## CHAPTER 13

# SHORTRAKER AND ROUGHEYE ROCKFISH 

by

## Paul D. Spencer and Rebecca F. Reuter

Relative to last years' final BSAI SAFE Report, the following changes have been made in the assessment of the Other Red Rockfish.

1) An age-structured population model for northern rockfish has been developed, and the assessment for northern rockfish is presented in a separate chapter. Thus, the remaining species of the "other red rockfish" complex are shortraker and rougheye rockfish, and the chapter name is changed to more clearly reflect the species composition.
2) The shortraker/rougheye stock complex is assessed with a Kalman filter rather than a straight averaging of survey biomass estimates.
3) The 2002 landings have been revised and the 2003 landings through September 27, 2003 have been included in the assessment.

The recommended 2004 ABC levels relative to the recommended 2003 levels are as follows:

|  | ABC |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | Eastern Bering Sea |  | Aleutian Islands |  |  |
|  | 2003 | 2004 | 2003 | 2004 | 2004 Total |
| Rougheye/Shortraker | 137 t |  | 830 t |  |  |
| Rougheye |  | 21 t |  | 174 t | 195 t |
| Shortraker |  | 84 t |  | 442 t | 526 t |
| Total | 137 t | 105 t | 830 t | 616 t |  |

The recommended 2004 OFL levels relative to the 2003 recommendations, assuming identical species complexes as used in 2003, are as follows:

|  | OFL |  |
| :--- | :---: | :---: |
|  | Eastern Bering Sea/Aleutian Islands |  |
|  | 2003 | 2004 |
| Rougheye/Shortraker | $1,289 \mathrm{t}$ |  |
| Rougheye |  | 259 t |
| Shortraker | $1,289 \mathrm{t}$ | 901 t |
| Total |  | 960 t |

## INTRODUCTION

Pacific ocean perch (POP), and four other associated species of rockfish (northern rockfish, S. polyspinis; rougheye rockfish, S. aleutianus; shortraker rockfish, S. borealis; and sharpchin rockfish, S. zacentrus) were managed as a complex in the eastern Bering Sea (EBS) and Aleutian Island (AI) management areas from 1979 to 1990. Known as the POP complex, these five species were managed as a single entity with a single TAC (total allowable catch) within each management area. In 1991, the North Pacific Fishery Management Council enacted new regulations that changed the species composition of the POP complex. For the eastern Bering Sea slope region, the POP complex was divided into two subgroups: 1) Pacific ocean perch, and 2) shortraker, rougheye, sharpchin, and northern rockfishes combined, also known as "other red rockfish" (ORR). For the Aleutian Islands region, the POP complex was divided into three subgroups: 1) Pacific ocean perch, 2) shortraker/rougheye rockfishes, and 3) sharpchin/northern rockfishes. In 2001, the other red rockfish complex in the eastern Bering Sea was split into two groups, rougheye/shortraker and sharpchin/northern, matching the complexes used in the Aleutian Islands. Additionally, separate TACs were established for the EBS and AI management areas, but the overfishing level (OFL) pertains to the entire BSAI area. These subgroups were established to protect Pacific ocean perch, shortraker rockfish, and rougheye rockfish (the three most valuable commercial species in the assemblage) from possible overfishing. In 2002, sharpchin rockfish were assigned to the "other rockfish" category, leaving only northern rockfish and the shortraker/rougheye complex as member of other red rockfish.

Sufficient age composition data exist to apply an age-structured model to northern rockfish, and the assessment of this species is presented in a separate chapter. In addition, a Kalman filter is applied to the remaining shortraker/rougheye complex. An advantage of the Kalman filter is that it utilizes both the catch estimates and the survey biomass estimates. In contrast, the method applied in previous assessments (straight averaging of survey biomass) utilizes only the catch information. The Kalman filter methodology for a single-species case was presented to the Plan Team at the Sept 2003 meeting, and the method is extended to apply to a two-species complex in this assessment.

