for the relationship is minimized at the mean of the stock data. The lower panel of Figure 2.13 shows the data (solid squares), the stock-recruitment relationship defined by the point estimates of the parameters (thick curve), and the $95 \%$ confidence interval around the stockrecruitment relationship (thin curves). This analysis is useful mostly because it indicates a considerable level of uncertainty regarding the shape of the stock-recruitment relationship. Moreover, this description of uncertainty should be regarded as an underestimate because of the problems noted in the paragraph above. The estimates given here are not recommended for use in estimating maximum sustainable yield.

## Exploitation

The model's estimated time series of the ratio between catch and age $3+$ biomass is shown in Table 2.27, together with the estimates provided in last year's final SAFE report (Thompson et al. 2001). The average value of this ratio over the entire time series is about 0.068 . The estimated values meet or exceed the average for every year after 1989 except 1994, whereas the estimated values fall below the average for every year prior to 1990.

## PROJECTIONS AND HARVEST ALTERNATIVES

## Amendment 56 Reference Points

Amendment 56 to the GOA Groundfish Fishery Management Plan defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL $\left(F_{O F L}\right)$, the maximum permissible ABC , and the fishing mortality rate used to set the maximum permissible ABC . The fishing mortality rate used to set ABC $\left(F_{A B C}\right)$ may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the GOA are managed under Tier 3 of Amendment 56. Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

3a) Stock status: $B / B_{40 \%}>1$

$$
\begin{aligned}
& F_{\text {OFL }}=F_{35 \%} \\
& F_{\text {ABC }} \leq F_{40 \%} \\
& \text { 3b) } \\
& \text { Stock status: } 1 / 20<B / B_{40 \%} \leq 1 \\
& \\
& F_{\text {OFL }}=F_{35 \%} \times\left(B / B_{40 \%}-1 / 20\right) \times 20 / 19 \\
& \text { 3c) } \\
& F_{\text {ABC }} \leq F_{40 \%} \times\left(B / B_{40 \%}-1 / 20\right) \times 20 / 19 \\
& \text { Stock status: } B / B_{40 \%} \leq 1 / 20 \\
& \\
& \\
& F_{\text {OFL }}=0 \\
& \\
& F_{\text {ABC }}=0
\end{aligned}
$$

Estimation of the $B_{40 \%}$ reference point used in the above formulae requires an assumption regarding the equilibrium level of recruitment. In this assessment, it is assumed that the equilibrium level of recruitment is equal to the post-1976 average (i.e., the arithmetic mean of all estimated recruitments from year classes spawned in 1977 or later). Other useful biomass reference points which can be calculated using this assumption are $B_{100 \%}$ and $B_{35 \%}$, defined analogously to $B_{40 \%}$. These reference points are estimated as follows:

| Reference point: | $B_{35 \%}$ | $B_{40 \%}$ | $B_{100 \%}$ |
| :--- | :--- | :--- | :--- |
| Spawning biomass: | $79,000 \mathrm{t}$ | $90,300 \mathrm{t}$ | $226,000 \mathrm{t}$ |

(For purposes of comparison, Model 1 estimates $B_{35 \%}$ and $B_{40 \%}$ at values of $76,000 \mathrm{t}$ and $86,900 \mathrm{t}$, respectively.)

For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. In this assessment, total fishing mortality was apportioned between gear types (early trawl, late trawl, longline, and pot) at a ratio of 290:53:173:484. These proportions result in a 2003 catch composition that matches the recent (19992001) average distribution of catches between the trawl and fixed-gear fisheries, between the early and late trawl fisheries, and between the longline and pot fisheries. It should be noted that this apportionment scheme is generally consistent with existing Steller sea lion protection measures. This apportionment results in the following estimates of $F_{35 \%}$ and $F_{40 \%}$ :

| $F_{35 \%}$ | $F_{40 \%}$ |
| :--- | :--- |
| 0.42 | 0.35 |

(For purposes of comparison, Model 1 estimates $F_{35 \%}$ and $F_{40 \%}$ at values of 0.50 and 0.41 , respectively.)

## Specification of OFL and Maximum Permissible ABC

Spawning biomass for 2003 is estimated at a value of $88,300 \mathrm{t}$. This is about $2 \%$ below the $B_{40 \%}$ value of $90,300 t$, thereby placing Pacific cod in sub-tier "b" of Tier 3. Given this, the model estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2003 as follows:

|  | Overfishing Level | Maximum Permissible ABC |
| :--- | :--- | :--- |
| Catch: | $70,100 \mathrm{t}$ | $59,900 \mathrm{t}$ |
| Fishing mortality rate: | 0.41 | 0.34 |

The age $3+$ biomass estimate for 2003 is $452,000 \mathrm{t}$. (For purposes of comparison, Model 1 estimates the

OFL for 2003 at a value of $72,800 t$, the maximum permissible $A B C$ for 2003 at a value of $61,400 t$, the $F_{O F L}$ at a value of 0.46 , the maximum permissible $F_{A B C}$ at a value of 0.38 , and the age $3+$ biomass for 2003 at a value of $481,000 \mathrm{t}$.)

## ABC Recommendation

It is important to remember that the maximum permissible $A B C$ computed under the stock assessment model is only a point estimate, around which there is significant uncertainty. For the past several years, the BSAI and GOA Pacific cod assessments have advocated a harvest strategy that formally addresses some of this uncertainty, namely the uncertainty surrounding parameters $M$ and $Q$ (see "Model Structure" above). For the assessments conducted in 2000 and 2001, the strategy was simplified by assuming that the ratio between the recommended $F_{A B C}$ and $F_{40 \%}$ estimate given in the 1999 assessment ( 0.87 ) was an appropriate factor by which to multiply the current maximum permissible $F_{A B C}$ to obtain a recommended $F_{A B C}$. The same strategy is recommended for setting the 2003 ABC . This strategy results in a recommended 2003 ABC of $52,800 \mathrm{t}$, corresponding to a fishing mortality rate of 0.30 . (For purposes of comparison, Model 1 results in a recommended 2003 ABC of $54,200 \mathrm{t}$, corresponding to a fishing mortality rate of 0.33 .)

## Standard Harvest and Recruitment Scenarios and Projection Methodology

A standard set of projections is required for each stock managed under Tiers 1,2 , or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2002 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2003 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2002. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2003, are as follow (" $m a x F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC , so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this
fraction is equal to the ratio of the $F_{A B C}$ value for 2003 recommended in the assessment to the max $F_{A B C}$ for 2003. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 1997-2001 average $F$, which was 0.30 . (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be 1) above its MSY level in 2003 or 2) above $1 / 2$ of its MSY level in 2003 and above its MSY level in 2013 under this scenario, then the stock is not overfished.)

Scenario 7: In 2003 and 2004, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2015 under this scenario, then the stock is not approaching an overfished condition.)

## Projections and Status Determination

Table 2.28 defines symbols used to describe projections of spawning biomass, fishing mortality rate, and catch corresponding to the seven standard harvest scenarios. These projections are shown in Tables 2.29-35.

Harvest scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock overfished? This depends on the stock's estimated spawning biomass in 2003:
a) If spawning biomass for 2003 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b) If spawning biomass for 2003 is estimated to be above $B_{35 \%}$, the stock is above its MSST.
c) If spawning biomass for 2003 is estimated to be above $1 / 2 B_{35 \%}$, but below $B_{35 \%}$, the stock's status relative to MSST is determined by referring to harvest scenario \#6 (Table 2.34). If the mean spawning biomass for 2013 is below $B_{35 \%}$, the stock is below its MSST. Otherwise, the stock is above its MSST.

Is the stock approaching an overfished condition? This is determined by referring to harvest scenario \#7 (Table 2.35):
a) If the mean spawning biomass for 2005 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b) If the mean spawning biomass for 2005 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c) If the mean spawning biomass for 2005 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2015. If the mean spawning biomass for 2015 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

In the case of GOA Pacific cod, spawning biomass for 2003 is estimated to be above $B_{35 \%}$. Therefore, the stock is above its MSST and is not overfished. Mean spawning biomass for 2005 is above $1 / 2 B_{35 \%}$ and below $B_{35 \%}$, and mean spawning biomass for 2015 is above $B_{35 \%}$. Therefore, the stock is not approaching an overfished condition.

## OTHER CONSIDERATIONS

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the BSAI Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Livingston, ed., 2002). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g., Hare and Mantua 2000). However, if regime shifts did occur in 1989 or 1999, it is not yet clear that they had any impact on Pacific cod.

The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), and Westrheim (1996). In terms of percent occurrence, the most important items in the diet of Pacific cod in the BSAI and GOA are polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, the most important dietary items are euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, the most important dietary items are walleye pollock, fishery offal, and yellowfin sole. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species
which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery operates to some extent in the same geographic areas used by Steller sea lion as foraging grounds (Livingston, ed., 2002).

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the longline fishery for Pacific cod. Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross (Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft . LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed sets was as follows:

| Gear | BS | AI | GOA |
| :--- | ---: | ---: | ---: |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513, 517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort was dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot

Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.

## Area Allocation of Harvests

For the 2002 fishery, allocation of ABC among regulatory areas followed the average biomass distribution estimated by the three most recent $(1996,1999$, and 2001) trawl surveys, giving the following apportionment: Western- $39 \%$, Central- $55 \%$, and Eastern-6\%. Assuming that this apportionment is retained for the 2003 fishery, the recommended ABC would be allocated approximately as follows: Western-20,600 t, Central-29,000 t, and Eastern-3,200 t.

## SUMMARY

The major results of the Pacific cod stock assessment are summarized in Table 2.36.

## REFERENCES

Albers, W. D., and P. J. Anderson. 1985. Diet of Pacific cod, Gadus macrocephalus, and predation on the northern pink shrimp, Pandalus borealis, in Pavlof Bay, Alaska. Fish. Bull., U.S. 83:601610.

Bakkala, R. G., and V. G. Wespestad. 1985. Pacific cod. In R. G. Bakkala and L. L. Low (editors), Condition of groundfish resources of the eastern Bering Sea and Aleutian Islands region in 1984, p. 37-49. U.S. Dep. Commer., NOAA Tech. Memo. NMFS F/NWC-83.

Calkins, D. G. 1998. Prey of Steller sea lions in the Bering Sea. Biosphere Conservation 1:33-44.
Fournier, D. A. 1983. An analysis of the Hecate Strait Pacific cod fishery using an age-structured model incorporating density-dependent effects. Can. J. Fish. Aquat. Sci. 40:1233-1243.
Fournier, D., and C. P. Archibald. 1982. A general theory for analyzing catch at age data. Can. J. Fish. Aquat. Sci. 38:1195-1207.
Grant, W. S., C. I. Zhang, and T. Kobayashi. 1987. Lack of genetic stock discretion in Pacific cod (Gadus macrocephalus). Can. J. Fish. Aquat. Sci. 44:490-498.
Hare, S. R., and N. J. Mantua. 2000. Empirical evidence for North Pacific regime shifts in 1977 and 1989. Progress in Oceanography 47:103-146.

Ketchen, K.S. 1964. Preliminary results of studies on a growth and mortality of Pacific cod (Gadus macrocephalus) in Hecate Strait, British Columbia. J. Fish. Res. Bd. Canada 21:1051-1067.
Livingston, P. A. 1989. Interannual trends in Pacific cod, Gadus macrocephalus, predation on three commercially important crab species in the eastern Bering Sea. Fish. Bull., U.S. 87:807-827.
Livingston, P. A. 1991. Pacific cod. In P. A. Livingston (editor), Groundfish food habits and predation on commercially important prey species in the eastern Bering Sea from 1984 to 1986, p. 31-88. U.S. Dept. Commer., NOAA Tech. Memo. NMFS F/NWC-207.

Livingston, P. A. (editor). 2003. Ecosystem Considerations for 2003. North Pacific Fishery Management Council, 605 West 4th Ave., Suite 306, Anchorage, AK 99501.
Low, L. L. 1974. A study of four major groundfish fisheries of the Bering Sea. Ph.D. Thesis, Univ. Washington, Seattle, WA 240 p.
McAllister, M. K., and J. N. Ianelli. 1997. Bayesian stock assessment using catch-age data and the sampling-importance resampling algorithm. Can. J. Fish. Aquat. Sci. 54:284-300.
Methot, R. D. 1986. Synthetic estimates of historical abundance and mortality for northern anchovy, Engraulis mordax. NMFS, Southwest Fish. Cent., Admin. Rep. LJ 86-29, La Jolla, CA.
Methot, R. D. 1989. Synthetic estimates of historical abundance and mortality for northern anchovy. In E. Edwards and B. Megrey (editors), Mathematical analysis of fish stock dynamics: Reviews and current applications, p. 66-82. Amer. Fish. Soc. Symposium 6.
Methot, R. D. 1990. Synthesis model: An adaptable framework for analysis of diverse stock assessment data. Int. N. Pac. Fish. Comm. Bull. 50:259-277.
Methot, R. D. 1998. Application of stock synthesis to NRC test data sets. In V. R. Restrepo (editor), Analyses of simulated data sets in support of the NRC study on stock assessment methods, p. 5980. U.S. Dep. Commer., NOAA Tech. Memo. NMFS-F/SPO-30.

Pitcher, K. W. 1981. Prey of Steller sea lion, Eumetopias jubatus, in the Gulf of Alaska. Fishery Bulletin 79:467-472.
Prentice, R. L. 1976. A generalization of the probit and logit methods for dose response curves. Biometrics 32:761-768.
Ricker, W. E. 1954. Stock and recruitment. J. Fish. Res. Board Can. 11:559-63.
Roberson, N. E. 2001. Age determination of Alaska Pacific cod (Gadus macrocephalus). MS thesis, University of Washington, Seattle. 44 p.
Shimada, A. M., and D. K. Kimura. 1994. Seasonal movements of Pacific cod (Gadus macrocephalus) in the eastern Bering Sea and adjacent waters based on tag-recapture data. U.S. Natl. Mar. Fish. Serv., Fish. Bull. 92:800-816.
Sinclair, E.S. and T. K. Zeppelin. 2002. Seasonal and spatial differences in diet in the western stock of Steller sea lions (Eumetopias jubatus). Journal of Mammalogy 83(4).
Thompson, G. G., and M. W. Dorn. 1997. Pacific cod. In Plan Team for Groundfish Fisheries of the Bering Sea/Aleutian Islands (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions, p. 121-158. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., and M. W. Dorn. 1998. Pacific cod. In Plan Team for Groundfish Fisheries of the Bering Sea/Aleutian Islands (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions, p. 113-181. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., and M. W. Dorn. 1999. Pacific cod. In Plan Team for Groundfish Fisheries of the Bering Sea/Aleutian Islands (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands regions, p. 151-230. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., and R. D. Methot. 1993. Pacific cod. In Plan Team for Groundfish Fisheries of the Bering Sea/Aleutian Islands (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Bering Sea/Aleutian Islands region as projected for 1994, chapter 2. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.

