## PART 5

## LM FOOD CHAIN

## Chapter 4. Description of Data, Constants, and Other Information Necessary to Run Model

### 5.4.1 Chemical Properties of PCB Contaminants

Polychlorinated biphenyls (PCBs) have been recognized as significant environmental contaminants since 1966 (Mullin et al., 1984). Their impact is particularly evident in the Great Lakes basin (Neidermeyer and Hickey, 1976; Hesselberg et al., 1990; Oliver et al., 1989; Eisenreich et al., 1989). In this modeling project, 40 PCB congeners or coeluters were targeted for simulation of their individual bioaccumulation by fish in the lake. Most of the PCB congeners were selected for their abundance and bioaccumulative tendency in the lake ecosystem. Other PCB congeners were included to make the targeted PCB group cover the full range of PCB hydrophobicity, and thus, a better representative subset of all existing 209 PCB congeners.

Hydrophobicity of a PCB congener is measured by its octanol-water partition coefficient ( $\mathrm{K}_{\mathrm{ow}}$ ) which is the most important chemical property governing bioaccumulation of the congener in organisms. Another important chemical property involved in modeling PCB contaminants is the organic carbon partition coefficient ( $\mathrm{K}_{\text {oc }}$ ) whose value can often be correlated to that of $\mathrm{K}_{\text {ow }}$. In this work, the following empirical relationship (Eadie et al., 1990) was used:

$$
\begin{equation*}
\log K_{o c}=1.94+0.72 \log K_{o w} \tag{5.4.1}
\end{equation*}
$$

The targeted PCB congeners or co-eluter congeners are listed in Table 5.4.1 with their octanol-water partition coefficients $\mathrm{K}_{\mathrm{ow}}$. The values of $\mathrm{K}_{\mathrm{ow}}$ are those of Hawker and Connell (1988). The molecular weight (MW) for each PCB congener is also listed for additional reference.

### 5.4.2 Site-Specific Data

### 5.4.2.1 Fish Food Web Structures

The structure of a food web shows how individual organisms in the food web are related to each other through feeding interactions. This dietary information is necessary for establishing appropriate linkages among individual submodels of a food web model and is important to the accurate simulation of chemical bioaccumulation in the food web.

The fish food webs of interest are those of two top predators in Lake Michigan, lake trout and coho salmon. These two species were selected for their important economic value. It is desirable to have a better understanding of the present and future concentrations of PCB contaminants in these two fish populations with the help of model simulations.

### 5.4.2.1.1 Lake Trout Food Web

It is believed that the lake trout in Lake Michigan are represented by three subpopulations at Sturgeon Bay, Sheboygan Reef, and Saugatuck (Figure 5.4.1). Movements of lake trout in Lake Michigan are believed to be considerably restricted in range (Brown et al., 1981). Each of the lake trout subpopulations has a site-specific food web structure.

Table 5.4.1. Targeted PCB Congeners and Their $\mathrm{K}_{\text {ow }}$

| Congener | IUPAC | Homolog | Molecular Weight | $\log K_{\text {ow }}$ |
| :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 154 | 4.09 |
| 4 | 3 | 1 | 188 | 4.69 |
| 2,3 | 5 | 2 | 223 | 4.97 |
| 2,4' | 8 | 2 | 223 | 5.07 |
| 3,4 | 12 | 2 | 223 | 5.22 |
| 3,4' | 13 | 2 | 223 | 5.29 |
| 4,4' | 15 | 2 | 223 | 5.3 |
| 2,4',4 | 17 | 3 | 257 | 5.25 |
| 2,2',3 | 16 | 3 | 257 | 5.16 |
| 2,4',6 | 32 | 3 | 257 | 5.44 |
| 2,2',5 | 18 | 3 | 257 | 5.24 |
| 2,3'5 | 26 | 3 | 257 | 5.66 |
| 2,4,4 | 28 | 3 | 257 | 5.67 |
| 2,4',5 | 31 | 3 | 257 | 5.67 |
| 2',3,4 | 33 | 3 | 257 | 5.6 |
| 3,4,4' | 37 | 3 | 257 | 5.83 |
| 2,2',3,4' | 42 | 4 | 292 | 5.76 |
| 2,2',3,5' | 44 | 4 | 292 | 5.75 |
| 2,2'4,5' | 49 | 4 | 292 | 5.85 |
| 2,2',5,5' | 52 | 4 | 292 | 5.84 |
| 2,3,3',4' | 56 | 4 | 292 | 6.11 |
| 2,3,4,4 | 60 | 4 | 292 | 6.11 |
| 2,3',4,4' | 66 | 4 | 292 | 6.2 |
| 2,3',4',5 | 70 | 4 | 292 | 6.2 |
| 2',3,4,5 | 76 | 4 | 292 | 6.13 |
| 2,4,4',5 | 74 | 4 | 292 | 6.2 |
| 3,3',4,4' | 77 | 4 | 292 | 6.36 |
| 2,3,3',4',6 | 110 | 5 | 326 | 6.48 |
| 3,4,4',5 | 81 | 4 | 292 | 6.36 |
| 2,2',3,4,5' | 87 | 5 | 326 | 6.29 |
| 2,2',3,3',6 | 84 | 5 | 326 | 6.04 |
| 2,2',3,5,5' | 92 | 5 | 326 | 6.35 |
| 2,2',3,4,6' | 89 | 5 | 326 | 6.07 |
| 2,2',3,4,4' | 85 | 5 | 326 | 6.3 |
| 2,2',4,4',5 | 99 | 5 | 326 | 6.39 |
| 2,2',4,5,5' | 101 | 5 | 326 | 6.38 |
| 2,3',4,4',5 | 118 | 5 | 326 | 6.74 |
| 2',3,4,4',5 | 123 | 5 | 326 | 6.74 |
| 2,2',3,4',5',6 | 149 | 6 | 361 | 6.67 |
| 2,3,3',4,4' | 105 | 5 | 326 | 6.65 |
| 2,2',3,3',4,6' | 132 | 6 | 361 | 6.58 |
| 2,2',4,4',5,5' | 153 | 6 | 361 | 6.92 |
| 2,2',3,5,5',6 | 151 | 6 | 361 | 6.64 |
| 2,2',3,4,4',5' | 138 | 6 | 361 | 6.83 |
| 2,3,3',4',5,6 | 163 | 6 | 361 | 6.99 |
| 2,2',3,4',5,5' | 146 | 6 | 361 | 6.89 |
| 2,2',3, ${ }^{\prime}, 4,4^{\prime}, 5$ | 170 | 7 | 395 | 7.27 |