## Information on Stock Structure

A variety of types of research can be used to infer stock structure of rougheye and shortraker rockfish, including larval distribution patterns and other life-history information, and genetic studies. In 2002, an analysis of archived Sebastes larvae was undertaken by Dr. Art Kendall; using data collected in 1990 off southeast Alaska (650 larvae) and the AFSC ichthyoplankton database (16,895 Sebastes larvae, collected on 58 cruises from 1972 to 1999, primarily in the Gulf of Alaska). The southeast Alaska larvae all showed the same morph, and were too small to have characteristics that would allow species identification. A preliminary examination of the AFSC ichthyoplankton database indicates that most larvae were collected in the spring, the larvae were widespread in the areas sampled, and most are small ( $5-7 \mathrm{~mm}$ ). The larvae were organized into three size classes for analysis: $<7.9 \mathrm{~mm}, 8.0-13.9 \mathrm{~mm}$, and $>14.0 \mathrm{~mm}$. A subset of the abundant small larvae was examined, as were all larvae in the medium and large groups. Species identification based on morphological characteristics is difficult because of overlapping characteristics among species, as few rockfish species in the north Pacific have published descriptions of the complete larval developmental series. However, all of the larvae examined could be assigned to four morphs identified by Kendall (1991), where each morph is associated with one or more species. Most of the small larvae examined belong to a single morph, which contains the species S. alutus (POP), S. polyspinus (northern rockfish), and S. ciliatus (dusky rockfish). Some larvae (18) belonged to a second morph which has been identified as S. borealis (shortraker rockfish) in the Bering Sea. The locations of these larvae were near Kodiak Island, the Semidi Islands, Chirkof Island, the Shumagin Islands, and near the eastern end of the Aleutian Islands. Another morph, represented by 58 samples in the Gulf of Alaska, could possibly represent rougheye rockfish, whose larvae have not been previously described.

For rougheye rockfish, fixed differences at a microsatellite locus and divergent mtDNA complements indicate two distinct species (Gharrett 2003). The ranges of the two species of rougheye are not coincident, although both species were caught in the same hauls in some areas. There are also two color morphs of rougheye rockfish, but these do not correspond exactly to the two species, and a way to distinguish the species morphometrically has not been identified. For the type A (western) rougheye, the microsatellite data showed weaker population structure than seen for shortraker, with one group in the central Aleutians and two large groups overlapping at Kodiak Island.

For shortraker rockfish, population structure was observed in microsatellite DNA analysis of 12 collections from Baranof Island to the western Aleutians revealed population structure roughly on a spatial scale consistent with our current management areas, although increased sample sizes may reveal finer spatial structuring (Matala et al. in press). The available data suggest are consistent with a neighborhood genetic model, suggesting that the expected dispersal of a particular specimen is much smaller than the species range. A parallel study with mtDNA revealed weaker stock structure than that observed in with the microsatellite data. The relationships among the mtDNA haplotypes suggest a population decline followed by a relatively recent (in geological time) population expansion (Gharrett 2003).

## CATCH HISTORY

Catches of shortraker and rougheye rockfish have been reported in a variety of species groups in the foreign and domestic Alaskan fisheries. Foreign catch records did not identify
rougheye and shortraker rockfish by species; instead, rougheye and shortraker rockfish were reported in management categories such as "other species" (1977, 1978), "POP complex" (19791985, 1989), and "rockfish without POP" (1986-1988). As mentioned above, the rougheye and shortraker rockfish have been managed in the domestic fishery as part of the "other red rockfish" or "shortraker/rougheye" complexes. Reported catches by management complex, and estimated catches by species, from 1993-2003 are shown in Table 13.1, with the species catches produced by computing the harvest proportions within management groups from the North Pacific Foreign Observer Program database, and applying these proportions to the estimated total catch (obtained from the NOAA Fisheries Alaska regional office). An identical procedure was used to obtain the estimates of catch by species from the 1977-1989 foreign and joint venture fisheries. Estimates of discarding by species complex are shown in Table 13.2. Rougheye and shortraker rockfish are relatively high valued species compared to northern rockfish, accounting for the lower discard rates for the "rougheye/shortraker" complex as compared to the "other red rockfish" complex.