Thompson, G. G., and A. M. Shimada. 1990. Pacific cod. In L. L. Low and R. E. Narita (editors), Condition of groundfish resources of the eastern Bering Sea-Aleutian Islands region as assessed in 1988, p. 44-66. U.S. Dep. Commer., NOAA Tech. Memo. NMFS F/NWC-178.

Thompson, G. G, and H. H. Zenger. 1993. Pacific cod. In Plan Team for Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska as projected for 1994, chapter 2. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G, and H. H. Zenger. 1994. Pacific cod. In Plan Team for Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska as projected for 1995, chapter 2. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., and H. H. Zenger. 1995. Pacific cod. In Plan Team for the Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska as projected for 1996, chapter 2. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., H. H. Zenger, and M. W. Dorn. 1997. Pacific cod. In Plan Team for Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska, p. 121-163. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., H. H. Zenger, and M. W. Dorn. 1998. Pacific cod. In Plan Team for Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska, p. 91-155. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., H. H. Zenger, and M. W. Dorn. 1999. Pacific cod. In Plan Team for Groundfish Fisheries of the Gulf of Alaska (editor), Stock assessment and fishery evaluation report for the groundfish resources of the Gulf of Alaska, p. 105-184. North Pacific Fishery Management Council, 605 W. 4th Avenue Suite 306, Anchorage, AK 99501.
Thompson, G. G., H. H. Zenger, and M. K. Dorn. 2000. Pacific cod. In Plan Team for the Groundfish Fisheries of the Gulf of Alaska (compiler), Stock Assessment and Fishery Evaluation Report for the Groundfish Resources of the Gulf of Alaska p. 91-110. North Pacific Fishery Management Council, 605 West $4^{\text {th }}$ Ave., Suite 306, Anchorage, AK 99501.
Thompson, G. G., H. H. Zenger, and M. K. Dorn. 2001. Assessment of the Pacific cod stock in the Gulf of Alaska. In Plan Team for the Groundfish Fisheries of the Gulf of Alaska (compiler), Stock Assessment and Fishery Evaluation Report for the Groundfish Resources of the Gulf of Alaska p. 2.1-2.73. North Pacific Fishery Management Council, 605 West $4^{\text {th }}$ Ave., Suite 306, Anchorage, AK 99501.
Walters, C. J., and D. Ludwig. 1981. Effects of measurement errors on the assessment of stockrecruitment relationships. Can. J. Fish. Aquat. Sci. 38:704-710.
Wespestad, V., R. Bakkala, and J. June. 1982. Current abundance of Pacific cod (Gadus macrocephalus) in the eastern Bering Sea and expected abundance in 1982-1986. NOAA Tech. Memo. NMFS F/NWC-25, 26 p.
Westrheim, S. J. 1996. On the Pacific cod (Gadus macrocephalus) in British Columbia waters, and a comparison with Pacific cod elsewhere, and Atlantic cod (G. morhua). Can. Tech. Rep. Fish. Aquat. Sci. 2092. 390 p.

Table 2.1--Summary of catches (t) of Pacific cod by fleet sector and gear type. All catches since 1980 include discards. Jt. Vent. = joint venture. Catches taken in the State-water fishery managed by ADFG are lumped into the "Pot" category. Catches for 2002 are through August.

| Fleet Sector |  |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\underline{\text { Foreign }}$ | $\underline{\text { Jt. Vent. }}$ | $\underline{\text { Domestic }}$ | $\underline{\text { Trawl }}$ | $\underline{\text { Longline }}$ | $\underline{\text { Pot }}$ | $\underline{\text { Other }}$ |  |  |
| 1978 | 11370 | 7 | 813 | 4547 | 6800 | 0 | 843 | 12190 |  |
| 1979 | 13173 | 711 | 1020 | 3629 | 9545 | 0 | 1730 | 14904 |  |
| 1980 | 34245 | 466 | 634 | 6464 | 27780 | 0 | 1101 | 35345 |  |
| 1981 | 34969 | 58 | 1104 | 10484 | 25472 | 0 | 175 | 36131 |  |
| 1982 | 26937 | 193 | 2335 | 6679 | 22667 | 0 | 119 | 29465 |  |
| 1983 | 29777 | 2426 | 4337 | 9512 | 26756 | 0 | 272 | 36540 |  |
| 1984 | 15896 | 4649 | 3353 | 8805 | 14844 | 0 | 249 | 23898 |  |
| 1985 | 9086 | 2266 | 3076 | 4876 | 9411 | 2 | 139 | 14428 |  |
| 1986 | 15211 | 1357 | 8444 | 6850 | 17619 | 141 | 402 | 25012 |  |
| 1987 | 0 | 1978 | 30961 | 22486 | 8261 | 642 | 1550 | 32939 |  |
| 1988 | 0 | 1661 | 32141 | 27145 | 3933 | 1422 | 1302 | 33802 |  |
| 1989 | 0 | 0 | 43293 | 37637 | 3662 | 376 | 1618 | 43293 |  |
| 1990 | 0 | 0 | 72517 | 59188 | 5919 | 5661 | 1749 | 72517 |  |
| 1991 | 0 | 0 | 76328 | 58093 | 7656 | 10464 | 115 | 76328 |  |
| 1992 | 0 | 0 | 80746 | 54593 | 15675 | 10154 | 325 | 80746 |  |
| 1993 | 0 | 0 | 56487 | 37806 | 8962 | 9708 | 11 | 56487 |  |
| 1994 | 0 | 0 | 47484 | 31446 | 6778 | 9160 | 100 | 47484 |  |
| 1995 | 0 | 0 | 68985 | 41875 | 10978 | 16055 | 77 | 68985 |  |
| 1996 | 0 | 0 | 68280 | 45991 | 10196 | 12040 | 53 | 68280 |  |
| 1997 | 0 | 0 | 68474 | 48405 | 10977 | 9065 | 26 | 68474 |  |
| 1998 | 0 | 0 | 62102 | 41569 | 9993 | 10510 | 29 | 62102 |  |
| 1999 | 0 | 0 | 68613 | 37167 | 12362 | 19015 | 70 | 68613 |  |
| 2000 | 0 | 0 | 65905 | 25457 | 11667 | 28728 | 54 | 65905 |  |
| 2001 | 0 | 0 | 51060 | 24382 | 9913 | 16613 | 151 | 51060 |  |
| 2002 | 0 | 0 | 44734 | 18541 | 11623 | 14565 | 4 | 44734 |  |
|  |  |  |  |  |  |  |  |  |  |

Table 2.2--History of Pacific cod ABC, TAC, total catch, and type of stock assessment model used to recommend ABC . ABC was not used in management of GOA groundfish prior to 1986. Catch for 2002 is current through August 30. The values in the column labeled "TAC" correspond to "optimum yield" for the years 1980-1986, "target quota" for the year 1987, and true TAC for the years 1988-2001.

| Year | ABC | TAC | Catch | Stock Assessment Model |
| :---: | :---: | :---: | :---: | :---: |
| 1980 | n/a | 60000 | 35345 | $\mathrm{n} / \mathrm{a}$ |
| 1981 | $\mathrm{n} / \mathrm{a}$ | 70000 | 36131 | $\mathrm{n} / \mathrm{a}$ |
| 1982 | $\mathrm{n} / \mathrm{a}$ | 60000 | 29465 | $\mathrm{n} / \mathrm{a}$ |
| 1983 | $\mathrm{n} / \mathrm{a}$ | 60000 | 36540 | $\mathrm{n} / \mathrm{a}$ |
| 1984 | $\mathrm{n} / \mathrm{a}$ | 60000 | 23898 | $\mathrm{n} / \mathrm{a}$ |
| 1985 | $\mathrm{n} / \mathrm{a}$ | 60000 | 14428 | $\mathrm{n} / \mathrm{a}$ |
| 1986 | 136000 | 75000 | 25012 | survey biomass |
| 1987 | 125000 | 50000 | 32939 | survey biomass |
| 1988 | 99000 | 80000 | 33802 | survey biomass |
| 1989 | 71200 | 71200 | 43293 | stock reduction analysis |
| 1990 | 90000 | 90000 | 72517 | stock reduction analysis |
| 1991 | 77900 | 77900 | 76328 | stock reduction analysis |
| 1992 | 63500 | 63500 | 80746 | stock reduction analysis |
| 1993 | 56700 | 56700 | 56487 | stock reduction analysis |
| 1994 | 50400 | 50400 | 47484 | stock reduction analysis |
| 1995 | 69200 | 69200 | 68985 | length-structured Synthesis model |
| 1996 | 65000 | 65000 | 68280 | length-structured Synthesis model |
| 1997 | 81500 | 69115 | 68474 | length-structured Synthesis model |
| 1998 | 77900 | 66060 | 62102 | length-structured Synthesis model |
| 1999 | 84400 | 67835 | 68613 | length-structured Synthesis model |
| 2000 | 76400 | 58715 | 65905 | length-structured Synthesis model |
| 2001 | 67800 | 52110 | 42022 | length-structured Synthesis model |
| 2002 | 57600 | 44230 | 44734 | length-structured Synthesis model |

Table 2.3---Total catch and discards of Pacific cod in the 2001 and 2002 fisheries. Data for 2002 are through October 5. "Catcher-Proc."="CP"=catcher-processor, "Catch"=total catch ( t ), "Disc."=discards (t), "Rate"=ratio of discards to total catch, "WGOA"=western GOA, "CGOA"=central GOA, "EGOA"=eastern GOA.

|  | Shoreside |  | Mothership |  | Catcher-Proc. |  |  |  | All |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Sector | Catch | Disc. | Catch | Disc. | Catch | Disc. | Catch | Disc. | Rate |
| 2001 | WGOA-Inshore | 8,374 | 105 | 0 | 0 | 4,087 | 97 | 12,461 | 202 | 0.016 |
| 2001 | WGOA-Offshore | 0 | 0 | 0 | 0 | 1,700 | 12 | 1,700 | 12 | 0.007 |
| 2001 | CGOA-Inshore | 24,423 | 1,606 | 0 | 0 | 832 | 42 | 25,255 | 1,648 | 0.065 |
| 2001 | CGOA-Offshore | 0 | 0 | 0 | 0 | 2,065 | 18 | 2,065 | 18 | 0.009 |
| 2001 | EGOA-Inshore | 132 | 24 | 0 | 0 | 0 | 0 | 132 | 24 | 0.182 |
| 2001 | EGOA-Offshore | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | n $/ \mathrm{a}$ |
| 2001 | Total | 32,929 | 1,735 | 0 | 0 | 8,684 | 169 | 41,613 | 1,904 | 0.046 |


|  | Shoreside |  | Mothership |  | Catcher-Proc. |  |  | All |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Sector | Catch | Disc. | Catch | Disc. | Catch | Disc. | Catch | Disc. | Rate |
| 2002 | WGOA-Inshore | 8,457 | 134 | 0 | 0 | 4,159 | 57 | 12,616 | 191 | 0.015 |
| 2002 | WGOA-Offshore | 0 | 0 | 0 | 0 | 1,618 | 81 | 1,618 | 81 | 0.050 |
| 2002 | CGOA-Inshore | 22,222 | 3,175 | 0 | 0 | 352 | 28 | 22,574 | 3,203 | 0.142 |
| 2002 | CGOA-Offshore | 0 | 0 | 0 | 0 | 2,245 | 113 | 2,245 | 113 | 0.050 |
| 2002 | EGOA-Inshore | 53 | 8 | 0 | 0 | 0 | 0 | 53 | 8 | 0.151 |
| 2002 | EGOA-Offshore | 0 | 0 | 0 | 0 | 48 | 0 | 48 | 0 | 0.000 |
| 2002 | Total | 30,732 | 3,317 | 0 | 0 | 8,422 | 279 | 39,154 | 3,596 | 0.092 |

Table 2.4--Catch of Pacific cod by year, gear, and period as used in the stock assessment model. Jig catches have been merged with pot catches for 1997-2002. Catch for 2002 is complete through period 2.

| Year |  | Trawl |  |  | Longline |  |  | Pot |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 |
| 1978 | 0 | 0 | 4547 | 0 | 0 | 6800 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 3629 | 0 | 0 | 9545 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 6464 | 0 | 0 | 27780 | 0 | 0 | 0 |
| 1981 | 387 | 3532 | 6565 | 10504 | 5312 | 9656 | 0 | 0 | 0 |
| 1982 | 1143 | 2041 | 3495 | 9912 | 2890 | 9865 | 0 | 0 | 0 |
| 1983 | 2861 | 2844 | 3807 | 10960 | 4651 | 11145 | 0 | 0 | 0 |
| 1984 | 3429 | 2008 | 3368 | 11840 | 425 | 2579 | 0 | 0 | 0 |
| 1985 | 2427 | 571 | 1878 | 9127 | 6 | 278 | 0 | 0 | 2 |
| 1986 | 2999 | 431 | 3420 | 15922 | 401 | 1296 | 5 | 59 | 77 |
| 1987 | 5377 | 7928 | 9181 | 5343 | 983 | 1935 | 219 | 141 | 282 |
| 1988 | 16021 | 6569 | 4555 | 2979 | 507 | 447 | 1081 | 23 | 318 |
| 1989 | 24614 | 12857 | 166 | 2378 | 356 | 928 | 241 | 103 | 32 |
| 1990 | 43279 | 7514 | 8395 | 5557 | 109 | 253 | 2577 | 1008 | 2076 |
| 1991 | 55977 | 631 | 1484 | 7260 | 325 | 70 | 9627 | 7 | 945 |
| 1992 | 51911 | 1189 | 1494 | 12692 | 750 | 2232 | 9926 | 66 | 487 |
| 1993 | 33632 | 2624 | 1550 | 8474 | 307 | 181 | 9699 | 19 | 1 |
| 1994 | 29152 | 1421 | 873 | 6678 | 48 | 52 | 8760 | 0 | 500 |
| 1995 | 38476 | 802 | 2597 | 10591 | 160 | 227 | 15490 | 50 | 592 |
| 1996 | 41450 | 3048 | 1493 | 9938 | 152 | 105 | 12066 | 27 | 0 |
| 1997 | 40727 | 1638 | 6040 | 10403 | 195 | 379 | 8981 | 3 | 107 |
| 1998 | 34693 | 3678 | 3197 | 9548 | 198 | 247 | 10538 | 1 | 1 |
| 1999 | 30124 | 1501 | 5542 | 11937 | 268 | 157 | 14438 | 3302 | 1344 |
| 2000 | 22133 | 2574 | 750 | 11446 | 114 | 107 | 17142 | 113 | 149 |
| 2001 | 15234 | 2035 | 7113 | 9677 | 96 | 140 | 14169 | 599 | 1997 |
| 2002 | 15836 | 2705 | $\mathrm{n} / \mathrm{a}$ | 11536 | 87 | $\mathrm{n} / \mathrm{a}$ | 14406 | 164 | n/a |