Table 5.4.1. Targeted PCB Congeners and Their $\mathrm{K}_{\mathrm{ow}}$ (Continued)

| Congener | IUPAC | Homolog | Molecular Weight | log Kow |
| :--- | :---: | :---: | :---: | :---: |
| $2,3,3^{\prime}, 4,4^{\prime}, 5,6^{\prime}$ | 190 | 7 | 395 | 7.46 |
| $2^{\prime}, 2^{\prime}, 3,3^{\prime}, 4,5,5^{\prime}$ | 172 | 7 | 395 | 7.33 |
| $2^{\prime}, 2^{\prime}, 3,3^{\prime}, 4,44^{\prime}, 6,6^{\prime}$ | 197 | 8 | 430 | 7.3 |
| $2^{2}, 2^{\prime}, 3,4,4^{\prime}, 5,5^{\prime}$ | 180 | 7 | 395 | 7.36 |
| $2^{2}, 2^{\prime}, 3,4,4^{\prime}, 5,6^{\prime}$ | 182 | 7 | 395 | 7.2 |
| $2^{2}, 2^{\prime}, 3,4^{\prime}, 5,55^{\prime}, 6$ | 187 | 7 | 395 | 7.17 |
| $2^{2}, 2^{\prime}, 3,3^{\prime}, 4,4^{\prime}, 5,6$ | 195 | 8 | 430 | 7.56 |
| $2^{2}, 2^{\prime}, 3,3^{\prime}, 4,5,5^{\prime}, 6,6^{\prime}$ | 208 | 9 | 464 | 7.71 |
| $2,2^{\prime}, 3,3^{\prime}, 4,44^{\prime}, 5^{\prime}, 6$ | 196 | 8 | 430 | 7.65 |
| $2,2^{\prime}, 3,4,4^{\prime}, 5,5^{\prime}, 6$ | 203 | 8 | 430 | 7.65 |
| $2,2^{\prime}, 3,3^{\prime}, 4^{\prime}, 5,5^{\prime}, 6$ | 201 | 8 | 430 | 7.62 |



Figure 5.4.1. Biota zones in Lake Michigan.

For each lake trout subpopulation, the food web was constructed using dietary data compiled from field sampling of lake trout and associated forage fish population. Lake trout (Salvelinus namaycush) were caught at the three locations during the spring, summer, and fall of 1994 and 1995. They were primarily captured via gill netting at depths ranging from 9 to 40 m . A minor portion of trout was captured by bottom trawling. Bottom trawling was used at depths of 10 to 50 m to obtain forage fish. Prey fish included alewife (Alosa pseudoharengus), rainbow smelt (Osmerus mordax), bloater (Coregonus hoyi), slimy sculpin (Cottus cognatus), and deepwater sculpin (Myoxocephalus thompsoni). The diets of lake trout and forage fish were determined by stomach analysis following a standard operating procedure established for the Lake Michigan Mass Balance Project (LMMBP) (U.S. Environmental Protection Agency, 1997a). For lake trout, the diet components were further classified into age classes.

The organisms in the base of Lake Michigan fish food webs are zooplankton, Mysis, and Diporeia. Their dietary information was obtained from literature sources. Mysis are reported to feed on zooplankton, phytoplankton, and "fresh" detrital material at the sediment surface and suspended in the water column (Beeton and Bowers, 1982; Grossnickle, 1982). Zooplankton are believed to feed on organic-rich particles, mainly phytoplankton in the water column (Peters and Downing, 1984). Diporeia are reported to feed on relatively "fresh" detrital material at the sediment surface (Evans et al., 1990; Gardner et al., 1990; Johnson, 1987; Lydy and Landrum, 1993; Marzolf, 1965; Quigley, 1988; Quigley and Vanderploeg, 1991).

Annual average dietary data for lake trout and its forage populations in the three biota zones of the lake are summarized in Tables 5.4.2a through 5.4.7. These data were used to construct a complete food web structure for each of the three lake trout populations in Lake Michigan.

### 5.4.2.1.2 Coho Salmon Food Web

The coho salmon in Lake Michigan are believed to move around large portions of the lake during the fish's lifetime (Patriarche, 1980). They were modeled as a single lake-wide population. The dietary
information of the coho salmon was compiled from field sampling. Coho salmon (Oncorhynchus kisutch) were sampled from angler's catches at various locations of the lake from May to November in 1994 and April to November in 1995.

The diet of coho salmon was determined by stomach analysis following a standard operating procedure established for the LMMBP (Elliott et al., 1996; Elliott and Holey, 1998; U.S. Environmental Protection Agency, 1997a). The prey species were further classified into age classes. The results are presented in Table 5.4.8.

Due to their extensive movement, coho salmon in the lake may encounter site-specific forage populations in different regions. This means that a given forage species in the coho salmon diet may belong to different subpopulations. The forage fish may have a location-dependent dietary history. Therefore, the food web structure below the top trophic level can vary with the movement of coho salmon. In order to construct an accurate food web structure for coho salmon in Lake Michigan, information on its migration pattern and food web structures of its forage populations in related locations is needed. The migration pattern of the coho salmon was established based on a general index of fish density, catch-per-unit-of-effort (CPE), in various locations on a monthly basis. In general, the fish aggregate in southern Lake Michigan during spring and travel to the southwestern region of the lake in summer. In the late summer and early autumn, most of the coho salmon are found in the northeastern region of the lake. They move back to the southeastern region during the winter. However, dietary information for forage fish in these locations were not readily available. Therefore, it was not possible to construct a comprehensive food web structure for coho salmon that reflects the seasonal or spatial variation of its forage food webs.

The most complete dietary information for forage fish was that collected from the Sturgeon Bay, Sheboygan Reef, and Saugatuck lake trout biota zones (Tables 5.4.3 through 5.4.7). In this study, these dietary data were used to construct three local food web structures for the coho salmon by linking each of them with the dietary data of the coho salmon as presented in Table 5.4.8.

Table 5.4.2a. Annual Dietary Composition of Lake Trout at Saugatuck (1994-1995)


Table 5.4.2a. Annual Dietary Composition of Lake Trout at Saugatuck (1994-1995) (Continued)

| Lake <br> Trout Age | Forage Fish Age | Alewife | Rainbow Smelt | Bloater | Slimy Sculpin | Deepwater Sculpin | Diporeia | Mysis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 11 | Age 3 | 10 |  |  |  |  |  |  |
|  | Age 4 |  |  | 30 |  |  |  |  |
|  | Age 5 |  |  | 25 |  |  |  |  |
|  | Age 6 | 10 |  |  |  |  |  |  |
|  | Age 7 |  |  | 25 |  |  |  |  |
| Age 12 | Age 1 |  | 5 |  |  |  |  |  |
|  | Age 2 |  |  |  |  |  |  |  |
|  | Age 3 | 10 |  |  |  |  |  |  |
|  | Age 4 |  |  |  |  |  |  |  |
|  | Age 5 | 20 |  | 15 |  |  |  |  |
|  | Age 6 |  | 10 | 30 |  |  |  |  |
|  | Age 7 | 10 |  |  |  |  |  |  |

Table 5.4.2b. Annual Dietary Composition of Lake Trout at Sheboygan Reef (1994-1995)

| Lake <br> Trout Age | Forage Fish Age | Alewife | Rainbow Smelt | Bloater | Slimy Sculpin | Deepwater Sculpin | Diporeia | Mysis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 | Age 1 | 85 |  |  |  |  | 15 |  |
| Age 2 | Age 1 | 80 |  |  |  | 10 | 5 | 5 |
| Age 3 | Age 1 Age 2 | 55 |  |  |  |  |  | 45 |
| Age 4 | Age 1 <br> Age 2 <br> Age 3 <br> Age 4 <br> Age 5 | $\begin{aligned} & 20 \\ & 20 \\ & 10 \end{aligned}$ |  | $\begin{aligned} & 10 \\ & 10 \end{aligned}$ |  | 10 |  | 20 |
| Age 5 | Age 1 <br> Age 2 <br> Age 3 <br> Age 4 <br> Age 5 <br> Age 6 | $\begin{aligned} & 20 \\ & 15 \\ & 15 \\ & 10 \end{aligned}$ |  | 20 10 |  |  |  | 10 |
| Age 6 | Age 2 <br> Age 3 <br> Age 4 <br> Age 5 | $\begin{aligned} & 30 \\ & 20 \\ & 10 \\ & 10 \end{aligned}$ |  | 10 20 |  |  |  |  |