## DATA

Fishery Catch
Catches from the domestic fishery prior to the domestic observer program were obtained from PACFIN records. Estimated domestic catches in 1990 were obtained from Guttormsen et al. 1992. Estimates of species-specific catch in 2003 were based on observer records through October 17, 2003, and estimated total catch through September 27, 2003. The time series of estimated catches from 1977-2003 are shown in Tables 13.3-13.4. Catches of shortraker and rougheye rockfish appear low in the mid-1980's, when the foreign fishery was reduced.

Estimates of catch by species can be compared to potential single-species ABC and OFL levels in order to evaluate whether excessive harvests may have occurred in the past (Tables 13.513.7). Beginning in 2001, the OFL levels for other red rockfish pertain to the entire BSAI area. Thus, the retrospective analysis of what single-species harvest limits might have been in these years is shown separately in Table 13.5, whereas years 1994 to 2000 for the AI and EBS areas are shown in Tables 13.6 and 13.7, respectively. The intent of this analysis is to investigate how our historical estimates of catch compare with species biomass estimates, and if disproportionate catch levels (relative to the biomass levels) have occurred in the past. Care should be taken not to interpret the results as evidence of overfishing, as this definition depends upon the definition of the stock or stock complex. It should also be noted that the definition of the ABC and OFL levels under past assessment procedures were highly sensitive to variability in survey biomass estimates, which was one motivation for application of a biological model to the existing data rather than sole reliance on observed biomass estimates measured with considerable uncertainty.

The estimated harvest of rougheye rockfish has occasionally exceeded their potential singlespecies harvest limits, sometimes by large amounts. For example, the 2001 BSAI rougheye rockfish catch of 615 t exceeds what the potential single-species BSAI OFL level might have been from applying an exploitation rate to the average of recent survey biomass estimates (350 t) (Table 13.5), and a similar situation occurred in 1996 when the estimated AI rougheye catch was 850 t and the potential OFL level was 587 t .

Note that observers can report shortraker and rougheye rockfish by species, or with a combined shortraker/rougheye species code. Although the combined code could not be used for estimating proportions, it has accounted for a large percentage of all shortraker and rougheye observed in recent years. The use of the combined code was especially prevalent on longline
vessels, where species identification is often made without benefit of close examination of a basket sample of fish. For example, in 2002 approximately $56 \%$ of the SR/RE observed on longline vessels was classified with the combined code. In 2003, the North Pacific Observer Program undertook changes to improve estimation of shortraker and rougheye in order to obtain representative species compositions, including making species identifications from basket sample where a detailed examination can occur.

Northern rockfish and shortraker rockfish have been the largest components of the eastern Bering Sea other red rockfish harvest from 1995 to 2000, as these two species ranged from $79 \%$ to $96 \%$ of the other red rockfish (Table 13.7). Often the estimated catches of these two species are similar, but because the population size of northern rockfish in recent assessments is estimated to be considerably smaller than shortraker rockfish (based upon average biomass of the post-1986 NMFS surveys), the northern rockfish have smaller harvest ABC levels.

The utility of the estimated catch by species is dependent on sampling a reasonable portion of the total catch for species composition. In the Aleutian Islands, the proportion of the total catch, by management group, sampled by observers and identified to species was above $50 \%$ from 1994 to 2002. In the eastern Bering Sea, the sampling ratio was above $40 \%$ from 1994 to 2000 (except 1997), but was approximately $34 \%$ for the 2002 data on the two management groups.