Table 2.5--Pacific cod length sample sizes from the commercial fisheries.

| Year | Trawl Fishery |  |  | Longline Fishery |  |  | Pot Fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1978 | 0 | 0 | 634 | 0 | 0 | 18670 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 14460 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 783 | 0 | 0 | 18671 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 461 | 0 | 0 | 19308 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 1390 | 0 | 0 | 22856 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 2896 | 0 | 0 | 127992 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 1039 | 0 | 0 | 47485 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 10141 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 87304 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 387 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 2432 | 0 | 0 | 0 |
| 1989 | 660 | 0 | 312 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 25396 | 10892 | 12025 | 9925 | 0 | 0 | 2783 | 2920 | 10711 |
| 1991 | 38514 | 0 | 131 | 12551 | 143 | 0 | 49453 | 139 | 0 |
| 1992 | 39683 | 0 | 2255 | 28817 | 577 | 3603 | 37177 | 664 | 5013 |
| 1993 | 26844 | 0 | 0 | 11748 | 0 | 0 | 20866 | 0 | 0 |
| 1994 | 12579 | 0 | 0 | 5201 | 0 | 0 | 16342 | 0 | 217 |
| 1995 | 26039 | 120 | 2402 | 24635 | 0 | 0 | 46625 | 0 | 1233 |
| 1996 | 17858 | 0 | 0 | 14706 | 0 | 0 | 35256 | 432 | 0 |
| 1997 | 22822 | 225 | 3746 | 7239 | 119 | 154 | 26880 | 252 | 1537 |
| 1998 | 52448 | 3465 | 6763 | 7981 | 410 | 148 | 31569 | 291 | 2902 |
| 1999 | 11550 | 232 | 1101 | 9013 | 86 | 396 | 33876 | 3719 | 3656 |
| 2000 | 6951 | 425 | 69 | 11426 | 47 | 20 | 28991 | 902 | 277 |
| 2001 | 6115 | 665 | 4560 | 12642 | 145 | 141 | 23290 | 0 | 3925 |
| 2002 | 6285 | 790 | n/a | 9583 | 67 | n/a | 16430 | 0 | n/a |

Table 2.6-Length frequencies of Pacific cod in the pre-1987 trawl fishery by year, period, and length bin.

| Yr. | Per | 1 | $\underline{2}$ | 3 | 4 | 5 | $\underline{6}$ | $\underline{1}$ | $\underline{8}$ | $\underline{9}$ | $\underline{10}$ | 11 | $\underline{12}$ | $\underline{13}$ | 14 | $\underline{15}$ | 16 | 17 | $\underline{18}$ | $\underline{19}$ | $\underline{20}$ | $\underline{21}$ | $\underline{22}$ | $\underline{23}$ | $\underline{24}$ | $\underline{25}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 3 | 0 | 0 | - | 0 | 0 | 1 | 1 | 5 | 9 | 5 | 4 |  | 40 | 93 | 125 | 106 |  |  |  |  | 3 |  | 0 |  |  |
| 1980 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | ${ }^{6}$ | 60 | 162 | 96 | 71 | 91 | 134 | 93 | 48 | 17 | 3 | 0 |  |  |
| 1981 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 29 | 85 | 148 | 145 | 47 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 1982 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 26 | 39 | 118 | 255 | 280 | 294 | 174 | 111 | 52 | 14 | 15 | 5 | 2 |  |  |
|  | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 11 | 24 | 106 | 332 | 388 | 403 | 439 | 375 | 310 | 252 | 143 | 76 | 23 | 7 |  |  |
| 984 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 7 | 49 | 135 | 265 | 127 | 140 | 122 | 70 | 47 |  |  | 13 |  |  | 4 |  |

Table 2.7-Length frequencies of Pacific cod in the pre-1987 longline fishery by year, period, and length bin.

|  | Per | 1 | $\underline{2}$ | $\underline{3}$ | $\underline{4}$ | $\underline{5}$ | $\underline{6}$ | 7 | $\underline{8}$ | $\underline{9}$ | $\underline{10}$ | $\underline{11}$ | $\underline{12}$ | $\underline{13}$ | $\underline{14}$ | 15 | 16 | $\underline{17}$ | $\underline{18}$ | $\underline{19}$ | $\underline{20}$ | $\underline{21}$ | $\underline{22}$ | $\underline{23}$ | $\underline{24}$ | $\underline{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 38 | 91 | 276 | 1160 | 2235 | 3077 | 4051 | 3359 | 2139 | 1261 | 696 | 224 |  | 6 | 1 |  |
| 1979 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 35 | 113 | 285 | 475 | 1124 | 1327 | 1744 | 2148 | 2534 | 2258 | 1401 | 651 | 271 | 75 | 12 |  |  |
| 1980 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 43 | 256 | 1184 | 3776 | 3199 | 1989 | 1555 | 1854 | 1998 | 1630 | 787 | 276 | 99 | 19 | 2 |  |
| 1981 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 29 | 83 | 263 | 1558 | 4685 | 5824 | 3243 | 1485 | 844 | 570 | 379 | 199 | 101 | 28 |  |  |
| 1982 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 40 | 106 | 280 | 498 | 1945 | 3992 | 5101 | 4586 | 3115 | 1729 | 815 | 351 | 181 | 80 | 26 | 6 |  |
| 1983 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 24 | 164 | 728 | 2661 | 11515 | 21037 | 24663 | 2224 | 17602 | 13130 | 7842 | 3868 | 1638 | 588 | 234 | 63 |  |
| 1984 | 3 | 0 |  | 0 | 0 | 0 | 1 | 1 | 5 | 40 | 135 | 341 | 885 | 4389 | 9372 | 10579 | 7666 | 4722 | 3612 | 2572 | 1666 | 958 | 380 | 134 | 23 |  |
| 1985 | 3 | 0 | 0 | 0 | 0 | 1 | 0 | 8 | 45 | 114 | 206 | 316 | 440 | 1036 | 990 | 1847 | 2170 | 1294 | 626 | 462 | 294 | 186 | 89 | 14 | 3 |  |
| 1986 |  |  |  |  |  |  |  |  |  | 133 |  |  |  | 2963 |  |  |  |  |  |  |  |  |  |  |  |  |


Table 2.9-Length frequencies of Pacific cod in the post-1986 longline fishery by year, period, and length bin.



| Yr. | Per | 1 | $\underline{2}$ | 3 | 4 | $\underline{5}$ | $\underline{6}$ | 7 | $\underline{8}$ | $\underline{9}$ | 10 | 11 | 12 | $\underline{13}$ | 14 | 15 | 16 | $\underline{17}$ | $\underline{18}$ | $\underline{19}$ | $\underline{20}$ | $\underline{21}$ | $\underline{22}$ | $\underline{23}$ | $\underline{24}$ | $\underline{25}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 2 | 174 | 34 | 34 | 121 | 104 | 87 | 104 | 469 | 992 | 1479 | 1653 | 1096 | 1566 | 3046 | 2576 | 1897 | 1131 | 469 | 226 | 69 | 52 | 7 | 17 | 0 |  |
| 1987 | 2 | 450 | 19 | 19 | 39 | 98 | 254 | 490 | 29 | 705 | 666 | 1234 | 1411 | 2822 | 4076 | 3116 | 1724 | 42 | 333 | 333 | 254 | 117 | 39 | 19 | 0 |  |
| 1990 | 2 | 251 | 0 | 11 | 103 | 217 | 137 | 57 | 114 | 240 | 286 | 435 | 549 | 1602 | 1774 | 1969 | 1683 | 973 | 549 | 194 | 160 | 80 | 34 | 11 | 11 |  |
| 1993 | 2 | 0 | 17 | 188 | 325 | 239 | 291 | 205 | 256 | 462 | 548 | 839 | 1318 | 2055 | 2620 | 3134 | 2055 | 1404 | 650 | 274 | 119 | 68 | 34 | 17 | 17 |  |
| 1996 | 2 | 0 | 35 | 232 | 875 | 1191 | 903 | 244 | 84 | 193 | 303 | 446 | 445 | 712 | 1043 | 1389 | 1668 | 1403 | 608 | 228 | 87 | 41 | 30 | 15 | 13 |  |
| 99 | 2 | 1 | 17 | 68 | 154 | 166 | 97 | 75 | 142 | 310 | 352 | 402 | 582 | 1093 | 1142 | 1448 | 1208 | 793 | 416 | 168 | 11 | 0 | 0 | 0 | 0 |  |
|  |  |  | 58 |  | 193 | 233 |  | 228 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 2.12--Biomass, standard error, $95 \%$ confidence interval (CI), and population numbers of Pacific cod estimated by NMFS' triennial bottom trawl survey of the GOA. All figures except population numbers are expressed in metric tons. Population numbers are expressed in terms of individual fish.

| Year | Biomass | Standard Error | Lower 95\% CI | Upper 95\% CI | Numbers |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1984 | 571,188 | 85,600 | 403,412 | 738,964 | $217,187,811$ |
| 1987 | 558,662 | 61,500 | 438,122 | 679,202 | $204,177,687$ |
| 1990 | 379,494 | 53,100 | 275,418 | 483,570 | $196,188,094$ |
| 1993 | 409,848 | 73,431 | 265,923 | 553,773 | $164,652,074$ |
| 1996 | 538,154 | 107,736 | 326,991 | 749,317 | $315,443,816$ |
| 1999 | 306,413 | 38,699 | 230,563 | 382,263 | $166,145,850$ |
| 2001 | 277,743 | 52,355 | 175,127 | 380,359 | $167,386,950$ |

Note: The 2001 survey did not cover the eastern GOA. To account for the missing stations, the 1999 survey estimates of biomass, biomass variance, and numbers for the eastern GOA were added to the respective 2001 values to produce the figures shown in the above table.

Table 2.13-Magnitude of hydroacoustic, longline, and bottom trawl survey removals $(\mathfrak{t})$ in the GOA from 1977 through 2002. Cells with an entry of zero indicate that survey removals amounted to less than 0.5 t , whereas cells with no entry indicate either that there was no survey in that region and year or that no data from any such surveys are available.

| Year | Gulf of Alaska |  |  |  |
| :--- | ---: | ---: | ---: | ---: |
|  | Acoustic | Longline | $\frac{\text { Trawl }}{}$ | Total <br> 1977 |
|  |  | 15 | 15 |  |
| 1978 |  |  | 32 | 32 |
| 1979 |  | 14 | 21 | 35 |
| 1980 |  | 25 | 65 | 90 |
| 1981 | 0 | 23 | 70 | 94 |
| 1982 |  | 20 | 41 | 61 |
| 1983 | 1 | 28 | 22 | 52 |
| 1984 | 0 | 24 | 104 | 128 |
| 1985 | 0 | 16 | 30 | 45 |
| 1986 | 0 | 17 | 194 | 210 |
| 1987 | 0 | 21 | 117 | 138 |
| 1988 | 0 | 66 | 1 | 68 |
| 1989 | 0 | 47 | 6 | 53 |
| 1990 | 0 | 48 | 38 | 87 |
| 1991 | 0 | 51 |  | 51 |
| 1992 | 0 | 68 |  | 68 |
| 1993 | 0 | 60 | 46 | 106 |
| 1994 | 0 | 42 |  | 43 |
| 1995 | 0 | 38 |  | 39 |
| 1996 | 1 | 39 | 35 | 75 |
| 1997 | 0 | 39 |  | 39 |
| 1998 | 0 | 30 |  | 30 |
| 1999 | 0 | 22 | 18 | 40 |
| 2000 | 1 | 15 |  | 16 |
| 2001 | 1 | 10 | 24 | 35 |
| 2002 | 4 | 16 |  | 20 |

Table 2.14-Symbols used in the Synthesis assessment model for Pacific cod (page 1 of 2).
Indices

| $a$ | age group |
| :--- | :--- |
| $g$ | gear type |
| $i$ | time interval |
| $j$ | size bin |
| $y$ | year |

Dimensions

| $a_{\text {min }}$ | age of youngest group |
| :--- | :--- |
| $a_{\text {max }}$ | age of oldest group |
| $g_{\text {max }}$ | number of gear types |
| $i_{\max }$ | number of time intervals in each year |
| $j_{\max }$ | number of size bins |
| $y_{\max }$ | number of years |

Special Values of Indices

| $a_{\text {rec }}$ | index of age group used to assess recruitment strength |
| :--- | :--- |
| $\mathrm{g}_{\text {sur }}$ | index of survey gear type |
| $i_{\text {spa }}$ | index of time interval during which spawning occurs |
| $i_{\text {sur }}$ | index of time interval during which survey occurs |

## Operators

$e(y \mid g) \quad$ returns the era containing year $y$ given gear type $g$
$l_{\text {mid }} \quad$ returns the length corresponding to the midpoint of bin $j$
$l_{\text {min }} \quad$ returns the smallest length contained in $\operatorname{bin} j$
$t_{\text {dur }} \quad$ returns the duration (in years) of time interval $i$
Continuous Variables

| $\alpha$ | age |
| :--- | :--- |
| $\lambda$ | length |
| $\tau$ | time |

Special Values of Continuous Variables

| $\alpha_{1}$ | first reference age used in length-at-age relationship (in years) |
| :--- | :--- |
| $\alpha_{2}$ | second reference age used in length-at-age relationship (in years) |
| $\lambda_{\text {min }}$ | minimum length used in assessment |
| $\lambda_{\text {max }}$ | maximum length used in assessment |
| $\tau_{\text {spa }}$ | annual time of spawning (in years) |
| $\tau_{\text {sur }}$ | annual time of survey (in years) |