Table 5.4.2b. Annual Dietary Composition of Lake Trout at Sheboygan Reef (1994-1995) (Continued)

| Lake | Forage |  |  |  |  |
| :--- | :--- | :---: | :--- | :--- | :--- |
| Trout | Fish | Alewife | Rainbow <br> Smelt | Bloater | Slimy <br> Sculpin |
| Age | Age |  | Deepwater <br> Sculpin | Diporeia |  |

Table 5.4.2c. Annual Dietary Composition of Lake Trout at Sturgeon Bay (1994-1995)

| Lake <br> Trout <br> Age | Forage <br> Fish <br> Age | Alewife | Rainbow <br> Smelt | Bloater | Slimy <br> Sculpin | Deepwater <br> Sculpin | Diporeia |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | Mysis

Table 5.4.2c. Annual Dietary Composition of Lake Trout at Sturgeon Bay (1994-1995) (Continued)

| Lake <br> Trout <br> Age | Forage <br> Fish <br> Age | Alewife | Rainbow <br> Smelt | Bloater | Slimy <br> Sculpin | Deepwater <br> Sculpin | Diporeia |
| :--- | :--- | :---: | :---: | :---: | :---: | :---: | :---: | Mysis

Table 5.4.3. Dietary Composition of Alewife in Lake Michigan (1994-1995)

|  | Prey | Saugatuck $(0-<75 \mathrm{~m})$ | Sturgeon Bay $(0-\sim 100 \mathrm{~m})$ | Sheboygan Reef (50-75 m) |
| :---: | :---: | :---: | :---: | :---: |
| Small: Fish Length < 120 mm |  |  |  |  |
| Age 1-2 | Diporeia Mysis | 10 | 45 | 40 |
|  | Zooplankton | 90 | 55 | 60 |
| Large: Fish Length > 120 mm |  |  |  |  |
| Age 3-7 | Diporeia | 10 | 75 | 20 |
|  | Mysis |  |  | 50 |
|  | Zooplankton | 90 | 25 | 30 |

Table 5.4.4. Dietary Composition of Bloater in Lake Michigan (1994-1995)

|  | Prey | Saugatuck $(0-<75 \mathrm{~m})$ | Sturgeon Bay $(0-\sim 100 \mathrm{~m})$ | Sheboygan Reef (50-75 m) |
| :---: | :---: | :---: | :---: | :---: |
| Small: Fish Length <= 160 mm |  |  |  |  |
| Age 1-3 | Diporeia <br> Mysis <br> Zooplankton | $\begin{aligned} & 80 \\ & 20 \end{aligned}$ | 100 | $\begin{aligned} & 35 \\ & 35 \\ & 30 \end{aligned}$ |
| Large: Fish Length (g) $\mathbf{>} \mathbf{1 6 0 ~ m m}$ |  |  |  |  |
| Age 4-7 | Diporeia Mysis Zooplankton | $\begin{aligned} & 75 \\ & 25 \end{aligned}$ | $\begin{aligned} & 70 \\ & 30 \end{aligned}$ | $\begin{aligned} & 25 \\ & 75 \end{aligned}$ |

Table 5.4.5. Dietary Composition of Rainbow Smelt in Lake Michigan (1994-1995)

|  | Prey | Saugatuck <br> $\mathbf{( 0 - < \mathbf { 7 5 } \mathbf { ~ m } )}$ | Sturgeon Bay <br> $(\mathbf{0}-\sim \mathbf{1 0 0} \mathbf{~ m})$ | Sheboygan Reef <br> $\mathbf{( 5 0 - 7 5} \mathbf{~})$ |
| :--- | :--- | :---: | :---: | :---: |
| All Ages | Diporeia |  | 10 |  |
|  | Mysis | 65 | 90 | 60 |
|  | Zooplankton | 35 |  | 40 |

Table 5.4.6. Dietary Composition of Slimy Sculpin in Lake Michigan (1994-1995)

|  | Prey | Saugatuck <br> $\mathbf{( 0 - < \mathbf { 7 5 } \mathbf { ~ m } )}$ | Sturgeon Bay <br> $\mathbf{( 0 - \sim 1 0 0 ~} \mathbf{~})$ | Sheboygan Reef <br> $(\mathbf{5 0}-\mathbf{7 5} \mathbf{~ m})$ |
| :--- | :--- | :---: | :---: | :---: |
| All Ages | Diporeia | 90 | 80 | 90 |
|  | Mysis | 10 | 20 | 10 |

Table 5.4.7. Dietary Composition of Deepwater Sculpin in Lake Michigan (1994-1995)

|  | Prey | Saugatuck <br> $(\mathbf{0}-<\mathbf{7 5} \mathbf{~ m})$ | Sturgeon Bay <br> $(\mathbf{0}-\sim \mathbf{1 0 0} \mathbf{~ m})$ | Sheboygan Reef <br> $\mathbf{( 5 0 - 7 5 ~ \mathbf { m } )}$ |
| :--- | :--- | :---: | :---: | :---: |
| All Ages | Diporeia | 70 | 45 | 80 |
|  | Mysis | 30 | 55 | 20 |

Table 5.4.8. Dietary Composition of Coho Salmon in Lake Michigan (1994-1995)

| Coho <br> Salmon Age | Forage Fish <br> Age | Alewife | Rainbow Smelt | Bloater | Diporeia | Mysis |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Age 1 | Age 1 | 40 | 10 | 10 |  |  |
|  | Age 2 | 40 |  |  |  |  |
| Age 2 | Age 1 | 25 |  |  |  |  |
|  | Age 2 | 10 |  |  |  |  |
|  | Age 3 | 20 | 5 |  |  |  |
|  | Age 4 | 20 |  |  |  |  |
|  | Age 5 | 10 |  |  |  |  |
|  | Age 6 | 10 |  |  |  |  |
|  | Age 7 | 10 |  |  |  |  |
|  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |

### 5.4.2.2 Fish Growth Rates

At a given body weight, W , fish growth rate, G , can be written as:

$$
\begin{equation*}
G=(d w / d t) / W \tag{5.4.2}
\end{equation*}
$$

where
$\begin{aligned}(d w / d t)= & \text { the derivative of fish weight } \mathrm{W} \text { with } \\ & \text { respect to fish age } t\end{aligned}$
With a set of weight-age data of a fish available, the average value for the fish growth rate for a given period of time can then be estimated by the following equation:

$$
\begin{equation*}
G=\ln \left(W_{1} / W_{o}\right) /\left(t_{1}-t_{o}\right) \tag{5.4.3}
\end{equation*}
$$

where
$\boldsymbol{W}_{1}=$ fish weight (g) at age $\mathrm{t}_{1}$ (day)
$\boldsymbol{W}_{\boldsymbol{o}}=$ fish weight $(\mathrm{g})$ at age $\mathrm{t}_{0}$ (day)
$G=$ fish average growth rate during age $t_{0}$ to $t_{1}$
The weight-age data for fish species in the food webs were obtained from field sampling conducted in 1994-1995 by the Great Lakes National Program Office (GLNPO) for the LMMBP. The methods of fish collection are described in Section 4.2.1. Each fish was weighed to the nearest gram. The lake trout and coho salmon were aged based on either decoding the information on a coded-wire tag (if found) or enumeration of annuli on scales in conjunction with use of fin clip information. More details on the fish aging procedure can be found in Lake Michigan Mass Balance Study Methods Compendium (U.S.