## Survey data

Biomass estimates for other red rockfish were produced from cooperative U.S.-Japan trawl survey from 1979-1985 on the eastern Bering Sea slope, and from 1980-1986 in the Aleutian Islands. U.S domestic trawl surveys were conducted in 1988, 1991, and 2002 on the eastern Bering Sea slope, and in 1991, 1994, 1997, 2000, and 2002 in the Aleutian Islands (Table 13.10). The 2002 eastern Bering Sea slope survey represents the initiation of a new survey time series distinct from the previous surveys in 1988 and 1991.

Previous assessments of rockfish have relied on an average of recent survey biomass estimates to obtain the current estimate of stock size. Consistent with the data used for the agestructured POP assessment, the AI survey biomass estimates are used as a suitable index of the BSAI rougheye and shortraker populations, as the bulk of these population is believed to be centered in the Aleutian Islands. Previous assessments for shortraker and rougheye rockfish have not used the cooperative U.S. - Japan AI trawl survey estimates, as these surveys were conducted with different vessels, survey gear, and sampling design relative to the U.S. domestic trawls surveys that began in 1991 (Skip Zenger, National Marine Fisheries Service, Seattle, WA, personal communication). Additionally, because previous assessment methods relied upon an average of survey biomass estimates to obtain the current estimate of stock size, the more recent survey were viewed most appropriate for this task. In this assessment, the early survey in the 1980s were used in the assessment model in order to provide some information on stock size during this portion of the time series, although it should be recognized that these data may not be strictly comparable with the most recent surveys.

The 2002 EBS slope survey represents the initiation of a new biennial survey. The most recent slope survey prior to 2002 (excluding some preliminary tows in 2000 intended for evaluating survey gear) was in 1991. The estimate of rougheye rockfish in the 2002 EBS slope survey was small (553 t) relative to the AI survey estimate (9,613 t), and the coefficient of variation for shortraker rockfish biomass in the EBS slope survey was unusually high (44\%). For these reasons, the 2002 EBS slope survey results are not used in this assessment, and the feasibility of
incorporating this time series will be evaluated in future years. Thus, the assessment procedure is conservative because the EBS biomass estimates of shortraker and rougheye rockfish are not used is determining the recommended total harvest levels.

## ANALYTICAL APPROACH

## Model structure

A simple surplus production model, the Gompertz-Fox model, was used to model the rougheye/shortraker complex, and the Kalman filter provided a method of statistically estimating the parameter values. The Gompertz-Fox model (Fox 1970) describes the rate of change of stock size as

$$
\begin{equation*}
\frac{d x}{d t}=a x(\ln (k)-\ln (x))-f x \tag{1}
\end{equation*}
$$

where $x$ is stock size, $k$ is carrying capacity, and $f$ is fishing mortality. The model is mathematically equivalent to a model of individual growth developed by Gompertz, and describes a situation where stocks at low sizes would show a sigmoidal increase in stock size to an asymptote. The GompertzFox model can be derived from the Pella-Tomlinson model (Pella and Tomlinson 1969) by taking the limit as $n$ (the parameter controlling the location of the peak of the production curve) approaches 1 . The peak of the production curve occurs at approximately $37 \%$ of the carrying capacity, in contrast to the logistic model where the peak occurs at $50 \%$ of the carrying capacity (Figure 13.1). The Gompertz-Fox model was chosen for this analysis because it is a simple model that offers some information on growth rate and carrying capacity, and it is easily transformed into a linear form suitable for the Kalman filter (Thompson 1996).

Under the Gompertz-Fox model, the rate of change of yield is modeled as $y=f x$, and the $f$ level corresponding to the maximum sustainable yield (MSY) is equivalent to the growth parameter a. Equilibrium biomass is $(b)$ is

$$
\begin{equation*}
b=k e^{-f / a} \tag{2}
\end{equation*}
$$

and the equilibrium stock size corresponding to MSY, $B_{m s y}$, is $k / e$.