Table 2.14-Symbols used in the Synthesis assessment model for Pacific cod (page 2 of 2).
Functions of Age or Length
$h(\lambda \mid \alpha) \quad$ probability density function describing distribution of length, conditional on age
$l(\alpha) \quad$ length at age
$p(\lambda) \quad$ proportion mature at length
$s(\lambda \mid g, y) \quad$ selectivity at length, conditional on gear type and year
$w(\lambda) \quad$ weight at length
$x(\alpha) \quad$ standard deviation associated with the length-at-age relationship, as a function of age
Arrays Generated by Synthesis
$b_{y} \quad$ biomass of population aged $a \geq a_{r e c}$ at start of year $y$
$c_{y} \quad$ spawning biomass at time of spawning in year $y$
$d_{y} \quad$ survey biomass at time of survey in year $y$
$n_{a, y, i} \quad$ population numbers at age $a$, year $y$, and time interval $i$
$u_{a, y} \quad$ population numbers at time of spawning at age $a$ and year $y$
$v_{a, y} \quad$ population numbers at time of survey at age $a$ and year $y$
$z_{a, i, j} \quad$ proportion of length distribution falling within size $\operatorname{bin} j$ at age $a$ and time interval $i$
Parameters Used by Synthesis
$F_{g, y, i} \quad$ instantaneous fishing mortality rate at each gear $g$, year $y$, and time $i$ for which catch $>0$
$K \quad$ Brody's growth parameter
$L_{1} \quad$ length at age $\alpha_{1}$
$L_{2} \quad$ length at age $\alpha_{2}$
$M \quad$ instantaneous natural mortality rate
$N_{a} \quad$ initial population numbers at each age $a>a_{\text {min }}$
$P_{1} \quad$ length at point of inflection in maturity schedule
$P_{2} \quad$ relative slope at point of inflection in maturity schedule
$Q \quad$ survey catchability
$R_{y} \quad$ recruitment at age $a_{\text {min }}$ in year $y$
$S_{1, g, e(y \mid g)}$ selectivity at minimum length in gear type $g$ and era $e$
$S_{2, g, e(y \mid g)}$ length at inflection in ascending part of selectivity schedule in gear type $g$ and era $e$
$S_{3, g, e(y \mid g)}$ relative slope at inflection in ascending part of selectivity schedule in gear type $g$ and era $e$
$S_{4, g, e(y \mid g)}$ length at maximum selectivity in gear type $g$ and era $e$
$S_{5, g, e(y \mid g)}$ selectivity at maximum length in gear type $g$ and era $e$
$S_{6, g, e(y \mid g)}$ length at inflection in descending part of selectivity schedule in gear type $g$ and era $e$
$S_{7, g, e(y \mid g)}$ relative slope at inflection in descending part of selectivity schedule in gear type $g$ and era $e$
$W_{1} \quad$ weight-length proportionality
$W_{2} \quad$ weight-length exponent
$X_{1} \quad$ standard deviation of length evaluated at age $\alpha_{1}$
$X_{2} \quad$ standard deviation of length evaluated at age $\alpha_{2}$

Table 2.15-Dimensions and special values of indices and variables used in the Pacific cod assessment. Symbols are defined in Table 2.14.

Dimensions

| $\underline{\text { Term }}$ | Value | Comments/Rationale |
| :--- | ---: | :--- | :--- |
| $a_{\text {min }}$ | 1 | assumed minimum age group observed in the trawl survey |
| $a_{\max }$ | 12 | a convenient place to insert an "age-plus" category |
| $g_{\max }$ | 5 | early trawl, late trawl, longline, pot, survey |
| $i_{\max }$ | 3 | January through March, June through August, September through December |
| $j_{\max }$ | 25 | bin boundaries are given in the "Data" section of the text |
| $y_{\text {max }}$ | 21 | 1978 through 1999 |

Special Values of Indices

| Term | Value | Comments/Rationale <br> $a_{\text {rec }}$ | 3 |
| :--- | ---: | :--- | :--- | | age traditionally used to indicate first significant recruitment to the fishery |
| :--- |
| $\mathrm{g}_{\text {sur }}$ |

Table 2.16-Partitioning the list of parameters used in the Synthesis model of Pacific cod into those that are estimated independently (i.e., outside) of Synthesis and those that are estimated conditionally (i.e., inside of Synthesis).

Parameters Estimated Independently

| $L_{1}$ | length at age $\alpha_{1}$ |
| :--- | :--- |
| $M$ | instantaneous natural mortality rate |
| $P_{1}$ | length at point of inflection in maturity schedule |
| $P_{2}$ | relative slope at point of inflection in maturity schedule |
| $Q$ | survey catchability |
| $W_{1}$ | weight-length proportionality |
| $W_{2}$ | weight-length exponent |
| $X_{1}$ | standard deviation of length evaluated at age $\alpha_{1}$ |
| $X_{2}$ | standard deviation of length evaluated at age $\alpha_{2}$ |
| Parameters Estimated Conditionally |  |
| $F_{g, y, i}$ | instantaneous fishing mortality rate at each gear $g$, year $y$, and time $i$ for which catch $>0$ |
| $K$ | Brody's growth parameter |
| $L_{2}$ | length at age $\alpha_{2}$ |
| $N_{a}$ | initial population numbers at each age $a>a_{m i n}$ |
| $R_{y}$ | recruitment at age $a_{m i n}$ in year $y$ |
| $S_{1, g, e(y \mid g)}$ | selectivity at minimum length in gear type $g$ and era $e$ |
| $S_{2, g, e(y \mid g)}$ | length at inflection in ascending part of selectivity schedule in gear type $g$ and era $e$ |
| $S_{3, g, e(y \mid g)}$ | relative slope at inflection in ascending part of selectivity schedule in gear type $g$ and era $e$ |
| $S_{4, g, e(y \mid g)}$ | length at maximum selectivity in gear type $g$ and era $e$ |
| $S_{5, g, e(y \mid g)}$ | selectivity at maximum length in gear type $g$ and era $e$ |
| $S_{6, g, e(y \mid g)}$ | length at inflection in descending part of selectivity schedule in gear type $g$ and era $e$ |
| $S_{7, g, e(y \mid g)}$ | relative slope at inflection in descending part of selectivity schedule in gear type $g$ and era $e$ |

Table 2.17-Pacific cod commercial fishery length sample sizes used in the multinomial distribution. (These values correspond to the square roots of the true sample sizes shown in Table 2.5.)

| Year | Trawl Fishery |  |  | Longline Fishery |  |  | Pot Fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1978 | 0 | 0 | 25 | 0 | 0 | 137 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 120 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 28 | 0 | 0 | 137 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 21 | 0 | 0 | 139 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 37 | 0 | 0 | 151 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 54 | 0 | 0 | 358 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 32 | 0 | 0 | 218 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 101 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 295 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 20 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 49 | 0 | 0 | 0 |
| 1989 | 26 | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 159 | 104 | 110 | 100 | 0 | 0 | 53 | 54 | 103 |
| 1991 | 196 | 0 | 11 | 112 | 12 | 0 | 222 | 12 | 0 |
| 1992 | 199 | 0 | 47 | 170 | 24 | 60 | 193 | 26 | 71 |
| 1993 | 164 | 0 | 0 | 108 | 0 | 0 | 144 | 0 | 0 |
| 1994 | 112 | 0 | 0 | 72 | 0 | 0 | 128 | 0 | 15 |
| 1995 | 161 | 11 | 49 | 157 | 0 | 0 | 216 | 0 | 35 |
| 1996 | 134 | 0 | 0 | 121 | 0 | 0 | 188 | 21 | 0 |
| 1997 | 151 | 15 | 61 | 85 | 11 | 12 | 164 | 16 | 39 |
| 1998 | 229 | 59 | 82 | 89 | 20 | 12 | 178 | 17 | 54 |
| 1999 | 107 | 15 | 33 | 95 | 9 | 20 | 184 | 61 | 60 |
| 2000 | 83 | 21 | 8 | 107 | 7 | 4 | 170 | 30 | 17 |
| 2001 | 78 | 26 | 68 | 112 | 12 | 12 | 153 | 0 | 63 |
| 2002 | 79 | 28 | n/a | 98 | 8 | n/a | 128 | 0 | n/a |

Table 2.18-Estimates of Pacific cod fishing mortality rates, expressed on an annual time scale (page 1 of 2). Empty cells indicate that no catch was recorded.

Model 1

| Year | Trawl |  |  | Longline |  |  | Pot |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1978 |  |  | 0.03 |  |  | 0.05 |  |  |  |
| 1979 |  |  | 0.02 |  |  | 0.07 |  |  |  |
| 1980 |  |  | 0.04 |  |  | 0.18 |  |  |  |
| 1981 | 0.00 | 0.02 | 0.03 | 0.05 | 0.04 | 0.05 |  |  |  |
| 1982 | 0.01 | 0.01 | 0.02 | 0.04 | 0.02 | 0.05 |  |  |  |
| 1983 | 0.02 | 0.02 | 0.02 | 0.04 | 0.03 | 0.06 |  |  |  |
| 1984 | 0.02 | 0.01 | 0.02 | 0.05 | 0.00 | 0.01 |  |  |  |
| 1985 | 0.01 | 0.00 | 0.01 | 0.04 | 0.00 | 0.00 |  |  |  |
| 1986 | 0.02 | 0.00 | 0.02 | 0.06 | 0.00 | 0.01 |  |  |  |
| 1987 | 0.03 | 0.07 | 0.06 | 0.03 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1988 | 0.08 | 0.06 | 0.03 | 0.02 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| 1989 | 0.13 | 0.12 | 0.00 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1990 | 0.23 | 0.07 | 0.06 | 0.03 | 0.00 | 0.00 | 0.02 | 0.01 | 0.02 |
| 1991 | 0.33 | 0.01 | 0.01 | 0.04 | 0.00 | 0.00 | 0.08 |  | 0.01 |
| 1992 | 0.32 | 0.01 | 0.01 | 0.08 | 0.01 | 0.02 | 0.08 | 0.00 | 0.01 |
| 1993 | 0.22 | 0.03 | 0.01 | 0.05 | 0.00 | 0.00 | 0.08 | 0.00 |  |
| 1994 | 0.19 | 0.01 | 0.01 | 0.04 | 0.00 | 0.00 | 0.07 |  | 0.00 |
| 1995 | 0.25 | 0.01 | 0.02 | 0.07 | 0.00 | 0.00 | 0.13 | 0.00 | 0.01 |
| 1996 | 0.28 | 0.04 | 0.01 | 0.07 | 0.00 | 0.00 | 0.11 | 0.00 |  |
| 1997 | 0.31 | 0.02 | 0.06 | 0.08 | 0.00 | 0.00 | 0.14 | 0.02 | 0.01 |
| 1998 | 0.30 | 0.06 | 0.04 | 0.08 | 0.00 | 0.00 | 0.22 | 0.01 | 0.01 |
| 1999 | 0.29 | 0.03 | 0.07 | 0.12 | 0.00 | 0.00 | 0.34 | 0.09 | 0.05 |
| 2000 | 0.24 | 0.05 | 0.01 | 0.12 | 0.00 | 0.00 | 0.40 | 0.01 | 0.01 |
| 2001 | 0.17 | 0.04 | 0.09 | 0.11 | 0.00 | 0.00 | 0.20 | 0.01 | 0.04 |
| 2002 | 0.18 | 0.05 | 0.06 | 0.13 | 0.00 | 0.00 | 0.20 | 0.00 | 0.03 |

Table 2.18-Estimates of Pacific cod fishing mortality rates, expressed on an annual time scale (page 2 of 2). Empty cells indicate that no catch was recorded.

Model 2

| Year | Trawl |  |  | Longline |  |  | Pot |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1978 |  |  | 0.03 |  |  | 0.05 |  |  |  |
| 1979 |  |  | 0.02 |  |  | 0.07 |  |  |  |
| 1980 |  |  | 0.04 |  |  | 0.18 |  |  |  |
| 1981 | 0.00 | 0.02 | 0.03 | 0.05 | 0.04 | 0.05 |  |  |  |
| 1982 | 0.01 | 0.01 | 0.02 | 0.04 | 0.02 | 0.05 |  |  |  |
| 1983 | 0.02 | 0.02 | 0.02 | 0.04 | 0.03 | 0.06 |  |  |  |
| 1984 | 0.02 | 0.01 | 0.02 | 0.05 | 0.00 | 0.01 |  |  |  |
| 1985 | 0.01 | 0.00 | 0.01 | 0.04 | 0.00 | 0.00 |  |  |  |
| 1986 | 0.02 | 0.00 | 0.02 | 0.06 | 0.00 | 0.01 |  |  |  |
| 1987 | 0.03 | 0.08 | 0.07 | 0.03 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1988 | 0.08 | 0.06 | 0.03 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| 1989 | 0.12 | 0.13 | 0.00 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| 1990 | 0.23 | 0.08 | 0.06 | 0.03 | 0.00 | 0.00 | 0.02 | 0.01 | 0.02 |
| 1991 | 0.32 | 0.01 | 0.01 | 0.04 | 0.00 | 0.00 | 0.07 |  | 0.01 |
| 1992 | 0.31 | 0.01 | 0.01 | 0.07 | 0.01 | 0.02 | 0.08 | 0.00 | 0.00 |
| 1993 | 0.20 | 0.03 | 0.01 | 0.05 | 0.00 | 0.00 | 0.08 | 0.00 |  |
| 1994 | 0.17 | 0.01 | 0.01 | 0.04 | 0.00 | 0.00 | 0.07 |  | 0.00 |
| 1995 | 0.23 | 0.01 | 0.02 | 0.06 | 0.00 | 0.00 | 0.12 | 0.00 | 0.01 |
| 1996 | 0.26 | 0.04 | 0.01 | 0.06 | 0.00 | 0.00 | 0.10 | 0.00 |  |
| 1997 | 0.27 | 0.02 | 0.06 | 0.07 | 0.00 | 0.00 | 0.13 | 0.02 | 0.01 |
| 1998 | 0.26 | 0.05 | 0.03 | 0.07 | 0.00 | 0.00 | 0.19 | 0.01 | 0.01 |
| 1999 | 0.24 | 0.02 | 0.06 | 0.09 | 0.00 | 0.00 | 0.29 | 0.07 | 0.04 |
| 2000 | 0.17 | 0.03 | 0.01 | 0.10 | 0.00 | 0.00 | 0.30 | 0.01 | 0.01 |
| 2001 | 0.13 | 0.03 | 0.07 | 0.09 | 0.00 | 0.00 | 0.16 | 0.01 | 0.03 |
| 2002 | 0.14 | 0.04 | 0.05 | 0.12 | 0.00 | 0.00 | 0.18 | 0.00 | 0.03 |