Environmental Protection Agency, 1997a) and Madenjian et al. (1998a, 1999). Forage fish were aged based on lengths and weights taken from the literature, and compared to the length and weight data collected for each of the fish species in this study.

A general relationship between age and weight for each fish was established through regression of the large amount of field data. The age-weight relationships for the lake trout in three biota zones, the migratory coho salmon, and their forage fish populations are presented in Tables 5.4.9a through 5.4.9c. Age-weight relationships for forage fish exhibit no regional variation, and a lake-wide average was obtained for each forage species. The results in Tables 5.4.9a, 5.4.9b, and 5.4.9c were used to estimate fish growth rates in the food web models.

The weight-age relationship for Mysis was estimated based on information from literature sources (Brafield and Llewellyn, 1962; Pothoven et al., 2000). The results are presented in Table 5.4.9d.

A constant value of 0.10 ( $1 /$ day) was adapted as the average growth rate for zooplankton in the lake (Connolly et al., 1992).

### 5.4.2.3 Energy Density of Food Web Components

In a bioenergetics-based food web model, energy balance is the basis for estimating chemical fluxes between fish and its prey species. It is, therefore, important to have a good knowledge of the energy content of the fish and its prey items.

Energy densities, D, of all fish species in this study were estimated based on lipid and protein fractions in individual organisms (Lucas, 1996).
$D=35.5 f_{L}+20.08 f_{p r}$
The terms $f_{L}$ and $f_{p r}$ are lipid and protein fractions in the fish body, respectively. The energy equivalents of lipid components ( $\mathrm{kJ} / \mathrm{g}$ ) is 35.5 , and the energy equivalents of protein components $(\mathrm{kJ} / \mathrm{g})$ is 20.08 . The standard value of energy equivalent for protein is $23.4 \mathrm{~kJ} / \mathrm{g}$-protein (Cho et al., 1982). It was adjusted to a lower value of $20.08 \mathrm{~kJ} / \mathrm{g}$-protein because after digestion, a portion of energy in the assimilated protein is lost by nitrogenous excretion
and is not available for further respiration. Energy contributions from other body components of a fish, such as carbohydrates, are negligible (Diana, 1995).

Fish lipid content was analyzed by extracting homogenized fish composite with 100 mL of $90 / 10$ ( $\mathrm{v}: \mathrm{v}$ ) petroleum ether/ethyl acetate. The extract was then evaporated and the residue was weighed as extractable lipid. Detailed procedures for fish lipid separation and determination are available in the Lake Michigan Mass Balance Study Methods Compendium (U.S. Environmental Protection Agency, 1997b) and Madenjian et al (2000). The values of protein fraction in the lake trout, coho salmon, and the other fish were compiled from or estimated based on various literature sources (Flath and Diana, 1985; Foltz and Norden, 1977; Gardner et al., 1985; Rottiers and Tucker, 1982; Schindler et al., 1971; Vijverberg and Frank, 1976). The lipid and protein fractions used for estimating energy content for all organisms in this study are compiled in Tables 5.4.10a through 5.4.10h.

### 5.4.2.4 Exposure Conditions

Environmental conditions to which fish are exposed play an important part in determining chemical exchange fluxes between a fish and its environment. Among the model parameters which characterize the environmental conditions for food webs, contaminant levels in water and sediment have direct influence on the contaminant level in exposed fish food webs, and temperature and oxygen content of the exposure environment regulate the chemical kinetics in fish food webs.

Due to the variation in Lake Michigan water characteristics, the exposure condition is different among fish food webs in different biota zones. To facilitate model calculations for fish food webs at Sturgeon Bay, Sheboygan Reef, and Saugatuck, exposure information for each of these three biota zones was required. Exposure data used are summarized here. All data for the LMMBP are available upon request to the GLNPO.

Table 5.4.9a. Average Weight-Age Relationships for Lake Trout in Lake Michigan (1994-1995)

| Age | Sheboygan Reef <br> Weight $\mathbf{( g )}$ | Saugatuck <br> Weight $\mathbf{( g )}$ | Sturgeon Bay <br> Weight $\mathbf{( g )}$ |
| :---: | :---: | :---: | :---: |
| 1 | 20 | 90 | 98 |
| 2 | 128 | 180 | 120 |
| 3 | 244 | 550 | 350 |
| 4 | 490 | 1100 | 800 |
| 5 | 900 | 2050 | 1500 |
| 6 | 1378 | 2850 | 2700 |
| 7 | 1900 | 3400 | 3200 |
| 8 | 2600 | 4000 | 4700 |
| 9 | 3400 | 5400 | 5000 |
| 10 | 4000 | 6500 | 5500 |
| 11 | 4400 | 6900 | 5600 |
| 12 | 4700 | 7100 | 5800 |
| 13 | 4900 | 7100 | 6000 |

Table 5.4.9b. Average Weight-Age Relationships for Coho Salmon in Lake Michigan (1994-1995)

| Age | Day | Weight (g) |
| :---: | :---: | :---: |
| 1 | 90 | 30 |
|  | 122 | 80 |
|  | 152 | 140 |
|  | 183 | 220 |
|  | 214 | 322 |
|  | 244 | 450 |
|  | 274 | 620 |
|  | 304 | 878 |
| 2 | 335 | 880 |
|  | 366 | 885 |
|  | 30 | 890 |
|  | 60 | 895 |
|  | 90 | 900 |
|  | 121 | 1400 |
|  | 151 | 1850 |
|  | 183 | 2190 |
|  | 214 | 2450 |
|  | 244 | 2670 |
|  | 274 | 2860 |
|  | 304 | 3050 |

Table 5.4.9c. Average Weight-Age Relationships of Forage Fish in Lake Michigan (1994-1995)

| Age | Alewife <br> Weight $\mathbf{( g )}$ | Bloater <br> Weight $\mathbf{( g )}$ | Rainbow Smelt <br> Weight $\mathbf{( g )}$ | Slimy Sculpin <br> Weight $\mathbf{( g )}$ | Deepwater Sculpin <br> Weight $(\mathbf{g})$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 3 | 3.7 | 5.3 | 0.6 | 0.6 |
| 2 | 15 | 12 | 8 | 1.2 | 1.8 |
| 3 | 27 | 26 | 13 | 2.2 | 3.5 |
| 4 | 37 | 38 | 19 | 4.6 | 7 |
| 5 | 45 | 50 | 22 | 10 | 13 |
| 6 | 50 | 65 | 25 | 10.6 | 19 |
| 7 | 53 | 88 | 28 |  | 24 |
| 8 | 55 | 110 | 30 | 29 |  |
| 9 |  |  | 32 |  | 34 |
| 10 |  |  |  |  | 38 |
| 11 |  |  |  |  | 40 |

Table 5.4.9d. Estimated Weight-Age Relationships of Mysis in Lake Michigan

| Month | Weight (g-wet) <br> Sturgeon Bay | Weight (g-wet) <br> Sheboygan Reef | Weight (g-wet) <br> Saugatuck |
| :---: | :---: | :---: | :---: |
| 0 | 0.00019 | 0.00001 | 0.00001 |
| 4 | 0.00194 | 0.00061 | 0.00095 |
| 8 | 0.00893 | 0.00330 | 0.00537 |
| 12 | 0.01691 | 0.00910 | 0.01706 |
| 16 | 0.03336 | 0.01860 | 0.04123 |