## The Kalman filter

A brief review of the Kalman filter is provided here, as more thorough presentations are provided in Meinhold and Singpurwalla (1983), Harvey (1990), and Pella (1993). The Kalman filter separates the system into a model of the state variable, which describes the true (but unobserved) state of nature, and a model of the observation variables, which describes how the observed data relate to the state variable. The state variable is modeled as

$$
\begin{equation*}
X_{t}=T_{t} X_{t-1}+c_{t}+R_{t} \eta_{t} \tag{3}
\end{equation*}
$$

where $X_{t}$ is a vector of $m$ state variables at time $t, T_{t}$ is a $m \times m$ matrix, $c_{t}$ is a $m \times 1$ vector of constants, $R_{t}$ is a $m \times g$ matrix and $\eta_{t}$ is a $g \times 1$ vector of random process errors with a mean of zero and a covariance matrix of $Q_{t}$. The inclusion of the $R_{t}$ vector is useful when a particular state
variable is affected by more than one type of random disturbance. Note that when there is only a single state variable the problem simplifies considerably and all terms become scalars. For the shortraker/rougheye complex, the state variables at each time step are the log biomass of each species. Finally, the state variable is described with a distribution with an estimated mean $\alpha_{t}$ and variance $P_{t}$.

The observation equation is

$$
\begin{equation*}
Y_{t}=Z_{t} X_{t}+d_{t}+\varepsilon_{t} \tag{4}
\end{equation*}
$$

where $Y_{t}$ is a $n \times 1$ vector of observed variables, $Z_{t}$ is a $n \times m$ matrix, $d_{t}$ is a $n \times 1$ vector and $\epsilon_{t}$ is a $n$ $\times 1$ vector of random observation errors with mean zero and covariance matrix $H_{t}$.

A distinct advantage of the Kalman filter is that both the process errors and observation errors are incorporated into the parameter estimation procedure. The method by which this occurs can be understood by invoking the Bayesian concepts of "prior" and "posterior" estimates of the state variable (Meinhold and Singpurwalla 1983). Denote $\alpha_{t-1}$ as the posterior estimate of $X_{t-1}$ using all the data up to and including time $t-1$. At time step $t$, a prior estimate of the state variable is made from the state equation (Eq. 3) and the posterior estimate from the previous step $\alpha_{t-1}$. Because this prior estimate of $X_{t}$ uses all the data up to time $t-1$, it is denoted as $\alpha_{t \mid t-1}$. The prior estimate can be used with Eq. 4 to predict the observation variables at time $t$. Upon observation of $Y_{t}$ there are now two estimates of the observed variables; the observed data $Y_{t}$ and the prediction from the prior estimate $\alpha_{t \mid t-1}$. The Kalman filter updates the prior and produces a posterior estimate, $\alpha_{t \mid t}$, that results in a value of $Y_{t}$ between these two points, and the extent to which the posterior estimate differs from the prior estimate is a function of the magnitude of prediction error and the observation error variance relative to the process error variance. The posterior estimates are then used as prior estimates in the next time step to continue the recursive procedure.

Parameter estimation can be obtained by minimizing the log likelihood of the data, and the log likelihood (without constant terms) is

$$
\begin{equation*}
\ln L=-\frac{1}{2} \sum_{t=1}^{T} \ln \left|F_{t}\right|-\frac{1}{2} \sum_{t=1}^{T} v_{t}^{\prime} F_{t}^{-1} v_{t} \tag{5}
\end{equation*}
$$

where $G_{t}$ is $Z_{t} P_{t \mid t-1} Z_{t}^{\prime}+H_{t}, P_{t \mid t-1}$ (the prior estimate of the variance of the state variable) is $T_{t} P_{t-1} T_{t}^{\prime}+$ $R_{t} Q_{t} R_{t}^{\prime}$, and $v_{t}$ (the one step ahead prediction error) is $y_{t}-Z_{t} \alpha_{t \mid-1}-d_{t}$.