Table 2.19-Estimates of Pacific cod recruitment at age 1 and initial numbers at age (in millions of fish).
Recruitment at age 1

| Year | Model 1 | Model 2 |
| :--- | ---: | ---: |
| 1978 | 555 | 557 |
| 1979 | 208 | 213 |
| 1980 | 325 | 327 |
| 1981 | 313 | 322 |
| 1982 | 234 | 240 |
| 1983 | 248 | 255 |
| 1984 | 231 | 235 |
| 1985 | 398 | 407 |
| 1986 | 255 | 268 |
| 1987 | 233 | 243 |
| 1988 | 354 | 373 |
| 1989 | 255 | 269 |
| 1990 | 342 | 367 |
| 1991 | 260 | 287 |
| 1992 | 219 | 244 |
| 1993 | 184 | 207 |
| 1994 | 185 | 216 |
| 1995 | 198 | 233 |
| 1996 | 268 | 280 |
| 1997 | 180 | 164 |
| 1998 | 171 | 135 |
| 1999 | 197 | 151 |
| 2000 | 259 | 209 |
| 2001 | 244 | 229 |


| Numbers at age |  |  |
| :---: | ---: | ---: |
| Age | Model 1 | Model 2 |
| 2 | 555 | 557 |
| 3 | 210 | 202 |
| 4 | 54 | 54 |
| 5 | 56 | 55 |
| 6 | 36 | 37 |
| 7 | 79 | 76 |
| 8 | 0 | 0 |
| 9 | 21 | 20 |
| 10 | 13 | 12 |
| 11 | 0 | 0 |
| 12 | 1 | 0 |

Table 2.20-Estimates of Pacific cod selectivity parameters. The first column in each half of the table lists the parameter families for which the remaining columns contain era-specific estimates.

| Model 1 |  |  | Model 2 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl (Jan-May) | 1978-86 | 1987-02 | Trawl (Jan-May) | 1978-86 | 1987-99 | 2000-02 |
| $S_{1, g, e(y \mid g)}$ | n /a | 0.00 | $S_{1, g, e(y \mid g)}$ | n /a | 0.00 | 0.00 |
| $S_{2, g, e(y \mid g)}$ | n/a | 64.11 | $S_{2, g, e(y \mid g)}$ | n/a | 63.94 | 59.34 |
| $S_{3, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 0.19 | $S_{3, g, e(y \mid g)}$ | n/a | 0.19 | 0.19 |
| $S_{4, g, e(y \mid g)}$ | n/a | 114.36 | $S_{4, g, e(y \mid g)}$ | n/a | 114.26 | 101.54 |
| $S_{5, g, e(y \mid g)}$ | n/a | 1.00 | $S_{5, g, e(y \mid g)}$ | n/a | 1.00 | 0.29 |
| $S_{6, g, e(y \mid g)}$ | n/a | 114.42 | $S_{6, g, e(y \mid g)}$ | n/a | 114.33 | 102.55 |
| $S_{7, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 0.15 | $S_{7, g, e(y \mid g)}$ | n/a | 0.17 | 9.35 |
| Trawl (Jun-Dec) | 1978-86 | 1987-02 | Trawl (Jun-Dec) | 1978-86 | 1987-99 | 2000-02 |
| $S_{1, g, e(y \mid g)}$ | 0.00 | 0.00 | $S_{1, g, e(y \mid g)}$ | 0.00 | 0.00 | 0.00 |
| $S_{2, g, e(y \mid g)}$ | 50.47 | 62.69 | $S_{2, g, e(y \mid g)}$ | 50.42 | 64.54 | 52.61 |
| $S_{3, g, e(y \mid g)}$ | 0.35 | 0.17 | $S_{3, g, e(y \mid g)}$ | 0.35 | 0.16 | 0.24 |
| $S_{4, g, e(y \mid g)}$ | 96.13 | 72.92 | $S_{4, g, e(y \mid g)}$ | 96.03 | 73.66 | 69.74 |
| $S_{5, g, e(y \mid g)}$ | 0.84 | 0.43 | $S_{5, g, e(y \mid g)}$ | 0.83 | 0.24 | 0.71 |
| $S_{6, g, e(y \mid g)}$ | 96.13 | 86.34 | $S_{6, g, e(y \mid g)}$ | 96.03 | 84.08 | 84.95 |
| $S_{7, g, e(y \mid g)}$ | 9.59 | 0.33 | $S_{7, g, e(y \mid g)}$ | 9.24 | 0.16 | 0.66 |
| Longline | 1978-86 | 1987-02 | Longline | 1978-86 | 1987-99 | 2000-02 |
| $S_{1, g, e(y \mid g)}$ | 0.00 | 0.00 | $S_{1, g, e(y \mid g)}$ | 0.00 | 0.00 | 0.00 |
| $S_{2, g, e(y \mid g)}$ | 53.66 | 63.15 | $S_{2, g, e(y \mid g)}$ | 53.52 | 62.41 | 62.28 |
| $S_{3, g, e(y \mid g)}$ | 0.33 | 0.25 | $S_{3, g, e(y \mid g)}$ | 0.33 | 0.25 | 0.28 |
| $S_{4, g, e(y \mid g)}$ | 80.34 | 93.48 | $S_{4, g, e(y \mid g)}$ | 80.23 | 93.38 | 92.00 |
| $S_{5, g, e(y \mid g)}$ | 0.93 | 0.77 | $S_{5, g, e(y \mid g)}$ | 0.89 | 0.83 | 0.38 |
| $S_{6, g, e(y \mid g)}$ | 109.11 | 94.41 | $S_{6, g, e(y \mid g)}$ | 109.51 | 94.31 | 92.99 |
| $S_{7, g, e(y \mid g)}$ | 10.00 | 1.90 | $S_{7, g, e(y \mid g)}$ | 10.00 | 1.31 | 1.64 |
| $\underline{\text { Pot }}$ | 1978-86 | 1987-02 | $\underline{\text { Pot }}$ | 1978-86 | 1987-99 | $\underline{2000-02}$ |
| $S_{1, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 0.00 | $S_{1, g, e(y \mid g)}$ | n/a | 0.00 | 0.00 |
| $S_{2, g, e(y \mid g)}$ | n/a | 65.63 | $S_{2, g, e(y \mid g)}$ | n/a | 65.94 | 61.47 |
| $S_{3, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 0.28 | $S_{3, g, e(y \mid g)}$ | n/a | 0.28 | 0.29 |
| $S_{4, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 76.73 | $S_{4, g, e(y \mid g)}$ | n/a | 76.31 | 76.10 |
| $S_{5, g, e(y \mid g)}$ | n/a | 0.23 | $S_{5, g, e(y \mid g)}$ | n/a | 0.21 | 0.19 |
| $S_{6, g, e(y \mid g)}$ | n/a | 76.73 | $S_{6, g, e(y \mid g)}$ | n/a | 76.31 | 76.10 |
| $S_{7, g, e(y \mid g)}$ | $\mathrm{n} / \mathrm{a}$ | 0.14 | $S_{7, g, e(y \mid g)}$ | n/a | 0.13 | 0.23 |
| Survey | 1984-02 |  | Survey | 1984-02 |  |  |
| $S_{1, g, e(y \mid g)}$ | 0.10 |  | $S_{1, g, e(y \mid g)}$ | 0.10 |  |  |
| $S_{2, g, e(y \mid g)}$ | 51.59 |  | $S_{2, g, e(y \mid g)}$ | 51.32 |  |  |
| $S_{3, g, e(y \mid g)}$ | 0.19 |  | $S_{3, g, e(y \mid g)}$ | 0.19 |  |  |
| $S_{4, g, e(y \mid g)}$ | 61.26 |  | $S_{4, g, e(y \mid g)}$ | 61.58 |  |  |
| $S_{5, g, e(y \mid g)}$ | 0.31 |  | $S_{5, g, e(y \mid g)}$ | 0.27 |  |  |
| $S_{6, g, e(y \mid g)}$ | 77.33 |  | $S_{6, g, e(y \mid g)}$ | 76.21 |  |  |
| $S_{7, g, e(y \mid g)}$ | 0.33 |  | $S_{7, g, e(y \mid g)}$ | 0.28 |  |  |

Table 2.21-Time series of GOA Pacific cod age 3+ biomass, spawning biomass, and survey biomass as estimated by Models 1 and 2. All biomass figures are in t .

| Year | Model 1 |  |  | Model 2 | Survey (obs) |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Age 3+ | Spawning Survey (est) | Age 3+ | Spawning Survey (est) |  |
| 1978 | 536 | 107 | 523 | 104 |  |
| 1979 | 579 | 122 | 566 | 119 |  |
| 1980 | 702 | 131 | 692 | 129 |  |
| 1981 | 730 | 136 | 723 | 134 |  |
| 1982 | 766 | 145 | 763 | 144 |  |
| 1983 | 796 | 156 | 798 | 156 |  |
| 1984 | 798 | 167 520 | 804 | 169506 | 571 |
| 1985 | 796 | 175 | 805 | 178 |  |
| 1986 | 792 | 182 | 803 | 185 |  |
| 1987 | 823 | $183 \quad 496$ | 836 | 188 484 | 559 |
| 1988 | 822 | 181 | 838 | 185 |  |
| 1989 | 811 | 180 | 831 | 185 |  |
| 1990 | 816 | 173 488 | 842 | 178 487 | 379 |
| 1991 | 778 | 159 | 810 | 166 |  |
| 1992 | 765 | 148 | 805 | 156 |  |
| 1993 | 733 | $141 \quad 459$ | 782 | $150 \quad 477$ | 410 |
| 1994 | 711 | 141 | 768 | 152 |  |
| 1995 | 681 | 140 | 745 | 153 |  |
| 1996 | 628 | 133 372 | 699 | 148 404 | 538 |
| 1997 | 580 | 122 | 659 | 139 |  |
| 1998 | 552 | 108 | 632 | 126 |  |
| 1999 | 517 | $95 \quad 303$ | 586 | 115 334 | 306 |
| 2000 | 470 | 83 | 519 | 103 |  |
| 2001 | 450 | $80 \quad 291$ | 471 | $99 \quad 279$ | 278 |
| 2002 | 464 | 79 | 454 | 92 |  |

Table 2.22-Distribution of Pacific cod lengths (in cm ) at age (mid-year) as defined by final parameter estimates. Lengths correspond to lower bounds of size bins. Columns sum to 1.0 .

| Len. | Age Group |  |  |  |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | $\underline{1}$ | $\underline{2}$ | $\underline{3}$ | $\underline{4}$ | $\underline{5}$ | $\underline{6}$ | $\underline{7}$ | $\underline{8}$ | $\underline{9}$ | $\underline{10}$ | $\underline{11}$ | $\underline{12+}$ |
| 105 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.023 |
| 100 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.061 |
| 95 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.004 | 0.020 | 0.140 |
| 90 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.008 | 0.041 | 0.107 | 0.221 |
| 85 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.011 | 0.066 | 0.171 | 0.272 | 0.241 |
| 80 | 0 | 0 | 0 | 0 | 0 | 0 | 0.010 | 0.087 | 0.234 | 0.333 | 0.334 | 0.179 |
| 75 | 0 | 0 | 0 | 0 | 0 | 0.006 | 0.095 | 0.284 | 0.364 | 0.299 | 0.199 | 0.092 |
| 70 | 0 | 0 | 0 | 0 | 0.002 | 0.079 | 0.315 | 0.373 | 0.246 | 0.125 | 0.057 | 0.033 |
| 65 | 0 | 0 | 0 | 0 | 0.040 | 0.319 | 0.382 | 0.198 | 0.073 | 0.024 | 0.008 | 0.008 |
| 60 | 0 | 0 | 0 | 0.007 | 0.269 | 0.409 | 0.168 | 0.042 | 0.009 | 0.002 | 0.001 | 0.001 |
| 55 | 0 | 0 | 0 | 0.137 | 0.461 | 0.166 | 0.027 | 0.004 | 0.001 | 0 | 0 | 0 |
| 50 | 0 | 0 | 0.015 | 0.484 | 0.205 | 0.021 | 0.002 | 0 | 0 | 0 | 0 | 0 |
| 45 | 0 | 0 | 0.275 | 0.329 | 0.023 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 | 0 | 0.001 | 0.372 | 0.038 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 39 | 0 | 0.026 | 0.256 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 36 | 0 | 0.207 | 0.074 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 33 | 0 | 0.447 | 0.009 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 30 | 0 | 0.271 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27 | 0.034 | 0.046 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 24 | 0.394 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 21 | 0.499 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 18 | 0.071 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.23-Schedules of Pacific cod weight ( kg ) and maturity proportions at length $(\mathrm{cm})$ as defined by final parameter estimates. Lengths correspond to lower bounds of size bins.

| Bin | Length | Weight | Maturity |
| ---: | ---: | ---: | ---: |
| 1 | 9 | 0.010 | 0 |
| 2 | 12 | 0.022 | 0.001 |
| 3 | 15 | 0.042 | 0.001 |
| 4 | 18 | 0.070 | 0.001 |
| 5 | 21 | 0.110 | 0.002 |
| 6 | 24 | 0.163 | 0.003 |
| 7 | 27 | 0.231 | 0.004 |
| 8 | 30 | 0.316 | 0.006 |
| 9 | 33 | 0.421 | 0.010 |
| 10 | 36 | 0.547 | 0.015 |
| 11 | 39 | 0.697 | 0.023 |
| 12 | 42 | 0.873 | 0.035 |
| 13 | 45 | 1.159 | 0.061 |
| 14 | 50 | 1.588 | 0.117 |
| 15 | 55 | 2.114 | 0.210 |
| 16 | 60 | 2.748 | 0.347 |
| 17 | 65 | 3.501 | 0.514 |
| 18 | 70 | 4.385 | 0.678 |
| 19 | 75 | 5.411 | 0.808 |
| 20 | 80 | 6.590 | 0.894 |
| 21 | 85 | 7.933 | 0.945 |
| 22 | 90 | 9.453 | 0.972 |
| 23 | 95 | 11.160 | 0.986 |
| 24 | 100 | 13.067 | 0.993 |
| 25 | 105 | 14.072 | 0.995 |

Table 2.24-Schedules of Pacific cod selectivities as defined by final parameter estimates. Lengths (cm) correspond to lower bounds of size bins. Trawl(1) = period 1 (January-May) trawl fishery, Trawl(2-3) = periods 2-3 (June-December) trawl fishery.