Table 5.4.10a. Average Lipid and Protein Fractions (\%) of Lake Trout in Lake Michigan (1994-1995)

| Age | Sheboygan Reef | Sturgeon Bay | Saugatuck | Protein $\%$ |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 2.3 | 4.8 | 2.3 | 17.37 |
| 2 | 3.66 | 4.68 | 3.66 |  |
| 3 | 7.9 | 9.21 | 7.13 |  |
| 4 | 9.36 | 11.81 | 9.52 |  |
| 5 | 12.48 | 17.04 | 18.77 |  |
| 6 | 15.56 | 18.3 | 21.06 |  |
| 7 | 18.6 | 19.13 | 18.56 |  |
| 8 | 19.36 | 20.52 | 19.12 |  |
| 9 | 19.34 | 20.15 | 20.68 |  |
| 10 | 19.1 | 22.63 | 23 |  |
| 11 | 20.73 | 22.5 | 21.7 |  |
| 12 | 22.4 | 20.53 | 19.7 |  |
| 13 | 20.2 | 20.9 | 30.6 |  |
| 14 | 20.1 | 21.4 |  |  |

Table 5.4.10b. Average Lipid and Protein Fractions (\%) of Coho Salmon in Lake Michigan (1994-1995)

| Age | Day | Lipid \% | Protein \% |
| :---: | :---: | :---: | :---: |
| 1 | 90 | 5.14 | 20.00 |
|  | 122 | 5.25 |  |
|  | 152 | 5.37 |  |
|  | 183 | 5.54 |  |
|  | 214 | 5.75 |  |
|  | 244 | 6.01 |  |
|  | 274 | 6.36 |  |
|  | 304 | 6.90 |  |
|  | 335 | 6.90 |  |
|  | 366 | 6.91 |  |
| 2 | 30 | 6.92 |  |
|  | 60 | 6.93 |  |
|  | 90 | 6.94 |  |
|  | 121 | 7.98 |  |
|  | 151 | 8.91 |  |
|  | 183 | 9.61 |  |
|  | 214 | 10.15 |  |
|  | 244 | 10.61 |  |
|  | 274 | 11.00 |  |
|  | 304 | 11.39 |  |

Table 5.4.10c. Average Lipid and Protein Fractions (\%) of Alewife in Lake Michigan (1994-1995)

| Age | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein \% |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 7.2 | 5.5 | 4 | 16.7 |
| 2 | 8.5 | 5.5 | 6 |  |
| 3 | 9 | 6 | 6 |  |
| 4 | 10.5 | 7.5 | 6 |  |
| 5 | 11.5 | 9 | 6 |  |
| 6 | 12 | 10 | 6 |  |
| 7 | 12.2 | 11 | 6 |  |
| 8 | 12.5 | 12 |  |  |

Table 5.4.10d. Average Lipid and Protein Fractions (\%) of Bloater in Lake Michigan (1994-1995)

| Age | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein $\%$ |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 5 | 4 | 5 | 16.3 |
| 2 | 5.5 | 4.5 | 7 |  |
| 3 | 8 | 5.5 | 8.5 |  |
| 4 | 11 | 6.5 | 9.5 |  |
| 5 | 12 | 8.5 | 12.5 |  |
| 6 | 12.5 | 10.5 | 14.5 |  |
| 7 | 13 | 11 | 15.5 |  |
| 8 | 13.5 |  |  |  |

Table 5.4.10e. Average Lipid and Protein Fractions (\%) of Rainbow Smelt in Lake Michigan (1994-1995)

| Age | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein \% |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 4.4 | 3.5 | 3 | 16.9 |
| 2 | 4.4 | 3.5 | 3 |  |
| 3 | 4.4 | 3.5 | 3 |  |
| 4 | 4.4 | 3.5 | 3 |  |
| 5 | 4.4 | 3.5 | 3 |  |
| 6 | 4.4 | 3.5 | 3 |  |
| 7 | 4.4 | 3.5 | 3 |  |
| 8 | 4.4 | 3.5 | 3 |  |
| 9 | 4.4 | 3.5 | 3 |  |

Table 5.4.10f. Average Lipid and Protein Fractions (\%) of Slimy Sculpin in Lake Michigan (1994-1995)

| Age | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein \% |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 6.4 | 3.5 | 8 | 15.9 |
| 2 | 6.5 | 4 | 8.1 |  |
| 3 | 6.6 | 4.5 | 8.2 |  |
| 4 | 6.8 | 5.3 | 8.4 |  |
| 5 | 7.1 | 5.2 | 8.5 |  |
| 6 | 7.2 | 5.2 | 8.5 |  |
| 7 | 7.3 |  |  |  |

Table 5.4.10g. Average Lipid and Protein Fractions (\%) of Deepwater Sculpin in Lake Michigan (19941995)

| Age | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein \% |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 8.8 | 2 | 7 | 14.4 |
| 2 | 8.9 | 3 | 7.1 |  |
| 3 | 9 | 4 | 7.2 |  |
| 4 | 9.1 | 5 | 7.3 |  |
| 5 | 9.4 | 5.5 | 7.5 |  |
| 6 | 9.7 | 6 | 7.8 |  |
| 7 | 9.9 | 7 | 7.9 |  |
| 8 | 10.1 | 7.2 | 8 |  |
| 9 | 10.3 | 7.2 | 8.1 |  |
| 10 | 10.5 | 7.5 | 8.2 |  |
| 11 | 10.6 |  |  |  |

Table 5.4.10h. Average Lipid and Protein Fractions (\%) of Zooplankton, Mysis, and Diporeia in Lake Michigan (1994-1995)

| Species | Sheboygan Reef | Saugatuck | Sturgeon Bay | Protein \% |
| :---: | :---: | :---: | :---: | :---: |
| Zooplankton | 2.91 | 2.79 | 1.57 | 7.1 |
| Mysis | 2.31 | 1.61 | 2.9 | 7 |
| Diporeia | 3.21 | 1.66 | 4.48 | 10 |

### 5.4.2.4.1 PCB Concentrations in Water

Lake Michigan water and particulate samples were collected at several stations within the Sturgeon Bay, Sheboygan Reef, and Saugatuck biota zones. Information regarding the sampling stations, collection procedures, sample preparation, and methods for PCB analysis are available in detail (U.S. Environmental Protection Agency, 1997a, 1997b). The organic carbon fraction in the suspended particles was also analyzed. The analysis procedures can also be found in the above documents.

No temporal variation of PCB concentrations was found for samples collected during 1994 and 1995. PCB concentrations in suspended particles were organic carbon normalized. There was substantial variation of PCB concentrations in suspended
particles among samples collected from different water depths. No substantial vertical variation was found for PCBs in the dissolved form. PCBs in suspended particles were divided into those collected at depth $<20 \mathrm{~m}$ and those collected at depth $>20 \mathrm{~m}$. For this study, it was assumed that the fish food webs were exposed to particulate PCB concentrations in the deeper layer. Median values for dissolved PCBs and those associated with suspended particles were used for model calibration. The PCB concentrations in the water column of the three biota zones are given in Table 5.4.11.