Application of the Gompertz-Fox model to the Kalman filter can be obtained by defining the state variable as log biomass, and using catch and survey biomass as observation variables. The log transformation of Eq. 1 is

$$
\begin{equation*}
\frac{d X}{d t}=a(B-X) \tag{6}
\end{equation*}
$$

where $X=\ln (x)$ and $B=\ln (b)=\ln \left(k \mathrm{e}^{-f / a}\right)$. The solution to this differential equation is

$$
\begin{equation*}
X_{t}=e^{-a t} X_{0}+\left(1-e^{-a t}\right) B_{t} \tag{7}
\end{equation*}
$$

where annual changes in $f_{t}$ result in $B_{t}=\ln \left(k e^{-f_{t} / a}\right)$. This solution can be also expressed in a recursive form as

$$
\begin{equation*}
X_{t+\Delta t}=e^{-a \Delta t} X_{t}+\left(1-e^{-a \Delta t}\right) B_{t} \tag{8}
\end{equation*}
$$

where $\Delta t$ is a discrete time period. For a single species case, defining $T_{t}=e^{-a \Delta t}$ and $c_{t}=\left(1-T_{t}\right) B_{t}$ produces the deterministic portion of the state equation (Eq. 3). For the two-species shortraker/rougheye example, a version of Eq. 8 would exist for each species. In this case, $T_{t}$ is a matrix of dimension 2 with the $e^{-a \Delta t}$ terms along the diagonal, and $c_{t}$ is a vector of length 2 with each term corresponding to each species.

For rougheye and shortraker rockfish, we typically have annual estimates of catch but triennial or biennial estimates of survey biomass, and this missing data complicates the observation equation. For years in which both data types are available,

$$
Y_{t}=\left[\begin{array}{l}
\ln \left(s_{-} r e_{t}\right) \\
\ln \left(s \_s r_{t}\right) \\
\ln \left(c_{-} r e_{t}\right) \\
\ln \left(c_{-} s r_{t}\right)
\end{array}\right], \quad Z_{t}=\left[\begin{array}{cc}
1 & 0 \\
0 & 1 \\
1 & 0 \\
0 & 1
\end{array}\right], \text { and } d_{t}=\left[\begin{array}{c}
\ln \left(q_{-} r e\right) \\
\ln \left(q_{-} s r\right) \\
\ln \left(f \_r e_{t}\right) \\
\ln \left(f \_s r_{t}\right)
\end{array}\right]
$$

where $s_{\_} r e_{t}$ and $s_{-} s r_{t}$ are the survey biomass estimates of rougheye and shortraker in year $t, c_{-} r e_{t}$ and $c_{-} s r_{t}$ are the aggregated catch of shortraker and rougheye during year $t, q_{-} r e$ and $q_{-} s r$ are the survey catchability coefficients, and $f_{-} r e_{t}$ and $f_{-} s r_{t}$ are the rates of removals from fishing. Note that this model formulation assumes the non-logged survey biomasses are proportional to the true biomass. Additionally, the aggregated catch during the year is used as an estimate of the rate of catch at the time of the survey, a reasonable approximation for BSAI rockfish because the survey occurs at the midpoint of the year. The observation equation simplifies when only catch data are available:

$$
Y_{t}=\left[\begin{array}{l}
\ln \left(c_{-} r e_{t}\right) \\
\ln \left(c_{-} s r_{t}\right)
\end{array}\right], \quad Z_{t}=\left[\begin{array}{ll}
1 & 0 \\
0 & 1
\end{array}\right], \text { and } d_{t}=\left[\begin{array}{l}
\ln \left(f_{-} r e_{t}\right) \\
\ln \left(f \_s r_{t}\right)
\end{array}\right]
$$

Although the observed data reflect the system at the midpoint of a year, it is expected that the instantaneous fishing mortality rate would change between calendar years; thus, a time-step of one-half year was chosen for the discretized model. At the beginning of the calendar year neither data type is available, and updating the prior estimates with observed data is not possible. In these cases, the posterior estimate is set equal to the prior estimate for the next time step (Kimura et al. 1996).