| Bin | Len. | Trawl (1) |  | Trawl (2-3) |  |  |  | Longline |  |  | Pot |  |  | Survey |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\underline{87-99}$ | $\underline{00-02}$ | $\underline{78-86}$ | $\underline{87-99}$ | $\underline{00-02}$ | $\underline{78-86}$ | $\underline{87-99}$ | $\underline{00-02}$ | $\underline{87-99}$ | $\underline{00-02}$ | $\underline{84-02}$ |  |  |
| 1 | 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 |  |  |
| 2 | 12 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 |  |  |
| 3 | 15 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 |  |  |
| 4 | 18 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 |  |  |
| 5 | 21 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.11 |  |  |
| 6 | 24 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.11 |  |  |
| 7 | 27 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.12 |  |  |
| 8 | 30 | 0.01 | 0.01 | 0.00 | 0.02 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.14 |  |  |
| 9 | 33 | 0.01 | 0.02 | 0.01 | 0.03 | 0.03 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.17 |  |  |
| 10 | 36 | 0.02 | 0.03 | 0.03 | 0.04 | 0.06 | 0.02 | 0.01 | 0.00 | 0.00 | 0.00 | 0.23 |  |  |
| 11 | 39 | 0.03 | 0.06 | 0.09 | 0.06 | 0.12 | 0.04 | 0.01 | 0.01 | 0.00 | 0.01 | 0.30 |  |  |
| 12 | 42 | 0.05 | 0.10 | 0.22 | 0.10 | 0.22 | 0.11 | 0.03 | 0.02 | 0.01 | 0.02 | 0.42 |  |  |
| 13 | 45 | 0.09 | 0.16 | 0.45 | 0.15 | 0.36 | 0.25 | 0.06 | 0.04 | 0.02 | 0.05 | 0.56 |  |  |
| 14 | 50 | 0.19 | 0.34 | 0.83 | 0.28 | 0.66 | 0.64 | 0.18 | 0.16 | 0.07 | 0.18 | 0.80 |  |  |
| 15 | 55 | 0.38 | 0.57 | 0.96 | 0.49 | 0.88 | 0.90 | 0.42 | 0.43 | 0.24 | 0.49 | 0.99 |  |  |
| 16 | 60 | 0.61 | 0.78 | 0.99 | 0.74 | 0.98 | 0.98 | 0.72 | 0.75 | 0.59 | 0.81 | 0.95 |  |  |
| 17 | 65 | 0.80 | 0.90 | 1.00 | 0.96 | 1.00 | 1.00 | 0.90 | 0.93 | 0.89 | 0.96 | 0.81 |  |  |
| 18 | 70 | 0.91 | 0.96 | 1.00 | 0.89 | 0.99 | 1.00 | 0.97 | 0.98 | 0.93 | 0.86 | 0.56 |  |  |
| 19 | 75 | 0.96 | 0.99 | 1.00 | 0.72 | 0.91 | 1.00 | 0.99 | 1.00 | 0.69 | 0.50 | 0.37 |  |  |
| 20 | 80 | 0.99 | 1.00 | 1.00 | 0.54 | 0.73 | 1.00 | 1.00 | 1.00 | 0.50 | 0.30 | 0.30 |  |  |
| 21 | 85 | 0.99 | 1.00 | 1.00 | 0.40 | 0.71 | 1.00 | 0.91 | 0.42 | 0.37 | 0.23 | 0.27 |  |  |
| 22 | 90 | 1.00 | 1.00 | 0.83 | 0.32 | 0.71 | 1.00 | 0.83 | 0.38 | 0.29 | 0.20 | 0.27 |  |  |
| 23 | 95 | 1.00 | 0.29 | 0.83 | 0.27 | 0.71 | 1.00 | 0.83 | 0.38 | 0.25 | 0.19 | 0.27 |  |  |
| 24 | 100 | 1.00 | 0.29 | 0.83 | 0.25 | 0.71 | 0.89 | 0.83 | 0.38 | 0.22 | 0.19 | 0.27 |  |  |
| 25 | 105 | 1.00 | 0.29 | 0.83 | 0.24 | 0.71 | 0.89 | 0.83 | 0.38 | 0.21 | 0.19 | 0.27 |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 2.25-Time series of Pacific cod age $3+$ biomass, spawning biomass, and survey biomass as estimated in last year's and this year's assessments.

| Year | Age 3+ Biomass |  | Spawning Biomass |  | Survey Biomass |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Last Year | This Year | Last Year | This Year | Last Year | This Year |
| 1978 | 555 | 523 | 111 | 104 |  |  |
| 1979 | 598 | 566 | 126 | 119 |  |  |
| 1980 | 721 | 692 | 135 | 129 |  |  |
| 1981 | 746 | 723 | 140 | 134 |  |  |
| 1982 | 781 | 763 | 148 | 144 |  |  |
| 1983 | 808 | 798 | 158 | 156 |  |  |
| 1984 | 808 | 804 | 170 | 169 | 526 | 506 |
| 1985 | 802 | 805 | 177 | 178 |  |  |
| 1986 | 797 | 803 | 183 | 185 |  |  |
| 1987 | 825 | 836 | 184 | 188 | 498 | 484 |
| 1988 | 822 | 838 | 180 | 185 |  |  |
| 1989 | 811 | 831 | 179 | 185 |  |  |
| 1990 | 816 | 842 | 171 | 178 | 490 | 487 |
| 1991 | 779 | 810 | 158 | 166 |  |  |
| 1992 | 764 | 805 | 147 | 156 |  |  |
| 1993 | 731 | 782 | 140 | 150 | 459 | 477 |
| 1994 | 708 | 768 | 140 | 152 |  |  |
| 1995 | 678 | 745 | 138 | 153 |  |  |
| 1996 | 625 | 699 | 131 | 148 | 371 | 404 |
| 1997 | 578 | 659 | 120 | 139 |  |  |
| 1998 | 553 | 632 | 106 | 126 |  |  |
| 1999 | 520 | 586 | 94 | 115 | 306 | 334 |
| 2000 | 471 | 519 | 83 | 103 |  |  |
| 2001 | 441 | 471 | 80 | 99 | 285 | 279 |
| 2002 | $\mathrm{n} / \mathrm{a}$ | 454 | $\mathrm{n} / \mathrm{a}$ | 92 |  |  |

Notes: Spawning biomass is computed as the sum of March female numbers at age times population weight at age times fraction mature at age.
"Survey biomass" is the model's estimate of what the actual survey should have observed.
All biomass figures are in 1000s of t .

Table 2.26-Time series of Pacific cod age 3 recruitment as estimated in last year's and this year's assessments.

| Year | Recruitment (millions of age 3 fish) |  |
| :---: | ---: | ---: |
|  | Last Year | This Year |
| 1978 | 55 | 54 |
| 1979 | 150 | 140 |
| 1980 | 268 | 265 |
| 1981 | 100 | 101 |
| 1982 | 157 | 156 |
| 1983 | 149 | 153 |
| 1984 | 112 | 114 |
| 1985 | 118 | 120 |
| 1986 | 111 | 111 |
| 1987 | 188 | 192 |
| 1988 | 122 | 127 |
| 1989 | 114 | 116 |
| 1990 | 171 | 178 |
| 1991 | 122 | 128 |
| 1992 | 162 | 175 |
| 1993 | 124 | 137 |
| 1994 | 103 | 116 |
| 1995 | 88 | 99 |
| 1996 | 89 | 103 |
| 1997 | 96 | 111 |
| 1998 | 131 | 133 |
| 1999 | 89 | 78 |
| 2000 | 80 | 64 |
| 2001 |  | 72 |

Table 2.27-Time series of Pacific cod catch divided by age 3+ biomass as estimated in last year's and this year's assessments (the entry for 2002 under "This Year" is based on catch through August, 2002; the entry for 2001 under "Last Year" was based on catch through August, 2001).

| Year | Catch Divided by Age 3+ Biomass |  |
| :--- | ---: | ---: |
|  | $\underline{\text { Last Year }}$ | $\underline{\text { This Year }}$ |
| 1978 | 0.02 | 0.02 |
| 1979 | 0.02 | 0.03 |
| 1980 | 0.05 | 0.05 |
| 1981 | 0.05 | 0.05 |
| 1982 | 0.04 | 0.04 |
| 1983 | 0.05 | 0.05 |
| 1984 | 0.03 | 0.03 |
| 1985 | 0.02 | 0.02 |
| 1986 | 0.03 | 0.03 |
| 1987 | 0.04 | 0.04 |
| 1988 | 0.04 | 0.04 |
| 1989 | 0.05 | 0.05 |
| 1990 | 0.09 | 0.09 |
| 1991 | 0.10 | 0.09 |
| 1992 | 0.11 | 0.10 |
| 1993 | 0.08 | 0.07 |
| 1994 | 0.07 | 0.06 |
| 1995 | 0.10 | 0.09 |
| 1996 | 0.11 | 0.10 |
| 1997 | 0.12 | 0.10 |
| 1998 | 0.11 | 0.10 |
| 1999 | 0.13 | 0.12 |
| 2000 | 0.14 | 0.13 |
| 2001 | 0.10 | 0.11 |
| 2002 | $\mathrm{n} / \mathrm{a}$ | 0.10 |

Table 2.28-Definitions of symbols and terms used in the Pacific cod projection tables.
Symbol Definition
SPR Equilibrium spawning per recruit, expressed as a percentage of the maximum level
L90\%CI Lower bound of the $90 \%$ confidence interval
Median Point that divides projection outputs into two groups of equal size ( $50 \%$ higher, $50 \%$ lower)
Mean Average value of the projection outputs
U90\%CI Upper bound of the $90 \%$ confidence interval
St. Dev. Standard deviation of the projection outputs

Table 2.29-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch (1000s of t ) under the assumption that $F=\max F_{A B C}$ in each year 20032015, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

## Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 88.3 | 88.3 | 88.3 | 88.3 | 0.00 |
| 2004 | 79.4 | 79.5 | 79.5 | 79.6 | 0.06 |
| 2005 | 77.4 | 77.9 | 78.0 | 78.7 | 0.40 |
| 2006 | 78.4 | 80.3 | 80.4 | 83.2 | 1.51 |
| 2007 | 79.4 | 83.7 | 84.2 | 90.2 | 3.54 |
| 2008 | 79.0 | 86.3 | 87.0 | 96.3 | 5.70 |
| 2009 | 78.8 | 87.9 | 89.0 | 101.9 | 7.29 |
| 2010 | 79.2 | 89.2 | 90.2 | 105.0 | 8.20 |
| 2011 | 79.8 | 89.6 | 91.1 | 106.5 | 8.66 |
| 2012 | 79.5 | 90.3 | 91.5 | 107.8 | 8.87 |
| 2013 | 80.0 | 90.1 | 91.7 | 107.7 | 8.85 |
| 2014 | 80.0 | 90.1 | 91.7 | 107.7 | 8.76 |
| 2015 | 80.2 | 90.3 | 91.7 | 107.7 | 8.72 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.34 | 0.34 | 0.34 | 0.34 | 0.000 |
| 2004 | 0.31 | 0.31 | 0.31 | 0.31 | 0.000 |
| 2005 | 0.30 | 0.30 | 0.30 | 0.30 | 0.002 |
| 2006 | 0.30 | 0.31 | 0.31 | 0.32 | 0.006 |
| 2007 | 0.31 | 0.33 | 0.32 | 0.35 | 0.013 |
| 2008 | 0.30 | 0.34 | 0.33 | 0.35 | 0.016 |
| 2009 | 0.30 | 0.35 | 0.34 | 0.35 | 0.016 |
| 2010 | 0.30 | 0.35 | 0.34 | 0.35 | 0.016 |
| 2011 | 0.31 | 0.35 | 0.34 | 0.35 | 0.016 |
| 2012 | 0.31 | 0.35 | 0.34 | 0.35 | 0.016 |
| 2013 | 0.31 | 0.35 | 0.34 | 0.35 | 0.015 |
| 2014 | 0.31 | 0.35 | 0.34 | 0.35 | 0.015 |
| 2015 |  |  | 0.35 | 0.015 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 59.9 | 59.9 | 59.9 | 59.9 | 0.00 |
| 2004 | 50.6 | 50.7 | 50.7 | 50.8 | 0.06 |
| 2005 | 51.4 | 52.0 | 52.0 | 53.0 | 0.50 |
| 2006 | 55.0 | 57.9 | 58.3 | 62.9 | 2.51 |
| 2007 | 56.3 | 64.5 | 65.3 | 77.8 | 6.72 |
| 2008 | 55.0 | 68.8 | 69.7 | 84.1 | 9.50 |
| 2009 | 54.8 | 71.5 | 71.7 | 88.3 | 10.66 |
| 2010 | 55.5 | 73.1 | 72.6 | 90.6 | 11.09 |
| 2011 | 55.4 | 73.1 | 73.1 | 91.2 | 11.22 |
| 2012 | 55.3 | 73.0 | 73.3 | 90.7 | 11.21 |
| 2013 | 56.1 | 73.3 | 73.2 | 91.1 | 10.92 |
| 2014 | 56.0 | 73.3 | 73.2 | 90.5 | 10.75 |
| 2015 | 55.7 |  | 90.9 | 10.78 |  |