### 5.4.2.4.2 PCB Concentrations in Sediment

Sediment sampling was not specifically conducted within the three biota zones. Sediment PCB concentrations in the three biota zones were

Table 5.4.11. PCB Concentrations in Lake Michigan Water Column (1994-1995)

| PCB <br> Congeners | Sturgeon Bay |  | Sheboygan Reef |  | Saugatuck |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Dissolved (ng/L) | Particulate ( $\mathrm{ng} / \mathrm{g}-\mathrm{OC}$ ) | Dissolved (ng/L) | Particulate (ng/g-OC) | Dissolved (ng/L) | Particulate (ng/g-OC) |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8+5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 12 | 0.002831 | 0 | 0.002265 | 6.94990 | 0.003126 | 0 |
| 13 | 0.001163 | 0.63374 | 0.00122 | 2.09185 | 0.0009 | 2.11576 |
| 15+17 | 0.003063 | 4.02012 | 0.002608 | 7.54759 | 0.004061 | 11.95844 |
| 16 | 0 | 1.00157 | 0 | 1.56798 | 0.001473 | 3.22824 |
| 32 | 0 | 1.37860 | 0 | 1.58024 | 0 | 4.22044 |
| 18 | 0.00333 | 3.57836 | 0.00377 | 5.47443 | 0.004623 | 10.4442 |
| 26 | 0.000941 | 0.22132 | 0.001258 | 0.37498 | 0.001582 | 2.96423 |
| 28+31 | 0.008012 | 12.42481 | 0.007067 | 17.84289 | 0.009846 | 55.58153 |
| 33 | 0.004408 | 2.28478 | 0.005054 | 3.43611 | 0.006045 | 9.97024 |
| 37+42 | 0.008967 | 14.5969 | 0.009517 | 15.35747 | 0.008866 | 16.86476 |
| 44 | 0.003189 | 5.38999 | 0.002878 | 7.30135 | 0.00581 | 20.89396 |
| 49 | 0.002259 | 3.96632 | 0.002054 | 6.86181 | 0.003302 | 14.05582 |
| 52 | 0.005627 | 9.48455 | 0.005518 | 16.12783 | 0.008475 | 35.36909 |
| 56+60 | 0.00134 | 7.20351 | 0.001344 | 13.76338 | 0.00198 | 30.47388 |
| 66 | 0.001664 | 18.39126 | 0.001893 | 29.53261 | 0.002783 | 63.51781 |
| 70+76 | 0.002179 | 7.46113 | 0.0021 | 16.42939 | 0.003036 | 33.74864 |
| 74 | 0.00103 | 4.18880 | 0.001039 | 5.84207 | 0.001371 | 14.10166 |
| 77+110 | 0.00291 | 13.79423 | 0.002586 | 28.80211 | 0.004342 | 49.83354 |
| 81 | 7.68E-05 | 1.52913 | 0 | 2.09813 | 0.000147 | 2.60703 |
| 87 | 0.00227 | 4.43503 | 0.002572 | 8.03297 | 0.002373 | 13.37302 |
| 92+84 | 0.005722 | 15.83466 | 0.007226 | 32.15896 | 0.01356 | 74.07366 |
| 89 | 0.00068 | 0.15860 | 0 | 0 | 0 | 1.59772 |
| 85 | 0.000507 | 4.76618 | 0.000569 | 8.63774 | 0.000681 | 13.30061 |
| 99 | 0.006156 | 25.2633 | 0.004236 | 36.02048 | 0.004228 | 49.72043 |
| 101 | 0.001328 | 10.76926 | 0.00278 | 17.31631 | 0.004522 | 34.31707 |
| 118 | 0.001236 | 10.49375 | 0.001156 | 19.16489 | 0.001713 | 35.9058 |
| 123+149 | 0.000705 | 6.59078 | 0.000862 | 13.43283 | 0.001331 | 21.52096 |
| 132+153+105 | 0.000724 | 18.7597 | 0.000958 | 31.21532 | 0.001451 | 58.52275 |
| 151 | 0 | 2.11833 | 0 | 3.88177 | 0 | 6.37438 |
| 163+138 | 0.002134 | 20.59195 | 0.002948 | 37.87159 | 0.002877 | 55.57444 |
| 146 | 0.00059 | 5.19236 | 0.000583 | 7.64438 | 0.000572 | 9.26933 |
| 170+190 | $4.74 \mathrm{E}-05$ | 2.54427 | 7.36E-05 | 5.30883 | 0.000131 | 8.24745 |
| 172+197 | 0 | 1.03453 | 0 | 1.85133 | 0 | 2.81659 |
| 180 | 0 | 1.91204 | 5.13E-05 | 5.94020 | 0 | 18.31716 |
| 187+182 | 0.002588 | 4.91753 | 0.000984 | 7.07428 | 0.000683 | 12.63331 |
| 208+195 | $4.34 \mathrm{E}-05$ | 0.88921 | 0 | 1.91721 | 0 | 2.42335 |
| 196+203 | 3.28E-05 | 1.50532 | $2.75 \mathrm{E}-05$ | 3.83510 | 0 | 5.10087 |
| 201 | 0.000168 | 3.05836 | 7.34E-05 | 6.59986 | 0.00018 | 9.01875 |

estimated based on samples collected at several nearby stations. These stations were selected for their closeness to a specific biota zone in distance, depth, and sediment characteristics. Because organic carbon normalized sediment PCB data showed limited horizontal variation, the estimate of sediment PCB exposure by using data from nearby stations was appropriate. Information regarding the sampling stations, collection procedures, sample preparation, and methods for PCB analysis are available in detail (U.S. Environmental Protection Agency, 1997a, 1997b). Organic carbon and dry fraction of sediment samples were also analyzed. The analysis procedures can also be found in the above documents.

Sediment data analysis revealed no significant temporal variation in PCB concentrations for samples collected during 1994 and 1995. PCB concentrations in sediment were organic carbon normalized. Median values for PCBs in sediment carbon were used for model calculations. The concentrations of PCBs dissolved in sediment pore water were estimated based on measured PCB data, organic carbon content, dry fraction in the sediment samples, and organic carbon-water partition coefficients for individual PCB congeners. The results of PCB concentrations in the sediment solids and pore water for the three biota zones are given in Table 5.4.12.

### 5.4.2.4.3 Exposure Temperature

Lake Michigan is a vast water body with a volume of $4,920 \mathrm{~km}^{3}$. It has a surface area of $57,800 \mathrm{~km}^{2}$, and its deepest point is 282 m (Coordinating Committee on Great Lakes Basic Hydraulic and Hydrologic Data, 1992). Physical characteristics of the lake vary with region and depth (Environment Canada and U.S. Environmental Protection Agency, 1997). To better reflect this reality, the model was constructed to simulate the exposure environment for each species, rather than as a whole for all species in a food web.

The prevailing annual cycles of exposure temperature for a lake-wide coho salmon population and for three lake trout and their forage populations at Sturgeon Bay, Sheboygan Reef, and Saugatuck were established and are presented in Figures 5.4.2a through 5.4.2c. The results were compiled based on site-specific information, such as annual water temperature profiles (U.S. Environmental Protection

Agency, 1995), species optimal temperature and depth at different life stages (Otto et al., 1976; Peterson et al., 1979; Stewart et al., 1983; Wismer and Christie, 1987; Wells, 1968), prey availability (Crowder and Crawford, 1984; Eck and Wells, 1986; Janssen and Brandt, 1980), spawning season (Janssen and Brandt, 1980), and spawn site preference (Jude et al., 1986; Rice, 1985). For simplicity, the exposure temperatures for different age groups in certain species were aggregated and average annual temperature cycles were determined for the species. The seasonal variation of surface water temperatures (U.S. Environmental Protection Agency, 1995) in the lake is also presented in the first panel of Figures 5.4.2a - 5.4.2b and Figure 5.4.2c for reference.