Table 2.30-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch ( 1000 s of t ) under the assumption that the ratio of $F$ to $\max F_{A B C}$ in each year 2003-2015 is fixed at a value of 0.87 , where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 88.7 | 88.7 | 88.7 | 88.7 | 0.00 |
| 2004 | 81.8 | 81.9 | 81.9 | 82.0 | 0.06 |
| 2005 | 80.7 | 81.2 | 81.2 | 82.0 | 0.40 |
| 2006 | 82.2 | 84.1 | 84.2 | 87.0 | 1.52 |
| 2007 | 83.5 | 87.9 | 88.4 | 94.7 | 3.63 |
| 2008 | 83.3 | 90.8 | 91.8 | 102.1 | 6.08 |
| 2009 | 83.2 | 93.0 | 94.3 | 108.6 | 8.07 |
| 2010 | 83.8 | 94.9 | 96.2 | 112.5 | 9.29 |
| 2011 | 84.4 | 96.0 | 97.6 | 114.7 | 9.95 |
| 2012 | 84.3 | 97.2 | 98.4 | 116.2 | 10.26 |
| 2013 | 84.9 | 97.8 | 98.8 | 117.0 | 10.30 |
| 2014 | 84.8 | 97.7 | 99.0 | 117.4 | 10.22 |
| 2015 | 85.1 | 97.9 | 99.2 | 117.3 | 10.17 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.30 | 0.30 | 0.30 | 0.30 | 0.000 |
| 2004 | 0.27 | 0.27 | 0.27 | 0.27 | 0.000 |
| 2005 | 0.27 | 0.27 | 0.27 | 0.27 | 0.001 |
| 2006 | 0.27 | 0.29 | 0.28 | 0.29 | 0.005 |
| 2007 | 0.28 | 0.30 | 0.29 | 0.30 | 0.009 |
| 2008 | 0.28 | 0.30 | 0.30 | 0.30 | 0.009 |
| 2009 | 0.28 | 0.30 | 0.30 | 0.30 | 0.009 |
| 2010 | 0.28 | 0.30 | 0.30 | 0.30 | 0.009 |
| 2011 | 0.28 | 0.30 | 0.30 | 0.30 | 0.008 |
| 2012 | 0.28 | 0.30 | 0.30 | 0.30 | 0.008 |
| 2013 | 0.28 | 0.30 | 0.30 | 0.30 | 0.008 |
| 2014 | 0.28 | 0.30 | 0.30 | 0.007 |  |
| 2015 |  |  | 0.30 | 0.007 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 52.8 | 52.8 | 52.8 | 52.8 | 0.00 |
| 2004 | 46.7 | 46.8 | 46.8 | 46.9 | 0.05 |
| 2005 | 48.3 | 48.8 | 48.9 | 49.7 | 0.46 |
| 2006 | 52.1 | 54.8 | 55.1 | 59.3 | 2.29 |
| 2007 | 53.7 | 61.3 | 61.7 | 70.9 | 5.56 |
| 2008 | 52.7 | 65.6 | 65.3 | 77.0 | 8.75 |
| 2009 | 52.6 | 67.1 | 67.2 | 81.4 | 8.95 |
| 2010 | 53.2 | 68.2 | 68.2 | 84.1 | 9.42 |
| 2011 | 53.9 | 68.3 | 68.8 | 84.9 | 9.58 |
| 2012 | 53.5 | 68.9 | 69.0 | 84.5 | 9.59 |
| 2013 | 54.3 | 68.6 | 69.1 | 84.9 | 9.36 |
| 2014 | 54.5 | 68.5 | 69.1 | 84.5 | 9.19 |
| 2015 | 54.4 | 68.6 | 69.2 | 84.4 | 9.14 |

Table 2.31-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch ( 1000 s of t ) under the assumption that $F=1 / 2$ max $F_{A B C}$ in each year 2003-2015, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 90.0 | 90.0 | 90.0 | 90.0 | 0.00 |
| 2004 | 89.2 | 89.3 | 89.3 | 89.4 | 0.06 |
| 2005 | 91.8 | 92.3 | 92.3 | 93.1 | 0.42 |
| 2006 | 96.0 | 97.9 | 98.1 | 101.1 | 1.64 |
| 2007 | 100.0 | 105.0 | 105.5 | 112.5 | 4.10 |
| 2008 | 102.0 | 111.3 | 112.3 | 124.3 | 7.23 |
| 2009 | 103.8 | 116.9 | 118.2 | 135.4 | 9.98 |
| 2010 | 105.5 | 121.6 | 122.9 | 143.3 | 11.87 |
| 2011 | 108.1 | 125.5 | 126.6 | 149.4 | 12.98 |
| 2012 | 109.5 | 128.0 | 129.1 | 152.6 | 13.57 |
| 2013 | 110.8 | 129.8 | 130.9 | 155.8 | 13.77 |
| 2014 | 111.8 | 130.9 | 132.0 | 156.0 | 13.76 |
| 2015 | 112.2 | 131.6 | 132.7 | 156.2 | 13.68 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.17 | 0.17 | 0.17 | 0.17 | 0.000 |
| 2004 | 0.17 | 0.17 | 0.17 | 0.17 | 0.000 |
| 2005 | 0.17 | 0.18 | 0.17 | 0.18 | 0.000 |
| 2006 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2007 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2008 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2009 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2010 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2011 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2012 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2013 | 0.18 | 0.18 | 0.18 | 0.18 | 0.000 |
| 2014 | 0.18 | 0.18 | 0.18 | 0.000 |  |
| 2015 |  |  | 0.18 | 0.000 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 31.6 | 31.6 | 31.6 | 31.6 | 0.00 |
| 2004 | 31.9 | 32.0 | 32.0 | 32.0 | 0.03 |
| 2005 | 35.4 | 35.7 | 35.7 | 36.0 | 0.17 |
| 2006 | 38.2 | 39.2 | 39.3 | 40.9 | 0.84 |
| 2007 | 39.4 | 42.6 | 42.9 | 47.4 | 2.57 |
| 2008 | 39.6 | 45.1 | 45.7 | 4.29 |  |
| 2009 | 39.9 | 46.9 | 47.6 | 56.7 | 5.35 |
| 2010 | 40.5 | 48.3 | 48.8 | 59.9 | 5.85 |
| 2011 | 40.9 | 48.9 | 49.6 | 60.0 | 6.07 |
| 2012 | 41.8 | 49.6 | 50.0 | 60.5 | 6.14 |
| 2013 | 41.7 | 49.7 | 50.3 | 60.8 | 6.05 |
| 2014 | 41.8 | 49.8 | 50.4 | 60.7 | 5.95 |
| 2015 |  | 49.9 | 50.5 | 61.2 | 5.92 |

Table 2.32-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch ( 1000 s of t ) under the assumption that $F=$ the 1997-2001 average in each year 2003-2015, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

## Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 89.2 | 89.2 | 89.2 | 89.2 | 0.00 |
| 2004 | 84.4 | 84.4 | 84.4 | 84.6 | 0.07 |
| 2005 | 83.7 | 84.2 | 84.3 | 85.1 | 0.43 |
| 2006 | 85.5 | 87.5 | 87.7 | 90.7 | 1.64 |
| 2007 | 87.4 | 92.3 | 92.8 | 99.7 | 4.05 |
| 2008 | 87.7 | 96.7 | 97.6 | 109.2 | 6.94 |
| 2009 | 88.3 | 100.6 | 101.7 | 117.9 | 9.30 |
| 2010 | 89.2 | 103.8 | 104.9 | 123.1 | 10.77 |
| 2011 | 90.7 | 106.3 | 107.3 | 127.3 | 11.56 |
| 2012 | 91.8 | 108.1 | 108.9 | 129.2 | 11.93 |
| 2013 | 92.6 | 109.3 | 109.9 | 130.5 | 11.99 |
| 2014 | 93.0 | 109.3 | 110.5 | 131.3 | 11.89 |
| 2015 | 93.1 | 109.7 | 110.9 | 130.6 | 11.80 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2004 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2005 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2006 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2007 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2008 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2009 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2010 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2011 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2012 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2013 | 0.25 | 0.25 | 0.25 | 0.000 |  |
| 2014 | 0.25 | 0.25 | 0.25 | 0.25 | 0.000 |
| 2015 | 0.25 |  |  | 0.000 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 45.0 | 45.0 | 45.0 | 45.0 | 0.00 |
| 2004 | 44.1 | 44.2 | 44.2 | 44.2 | 0.02 |
| 2005 | 46.4 | 46.6 | 46.6 | 47.0 | 0.21 |
| 2006 | 49.1 | 50.6 | 50.7 | 52.9 | 1.20 |
| 2007 | 49.9 | 54.4 | 54.8 | 61.1 | 3.58 |
| 2008 | 49.6 | 57.2 | 57.9 | 67.4 | 5.81 |
| 2009 | 49.6 | 59.0 | 59.9 | 72.1 | 7.08 |
| 2010 | 50.1 | 60.5 | 61.1 | 74.3 | 7.62 |
| 2011 | 50.8 | 61.0 | 61.9 | 75.5 | 7.84 |
| 2012 | 50.7 | 61.8 | 62.3 | 75.5 | 7.89 |
| 2013 | 51.6 | 61.6 | 62.4 | 76.0 | 7.75 |
| 2014 | 51.3 | 61.6 | 62.5 | 75.8 | 7.62 |
| 2015 | 51.6 | 61.8 | 62.6 | 76.2 | 7.58 |

Table 2.33-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch ( 1000 s of t ) under the assumption that $F=0$ in each year 2003-2015, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 91.8 | 91.8 | 91.8 | 91.8 | 0.00 |
| 2004 | 100.9 | 101.0 | 101.0 | 101.1 | 0.07 |
| 2005 | 112.6 | 113.1 | 113.1 | 113.9 | 0.43 |
| 2006 | 125.9 | 127.9 | 128.1 | 131.1 | 1.66 |
| 2007 | 139.0 | 144.2 | 144.7 | 152.1 | 4.27 |
| 2008 | 149.2 | 159.3 | 160.4 | 174.1 | 7.99 |
| 2009 | 157.7 | 173.0 | 174.7 | 194.8 | 11.90 |
| 2010 | 165.0 | 185.2 | 187.1 | 212.6 | 15.23 |
| 2011 | 171.8 | 195.6 | 197.6 | 228.4 | 17.68 |
| 2012 | 177.9 | 203.9 | 205.6 | 238.2 | 19.32 |
| 2013 | 182.2 | 209.6 | 211.6 | 247.6 | 20.28 |
| 2014 | 185.4 | 214.6 | 215.7 | 253.0 | 20.87 |
| 2015 | 187.2 | 217.3 | 218.6 | 254.8 | 21.07 |


| Fishing Mortality Projections <br> Year <br> L90\%CI |  | Median | Mean | U90\%CI | St. Dev. |
| :---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 0 |
| 200 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 |  |
| 2015 | 0 | 0 | 0 | 0 |  |

Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 |  |

Table 2.34-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t ), fishing mortality, and catch ( 1000 s of t ) under the assumption that $F=F_{\text {OFL }}$ in each year 2003-2015, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 87.6 | 87.6 | 87.6 | 87.6 | 0.00 |
| 2004 | 76.0 | 76.1 | 76.1 | 76.2 | 0.06 |
| 2005 | 73.0 | 73.5 | 73.6 | 74.3 | 0.40 |
| 2006 | 73.5 | 75.3 | 75.5 | 78.2 | 1.49 |
| 2007 | 74.1 | 78.4 | 78.8 | 84.7 | 3.45 |
| 2008 | 73.5 | 80.5 | 81.2 | 89.8 | 5.34 |
| 2009 | 73.2 | 81.9 | 82.6 | 93.5 | 6.46 |
| 2010 | 73.7 | 82.8 | 83.4 | 95.3 | 6.95 |
| 2011 | 73.9 | 83.0 | 83.8 | 96.2 | 7.12 |
| 2012 | 73.8 | 83.3 | 83.9 | 96.5 | 7.16 |
| 2013 | 74.2 | 83.1 | 83.8 | 96.3 | 7.06 |
| 2014 | 74.1 | 82.9 | 83.8 | 95.8 | 6.95 |
| 2015 | 74.0 | 82.9 | 83.8 | 95.9 | 6.90 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.41 | 0.41 | 0.41 | 0.41 | 0.000 |
| 2004 | 0.35 | 0.35 | 0.35 | 0.35 | 0.000 |
| 2005 | 0.34 | 0.34 | 0.34 | 0.34 | 0.002 |
| 2006 | 0.34 | 0.35 | 0.35 | 0.36 | 0.007 |
| 2007 | 0.34 | 0.36 | 0.36 | 0.39 | 0.017 |
| 2008 | 0.34 | 0.38 | 0.38 | 0.023 |  |
| 2009 | 0.34 | 0.38 | 0.38 | 0.42 | 0.026 |
| 2010 | 0.34 | 0.38 | 0.38 | 0.42 | 0.027 |
| 2011 | 0.34 | 0.39 | 0.39 | 0.42 | 0.027 |
| 2012 | 0.34 | 0.39 | 0.39 | 0.42 | 0.027 |
| 2013 | 0.34 | 0.38 | 0.38 | 0.42 | 0.026 |
| 2014 | 0.34 | 0.39 | 0.42 | 0.026 |  |
| 2015 | 0.34 |  | 0.42 | 0.026 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 70.1 | 70.1 | 70.1 | 70.1 | 0.00 |
| 2004 | 55.5 | 55.6 | 55.6 | 55.7 | 0.07 |
| 2005 | 55.0 | 55.7 | 55.8 | 56.8 | 0.56 |
| 2006 | 58.4 | 61.7 | 62.0 | 67.2 | 2.80 |
| 2007 | 59.4 | 68.4 | 69.4 | 83.0 | 7.76 |
| 2008 | 57.9 | 72.6 | 74.3 | 94.1 | 11.50 |
| 2009 | 57.6 | 75.1 | 76.5 | 98.7 | 12.88 |
| 2010 | 57.8 | 76.3 | 77.4 | 100.1 | 13.32 |
| 2011 | 58.0 | 76.2 | 77.7 | 100.6 | 13.46 |
| 2012 | 58.0 | 76.8 | 77.7 | 99.5 | 13.36 |
| 2013 | 58.4 | 76.1 | 77.4 | 99.8 | 12.99 |
| 2014 | 57.9 | 76.0 | 77.3 | 99.8 | 12.89 |
| 2015 | 57.8 | 76.0 | 77.4 | 99.7 | 12.96 |

Table 2.35-Equilibrium reference points and projections for GOA Pacific cod spawning biomass (1000s of t , fishing mortality, and catch ( 1000 s of t ) under the assumption that $F=\max F_{A B C}$ in each year 20032004 and $F=F_{O F L}$ thereafter, where future recruitment is drawn from a distribution based on estimated recruitments spawned during the period 1977-2000. See Table 2.28 for symbol definitions.

## Equilibrium Reference Points

| SPR | Spawning Biomass | Fishing Mortality | Catch |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 225.8 | 0 | 0 |  |  |
| 40\% | 90.3 | 0.35 | 74.4 |  |  |
| 35\% | 79.0 | 0.42 | 80.3 |  |  |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| 2003 | 88.3 | 88.3 | 88.3 | 88.3 | 0.00 |
| 2004 | 79.4 | 79.5 | 79.5 | 79.6 | 0.06 |
| 2005 | 76.9 | 77.4 | 77.5 | 78.2 | 0.40 |
| 2006 | 75.6 | 77.4 | 77.6 | 80.2 | 1.48 |
| 2007 | 75.0 | 79.3 | 79.7 | 85.7 | 3.43 |
| 2008 | 73.9 | 80.9 | 81.5 | 90.0 | 5.32 |
| 2009 | 73.3 | 81.9 | 82.7 | 93.6 | 6.46 |
| 2010 | 73.6 | 82.8 | 83.4 | 95.3 | 6.95 |
| 2011 | 73.9 | 82.9 | 83.7 | 96.1 | 7.13 |
| 2012 | 73.8 | 83.3 | 83.8 | 96.4 | 7.16 |
| 2013 | 74.2 | 83.0 | 83.8 | 96.3 | 7.06 |
| 2014 | 74.1 | 82.9 | 83.8 | 95.8 | 6.95 |
| 2015 | 74.0 | 82.9 | 83.8 | 95.9 | 6.90 |

## Fishing Mortality Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0.34 | 0.34 | 0.34 | 0.34 | 0.000 |
| 2004 | 0.31 | 0.31 | 0.31 | 0.31 | 0.000 |
| 2005 | 0.36 | 0.36 | 0.36 | 0.36 | 0.002 |
| 2006 | 0.35 | 0.36 | 0.36 | 0.37 | 0.007 |
| 2007 | 0.35 | 0.37 | 0.37 | 0.40 | 0.023 |
| 2008 | 0.34 | 0.38 | 0.38 | 0.42 | 0.026 |
| 2009 | 0.34 | 0.38 | 0.38 | 0.42 | 0.027 |
| 2010 | 0.34 | 0.38 | 0.38 | 0.42 | 0.027 |
| 2011 | 0.34 | 0.39 | 0.38 | 0.42 | 0.027 |
| 2012 | 0.34 | 0.39 | 0.39 | 0.42 | 0.026 |
| 2013 | 0.34 | 0.38 | 0.38 | 0.42 | 0.026 |
| 2014 | 0.34 | 0.39 | 0.42 | 0.026 |  |

## Catch Projections

| Year | L90\%CI | Median | Mean | U90\%CI | St. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 59.9 | 59.9 | 59.9 | 59.9 | 0.00 |
| 2004 | 50.6 | 50.7 | 50.7 | 50.8 | 0.06 |
| 2005 | 60.4 | 61.1 | 61.1 | 62.2 | 0.59 |
| 2006 | 61.0 | 64.4 | 64.7 | 70.0 | 2.85 |
| 2007 | 60.4 | 69.4 | 70.5 | 84.1 | 7.76 |
| 2008 | 58.1 | 72.8 | 74.5 | 94.3 | 11.46 |
| 2009 | 57.5 | 75.0 | 76.4 | 98.6 | 12.86 |
| 2010 | 57.7 | 76.2 | 77.3 | 100.0 | 13.32 |
| 2011 | 58.0 | 76.1 | 77.6 | 100.5 | 13.46 |
| 2012 | 57.9 | 76.8 | 77.6 | 99.5 | 13.37 |
| 2013 | 58.4 | 76.1 | 77.4 | 99.8 | 12.99 |
| 2014 | 57.9 | 76.0 | 77.3 | 99.8 | 12.89 |
| 2015 | 57.8 | 76.0 | 77.4 | 99.7 | 12.96 |

Table 2.36--Summary of major results for the stock assessment of Pacific cod in the GOA region.

| Natural mortality rate: |  | 0.37 |
| :---: | :---: | :---: |
| Reference fishing mortalities: | Rate | Value |
|  | $F_{35 \%}$ | 0.42 |
|  | $F_{40 \%}$ | 0.35 |
|  | $\max F_{A B C}$ | 0.34 |
| Reference spawning biomass: | Type | Value |
|  | $B_{35 \%}$ | 79,000 t |
|  | $B_{40 \%}$ | 90,300 t |
| Projected biomass for 2003: | Type | Value |
|  | Age 3+ | 452,000 t |
|  | Spawning (at max $F_{A B C}$ ) | 88,300 t |
| Recommended ABC for 2003: | Units | Value |
|  | Fishing Mortality | 0.30 |
|  | Catch | 52,800 t |
| Overfishing level for 2003: | Units | Value |
|  | Fishing Mortality | 0.41 |
|  | Catch | 70,100 t |

2000 Period 1 Trawl Catch Size Composition


2000 Period 1 Longline Catch Size Composition


2000 Period 1 Pot Catch Size Composition


Fig
ure 2.1-Estimated and observed size compositions from the 2000 period 1 fisheries (Model 1).

2001 Period 1 Trawl Catch Size Composition


2001 Period 1 Longline Catch Size Composition


2001 Period 1 Pot Catch Size Composition


Fig
ure 2.2-Estimated and observed size compositions from the 2001 period 1 fisheries (Model 1).

2002 Period 1 Trawl Catch Size Composition


2002 Period 1 Longline Catch Size Composition


2002 Period 1 Pot Catch Size Composition


Fig
ure 2.3-Estimated and observed size compositions from the 2002 period 1 fisheries (Model 1).

2000 Period 1 Trawl Catch Size Composition


2000 Period 1 Longline Catch Size Composition


2000 Period 1 Pot Catch Size Composition


Fig
ure 2.4-Estimated and observed size compositions from the 2000 period 1 fisheries (Model 2).

2001 Period 1 Trawl Catch Size Composition


2001 Period 1 Longline Catch Size Composition


2001 Period 1 Pot Catch Size Composition


Fig
ure 2.5-Estimated and observed size compositions from the 2001 period 1 fisheries (Model 2).

2002 Period 1 Trawl Catch Size Composition


2002 Period 1 Longline Catch Size Composition


2002 Period 1 Pot Catch Size Composition


Figure 2.6-Estimated and observed size compositions from the 2002 period 1 fisheries (Model 2).

1996 Bottom Trawl Survey Size Composition


1999 Bottom Trawl Survey Size Composition


2001 Bottom Trawl Survey Size Composition


Figure 2.7-Estimated and observed size compositions from the 3 most recent surveys (Model 1).

1996 Bottom Trawl Survey Size Composition


1999 Bottom Trawl Survey Size Composition


2001 Bottom Trawl Survey Size Composition


Figure 2.8-Estimated and observed size compositions from the 3 most recent surveys (Model 2).


Figure 2.9-Comparison of biomass estimates resulting from Models 1 and 2.


Figure 2.10-Retrospective analysis of estimated survey biomass, 1997-present. The vertical error bars around the observed survey biomass represent 1.96 standard deviations on either side of the mean.


Figure 2.11-Retrospective analysis of estimated age 3+ biomass, 1997-present.


Figure 2.12-Pacific cod recruitment at age 3 as estimated by the assessment model.


Figure 2.13-Some aspects of uncertainty surrounding the stock-recruitment relationship. The upper panel shows a $95 \%$ confidence ellipse for the estimated parameters of the stock-recruitment relationship, with dashed lines indicating the location of the point estimates. The lower panel shows the data (small squares), the estimated relationship (bold curve), and the $95 \%$ confidence interval around the curve (thin curves), with dashed lines indicating the locations of the data means. See text for details and caveats.

Appendix 2A: Approximate Functional Representations of Population Dynamics Used in Synthesis

These equations are similar to those used in Synthesis. Symbols are defined in Table 2.14.

## Functions of Length or Age

Weight at length:

$$
w(\lambda)=W_{1} \lambda^{W_{2}}
$$

Proportion mature at length:

$$
p(\lambda)=\frac{1}{1+\exp _{\left(-P_{1}\left(P_{2}-\lambda\right)\right.}}
$$

Length at age:

$$
l(\alpha)=L_{1}+\left(L_{2}-L_{1}\right)\left(\frac{1-\exp _{( }-K\left(\alpha-\alpha_{1}\right)_{)}}{1-\exp _{( }-K\left(\alpha_{2}-\alpha_{1}\right)_{)}}\right)
$$

Standard deviation of length at age:

$$
x(\alpha)=X_{1}+\left(X_{2}-X_{1}\right)\left(\frac{l(\alpha)-L_{1}}{L_{2}-L_{1}}\right)
$$

Probability density function describing distribution of length, conditional on age:

$$
h(\lambda \mid \alpha)=\sqrt{\frac{1}{2 \pi}}\left(\frac{1}{x(\alpha)}\right) \exp \left(-\left(\frac{1}{2}\right)\left(\frac{\lambda-l(\alpha)}{x(\alpha)}\right)^{2}\right)
$$

Selectivity at length $\lambda \leq S_{g, 4, e(y \mid g)}$ (ascending limb), conditional on gear type and year:

$$
\begin{aligned}
& s(\lambda \mid g, y)=S_{g, 1, e(y \mid g)}+ \\
& \quad\left(1-S_{g, 1, e(y \mid g)}\right)\left(\frac{\frac{1}{1+\exp _{( }-S_{g, 3, e(y \mid g)}\left(\lambda-S_{g, 2, e(y \mid g)}\right)}-\frac{1}{1+\exp _{( }-S_{g, 3, e(y \mid g)}\left(\lambda_{\min }-S_{g, 2, e(y \mid g)}\right)}}{\frac{1}{1+\exp _{( }-S_{g, 3, e(y \mid g)}\left(S_{g, 4, e(y \mid g)}-S_{g, 2, e(y \mid g)}\right)}-\frac{1}{1+\exp _{( }-S_{g, 3, e(y \mid g)}\left(\lambda_{\min }-S_{g, 2, e(y \mid g)}\right) \mid}}\right)
\end{aligned}
$$

Selectivity at length $\lambda \geq S_{g, 4, e(y \mid g)}$ (descending limb), conditional on gear type and year:

$$
\left.\begin{array}{l}
s(\lambda \mid g, y)=1+ \\
\quad\left(1-S_{g, 5, e(y \mid g)}\right)\left(\frac{\frac{1}{1+\exp _{( }-S_{g, 7, e(y \mid g)}\left(\lambda-S_{g, 6, e(y \mid g)}\right)}-\frac{1}{1+\exp _{( }-S_{g, 7, e(y \mid g)}\left(S_{g, 4}-S_{g, 6, e(y \mid g)}\right)}}{\left.\frac{1}{\left.1+\exp _{\left(-S_{g, 7, e(y \mid g)}\left(\lambda_{\max }-S_{g, 6, e(y \mid g)}\right)\right.}\right)}-\frac{1}{1+\exp _{( }-S_{g, 7, e(y \mid g)}\left(S_{g, 4, e(y \mid g)}-S_{g, 6, e(y \mid g)}\right)}\right)}\right.
\end{array}\right) .
$$

## Numbers at Age

Matrix for converting numbers at length into numbers at age:

$$
z_{a, i, j}=\frac{\int_{l_{\min }(j)}^{l_{\min }(j+1)} h\left(\lambda \mid a+t_{d u r}(i)\right) \mathrm{d} \lambda}{\int_{\lambda_{\min }}^{\lambda_{\max }} h\left(\lambda \mid a+t_{d u r}(i)\right) \mathrm{d} \lambda}
$$

For all $y$ :

$$
n_{a_{m i n}, y, 1}=R_{y}
$$

For all $a>a_{\text {min }}$ :

$$
n_{a, 1,1}=N_{a}
$$

For all $i<i_{\max }$ :

$$
n_{a, y, i+1}=n_{a, y, i} \sum_{j=1}^{j_{\max }}\left(z_{a, i, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i} s\left(l_{m i d}(j) \mid g, y\right)\right) t_{d u r}(i)\right)\right)
$$

For all $a<a_{\max }$ and all $y<y_{\max }$ :

$$
n_{a+1, y+1,1}=n_{a, y, i_{\max }} \sum_{j=1}^{j_{\max }}\left(z_{a, i_{\max }, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i_{\max }} s\left(l_{\operatorname{mid}}(j) \mid g, y\right)\right) t_{d u r}\left(i_{\max }\right)\right)\right)
$$

For all $y<y_{\max }$ :

$$
\begin{aligned}
n_{a_{\max }, y+1,1}= & n_{a_{\max }-1, y, i_{\max } \sum_{j=1}^{j_{\max }}\left(z_{a_{\max }-1, i_{\max }, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i_{\max }} s\left(l_{\operatorname{mid}}(j) \mid g, y\right)\right) t_{\text {dur }}\left(i_{\max }\right)\right)\right)} \\
& +n_{a_{\max }, y, i_{\max }} \sum_{j=1}^{j_{\max }}\left(z_{a_{\max }, i_{\max }, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i_{\max }} s\left(l_{\operatorname{mid}}(j) \mid g, y\right)\right) t_{d u r}\left(i_{\max }\right)\right)\right)
\end{aligned}
$$

At time of spawning:

$$
u_{a, y}=n_{a, y, i_{s p a}} \sum_{j=1}^{j_{\max }}\left(z_{a, i_{\text {spa }}, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i_{s p a}} s\left(l_{m i d}(j) \mid g, y\right)\right)\left(\tau_{s p a}-\sum_{i=1}^{i_{s p a}-1} t_{d u r}(i)\right)\right)\right)
$$

At time of survey:

$$
v_{a, y}=n_{a, y, i_{s u r}}^{j_{\max }}\left(\sum_{j=1}\left(z_{a, i_{s u r}, j} \exp \left(\left(-M-\sum_{g=1}^{g_{\max }} F_{g, y, i_{s u r}} s\left(l_{m i d}(j) \mid g, y\right)\right)\left(\tau_{s u r}-\sum_{i=1}^{i_{s u r}-1} t_{d u r}(i)\right)\right)\right)\right.
$$

## Biomass

Start-of-year biomass at ages $a>a_{r e c}$ :

$$
b_{y}=\sum_{a=a_{\text {rec }}}^{a_{\text {max }}}\left(\begin{array}{c}
j_{\text {max }} \\
\left.\left.n_{a, y, 1} \sum_{j=1} z_{a, 1, j} w\left(l_{\text {mid }}(j)\right)\right), ~\right)
\end{array}\right.
$$

Female spawning biomass:

$$
c_{y}=\frac{1}{2} \sum_{a=a_{\text {min }}}^{a_{\max }}\left(u_{a, y} \sum_{j=1}^{j_{\max }} z_{a, i_{\text {spa }},} w\left(l_{\text {mid }}(j)\right) p\left(l_{\text {mid }}(j)\right)\right)
$$

Survey biomass:

$$
d_{y}=Q \sum_{a=a_{\text {min }}}^{a_{\text {max }}}\left(v_{a, y} \sum_{j=1}^{j_{\text {max }}} z_{a, i_{\text {sur }}, j} w\left(l_{\text {mid }}(j)\right) s\left(l_{\text {mid }}(j) \mid g_{\text {sur }}, y\right)\right)
$$