### 5.4.2.4.4 Oxygen Concentration in Water

The oxygen concentration in water that organisms vent through their gill membranes was determined by water temperature. In this study, the dissolved oxygen content in water $\left[\mathrm{O}_{2}\right]$ was estimated according to an empirical correlation between oxygen solubility ( $\mathrm{mg} / \mathrm{L}$ ) and water temperature (Greenberg et al., 1992).

$$
\begin{align*}
\operatorname{Ln}\left[O_{2}\right] & =-139.34411+\left(1.575701 \times 10^{5} / T\right) \\
& -\left(6.642308 \times 10^{7} / T^{2}\right)+\left(1.2438 \times 10^{10} / T^{3}\right) \\
& -\left(8.621949 \times 10^{11} / T^{4}\right) \tag{5.4.5}
\end{align*}
$$

where
$\boldsymbol{T}=$ temperature $\left({ }^{\circ} \mathrm{K}\right)$

### 5.4.3 Physiological Data of Fish and Other Organisms

### 5.4.3.1 Species-Specific Respiration Rates

In the bioenergetics-based food web model (LM Food Chain), fish respiration (or metabolism) rate is a key model parameter which determines the dynamics of chemical uptake from water and food. Fish respiration rate is dependent on fish weight, temperature, and degree of fish activity. For most of the fish species in the Lake Michigan food webs, an extensive study of respiration as a function of weight,

Table 5.4.12. PCB Concentrations in Lake Michigan Surface Sediment (1994-1995)

| PCB <br> Congeners | Sturgeon Bay |  | Sheboygan Reef |  | Saugatuck |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pore Water (ng/L) | $\begin{gathered} \text { Particle } \\ \text { (ng/g-OC) } \end{gathered}$ | Pore Water (ng/L) | $\begin{gathered} \text { Particle } \\ \text { (ng/g-OC) } \end{gathered}$ | Pore Water (ng/L) | $\begin{gathered} \text { Particle } \\ \text { (ng/g-OC) } \end{gathered}$ |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8+5 | 0.0279054 | 10.00643 | 0.0382893 | 13.72811 | 0.0908558 | 32.58473 |
| 12 | 0.0014341 | 0.71635 | 0 | 0 | 0 | 0 |
| 13 | 0.0013447 | 0.75432 | 0.0029074 | 1.63073 | 0.0061499 | 3.44944 |
| 15+17 | 0.0096751 | 5.29398 | 0.0252991 | 13.84212 | 0.0681406 | 37.28868 |
| 16 | 0.0038924 | 1.76031 | 0.0074487 | 3.36831 | 0.0171774 | 7.76701 |
| 32 | 0.0024807 | 1.41452 | 0.0026843 | 1.53063 | 0.0277008 | 15.80006 |
| 18 | 0.0071207 | 3.67665 | 0.018433 | 9.51698 | 0.0504009 | 26.02673 |
| 26 | 0.0026656 | 2.76099 | 0.0035926 | 3.72098 | 0.0167924 | 17.39422 |
| 28+31 | 0.049642 | 52.27209 | 0.067541 | 71.12439 | 0.2105729 | 221.7646 |
| 33 | 0.0125176 | 11.73799 | 0.0149408 | 14.00947 | 0.0549273 | 51.50962 |
| 37+42 | 0.009989 | 13.00778 | 0.019252 | 25.29948 | 0.061323 | 80.17273 |
| 44 | 0.0149115 | 17.92995 | 0.023162 | 27.84947 | 0.0644197 | 77.46354 |
| 49 | 0.0078881 | 11.19489 | 0.0098061 | 13,91655 | 0.0308258 | 43.75034 |
| 52 | 0.0174969 | 24.42368 | 0.0227162 | 31.70843 | 0.0629656 | 87.89781 |
| 56+60 | 0.0183856 | 40.15014 | 0.0281371 | 61.44742 | 0.0645118 | 140.8908 |
| 66 | 0.0446862 | 113.2876 | 0.05182 | 131.3768 | 0.1180745 | 299.3598 |
| 70+76 | 0.0180978 | 43.29467 | 0.0239207 | 57.22649 | 0.0613122 | 146.6866 |
| 74 | 0.0075816 | 19.2216 | 0.007947 | 20.14776 | 0.0219209 | 55.5777 |
| 77+110 | 0.0175742 | 69.76274 | 0.026113 | 102.5709 | 0.0514251 | 200.6624 |
| 81 | 0.0004445 | 1.46919 | 0.0006704 | 2.21588 | 0.0018093 | 5.98041 |
| 87 | 0.0049453 | 14.55518 | 0.0062292 | 18.33357 | 0.0162384 | 47.79412 |
| 92+84 | 0.0089757 | 22.56819 | 0.0178462 | 44.87101 | 0.0322442 | 81.07549 |
| 89 | 0.000158 | 0.3229 | 0.0011103 | 2.26910 | 0.0019127 | 3.90920 |
| 85 | 0.00761 | 22.77225 | 0.00995 | 29.77426 | 0.0154561 | 46.2521 |
| 99 | 0.0065556 | 22.77379 | 0.007798 | 27.0894 | 0.0150233 | 52.19077 |
| 101 | 0.0116518 | 39.81201 | 0.0138112 | 47.18955 | 0.029209 | 99.80335 |
| 118 | 0.0104703 | 64.97851 | 0.0130125 | 80.75468 | 0.0212667 | 131.9822 |
| 123+149 | 0.0034083 | 19.95935 | 0.0040488 | 23.71038 | 0.0083891 | 49.12844 |
| 132+153+105 | 0.0133833 | 79.90554 | 0.0151232 | 95.42261 | 0.0230193 | 145.2468 |
| 151 | 0.0011542 | 6.06862 | 0.0012538 | 6.59256 | 0.0026042 | 13.69295 |
| 163+138 | 0.0099996 | 82.25939 | 0.0127299 | 104.7201 | 0.0209239 | 172.1294 |
| 146 | 0.0014001 | 11.14243 | 0.0015557 | 12.38059 | 0.0028539 | 22.71165 |
| 170+190 | 0.0009791 | 17.12468 | 0.0009686 | 16.94218 | 0.0016554 | 28.95361 |
| 172+197 | 0.0002736 | 4.51562 | 0.0002758 | 4.53335 | 0.0005621 | 9.23983 |
| 180 | 0.0017814 | 30.90044 | 0.0018309 | 31.75848 | 0.0030391 | 52.71641 |
| 187+182 | 0.0010107 | 13.11631 | 0.000956 | 12.40633 | 0.0016946 | 21.99273 |
| 208+195 | 0.0001999 | 5.46997 | 0.000171 | 4.67938 | 0.0002984 | 8.16719 |
| 196+203 | 0.0006348 | 17.80964 | 0.0005772 | 16.19255 | 0.0011243 | 31.54036 |
| 201 | 0.0006772 | 18.07668 | 0.0004869 | 12.99829 | 0.0012951 | 34.57076 |



Figure 5.4.2a. Typical annual cycles of exposure temperature for Lake Michigan food webs at Saugatuck and Sturgeon Bay.


Figure 5.4.2b. Typical annual cycles of exposure temperature for Lake Michigan food web at Sheboygan Reef.


Figure 5.4.2c. Typical annual cycles of exposure temperature for coho salmon in Lake Michigan.
temperature, and swimming speed was conducted, and results were reported (Lantry and Stewart, 1993; Rudstam, 1989; Rudstam et al., 1994; Stewart et al., 1983; Stewart and Binkowski, 1986). In general, a fish's daily respiration rate, in $\mathrm{g}-\mathrm{O}_{2} /$ day, can be formulated as:
$R=\alpha W^{\beta} \cdot e^{\rho T} \cdot e^{v U}$
where $\alpha, \beta, \rho, v$ are species-specific empirical constants, W is weight, and U is the swimming speed of the fish, in $\mathrm{cm} / \mathrm{s}$.

For a given aquatic species, the swimming speed can be expressed as a function of body weight and water temperature:
$U=\omega W^{\delta} e^{\phi T}$
where $\omega, \delta, \phi$ are species-specific empirical constants.

The values of the species-specific empirical constants used to estimate the respiration rate were collected from literature sources (Lantry and Stewart, 1993; Rudstam, 1989; Rudstam et al., 1994; Stewart et al., 1983; Stewart and Binkowski, 1986) and are listed in Table 5.4.13. For slimy and deepwater sculpin, there was insufficient information available to generate species-specific respiration rates. As an alternative, their respiration rates were estimated using the generalized fish respiration equation. The constants used for the calculation of their respiration rates were also given in the table.

In this study, a value of $13.56 \mathrm{~kJ} / \mathrm{g}-\mathrm{O}_{2}$ (Elliott and Davison, 1975; Brafield and Llewellyn, 1982; Crisp, 1984) was used as the respiratory energy equivalent, or oxycalorific coefficient, for converting oxygen respiration to energy utilized by fish.

For zooplankton, a simple equation was used to estimate its respiration, in $\mathrm{kJ} / \mathrm{gwet} /$ day, as a function of water temperature (Connolly et al., 1992):

$$
\begin{equation*}
R=0.60 e^{\rho T} \tag{5.4.8}
\end{equation*}
$$

### 5.4.3.2 Respiration Rates Adjusted for Specific Dynamic Action (SDA)

The respiration rate estimated with Equation 5.4.6 represents the average energy requirement for the resting metabolism of a fish. It has been reported that there is an increase in respiration rate for a recently fed fish (Kayser, 1963). The additional respiration activity is often referred to as Specific Dynamic Action (SDA). The origin of the extra respiration is believed to be due to the energy necessary for the digestion of ingested foods, the absorption of nutrients, the deaminization of amino acids, and the synthesis of the products of nitrogenous excretion. In homothermic animals, it has been shown that SDA represents $30 \%$ of the caloric content of the ingested protein, $13 \%$ for a lipid, and 5\% for a carbohydrate (Lucas, 1996). Due to the difficulty in experimentally discriminating SDA from additional respiration associated with excitement and activity with feeding, different SDA

Table 5.4.13. Bioenergetic Parameters of Lake Michigan Fishes

| Parameter | Mysis | Slimy <br> Sculpin | Deepwater <br> Sculpin | Alewife | Rainbow <br> Smelt | Bloater | Lake Trout | Coho <br> Salmon |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\alpha$ | 0.00182 | $0.043^{*}$ | $0.043^{*}$ | 0.00367 | 0.0027 | 0.0018 | 0.00463 | 0.00264 |
| $\left(\mathrm{gO}_{2} /\right.$ gwet/day $)$ |  |  |  |  |  |  |  |  |
| $\beta$ | -0.161 | -0.3 | -0.3 | -0.2152 | -0.216 | -0.12 | -0.295 | -0.217 |
| $\rho$ | 0.0752 | 0.03 | 0.03 | 0.0548 | 0.036 | 0.047 | 0.059 | 0.06818 |
| $\omega$ | 0 | 1.19 | 1.19 | 5.78 | 0 | 7.23 | 11.7 | 9.7 |
| $\delta$ | 0 | 0.32 | 0.32 | -0.045 | 0 | 0.25 | 0.05 | 0.13 |
| $\phi$ | 0 | 0.045 | 0.045 | 0.149 | 0 | 0 | 0.0405 | 0.0405 |
| v | 0 | 0.0176 | 0.0176 | 0.03 | 0 | 0.025 | 0.0232 | 0.0234 |

*With a unit of gwet/gwet/day.
values were cited in the literature that ranged from $9 \%$ to $20 \%$ of the energy contained in the diet (Jobling, 1981).

In this study, the SDA is modeled as a portion of a fish's dietary ingestion. The respiration rate adjusted for SDA can then be written as:

$$
\begin{equation*}
R_{S D A}=R \cdot Q_{O X}+S D A\left(R_{S D A}+G \cdot D_{f}\right) \tag{5.4.9}
\end{equation*}
$$

where
$\boldsymbol{R}_{\text {SDA }}=$ SDA adjusted respiration rate, $\mathrm{g}-\mathrm{O}_{2} /$ day
$\boldsymbol{R}=$ resting respiration rate calculated with empirical equations, $\mathrm{g}-\mathrm{O}_{2} /$ day
$\boldsymbol{Q}_{o x}=$ respiratory energy equivalent or oxycalorific coefficient, $\mathrm{kJ} / \mathrm{g}-\mathrm{O}_{2}$

SDA $=$ fraction of assimilated energy spent on specific dynamic action
$G \quad=\quad$ fish growth rate, 1/day
$\boldsymbol{D}_{\boldsymbol{f}}=$ energy density of the fish

The final respiration rate, in $\mathrm{kJ} /$ day, was then estimated as:

$$
\begin{equation*}
R_{S D A}=\left(R \cdot Q_{o x}+S D A \cdot G \cdot D_{f}\right) /(1-S D A) \tag{5.4.10}
\end{equation*}
$$

### 5.4.4 Calibrated Model Parameters

There are several constants and variables in the model's equation whose values are either not readily available or inconclusive. Their values were determined through model calibration to site-specific conditions. The calibrated parameters include food assimilation efficiency ( $\beta$ ) for each species or age group, the chemical assimilation efficiency ( $\alpha$ ) for each species or age group for each PCB congener, the chemical relative gill transfer coefficient ( $\mathrm{E}_{\mathrm{c}} / \mathrm{E}_{\mathrm{o}}$ ) for each species (or age group) for each PCB congener, and the fraction of ingested energy for SDA for each species or age group.

An acceptable value range for each of the calibrated model parameters and its general trend for PCB congeners or species in different trophic levels was established based on information from the literature and experience gained in previous modeling work. Depending upon species and its diet, food assimilation efficiency has a value ranging from 0.05 to 0.85 (Brocksen et al., 1968; Brocksen and Brugge, 1974; Elliott, 1976; Averett, 1969). The value for the
chemical assimilation efficiency can vary from 0.2 to 0.8 and is reported to be correlated with the $\mathrm{K}_{\text {ow }}$ value for the chemical (Gobas, 1988). The chemical relative gill transfer coefficient $\left(E_{c} / E_{0}\right)$ ranges from 0.1 to 1.0 and is also believed to be related to $\mathrm{K}_{\mathrm{ow}}$ for the chemical (McKim et al., 1985). Energy fraction for SDA has a value ranging from 0.00 to 0.20 . These data were used to guide our model calibrations for appropriate parameterization.

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