An initial estimate of the mean and variance of the state variable ( $\alpha_{0}$ and $P_{0}$, respectively) is required to begin the recursive calculations, and can be obtained in several ways. These terms could also be estimated freely along with the other model parameters, or a diffuse prior may be placed upon them (Pella 1993). However, freely estimating these parameters increases the complexity of the estimation procedure and is not recommended (Pella 1993). For this analysis, a concentrated likelihood function was used to obtain maximum likelihood estimates of the initial state variables, which were then used in a standard Kalman filter (Rosenberg 1973).

## Catch estimation error

As mentioned above, species-specific catches of shortraker and rougheye are often made from application of an observed proportion of the catch (from observer sampling) to the estimated aggregated catch for the species complex. For example, in years where shortraker and rougheye catches are reported as a two species complex, the species-specific catches would be obtained by

$$
\begin{aligned}
& C_{R E}=p_{R E} * C_{R E / S R} \\
& C_{S R}=p_{S R} * C_{R E / S R}
\end{aligned}
$$

where pre and psr are the proportion of rougheye and shortraker observed in observer sampling and Cre/sr is the aggregated catch. This estimation procedure produces quantities that can be viewed as the product of two random variables. While overall catch data are often viewed as relatively precisely observed as compared to other fisheries information, the proportions from observer sampling adds additional error. In addition two species-specific estimates of catch are likely to be correlated because they are functions of with some variables in common. For this assessment, it was assumed that the aggregated species complex catch were lognormally distributed, the species proportions from observer sampling followed a multinomial distribution, and these two random variables were independent. The variances and covariances of the log of estimated catches can be obtained from the Delta method (Seber 1982), with the variances equal to

$$
\begin{aligned}
& V\left(\ln \left(C_{R E}\right)\right)=\sigma^{2}+\frac{p_{S R}}{N p_{R E}} \\
& V\left(\ln \left(C_{S R}\right)\right)=\sigma^{2}+\frac{p_{R E}}{N p_{S R}}
\end{aligned}
$$

and the covariance between the catches equal to

$$
\operatorname{Cov}\left(\ln \left(C_{R E}\right), \ln \left(C_{S R}\right)\right)=\left(e^{\sigma^{2}}-1\right)-\frac{1}{N}
$$

where $N$ is the assumed sample size for the multinomial distribution, $\sigma$ is approximately the coefficient of variation of the aggregated complex catch, and the levels of pre and Psr are taken at their expected values.

An additional complication arises when the species-specific catch estimation procedure is applied across several areas and/or fisheries, and the total catch for each species is a sum of several random variables. In this case, define $S_{R E}$ and $S_{S R}$ as

$$
\begin{aligned}
& S_{R E}=\sum_{i} p_{R E, i} * C_{R E / S R, i} \\
& S_{S R}=\sum_{i} p_{S R, i} * C_{R E / S R, i}
\end{aligned}
$$

where $i$ indexes the total number terms in the summation. The means and variances of each of the terms within this summation are additive, and application of the Delta method yields the covariances of the log catches:

$$
\sum_{i} V\left(C_{S R / R E, i}\right) * \frac{p_{R E, i} p_{S R, i}}{S_{R E, i} S_{S R, i}}+\operatorname{Cov}\left(p_{R E, i} p_{S R, i}\right) * \frac{\left(C_{S R / R E, i}\right)^{2}}{S_{R E, i} S_{S R, i}}
$$

## Parameters Estimated Independently

The survey catchability coefficient for each species was fixed at 1.0. The parameters relating to the estimation error on catches were fixed such that $N=100$ and $\sigma=0.15$. Because of the longevity and perceived low population growth rates of shortraker and rougheye rockfish, the process error CV was set to the relatively low value of 0.05 .

## Parameters Estimated Conditionally

The parameter estimated conditionally in the model include the $a, k$, and $f_{t}$ parameters for each species. The estimation of $a$ for each species proved problematic with this dataset, and lognormal priors were utilized to stabilize parameter values. The mean of the lognormal prior was equal to the assumed natural mortality rate $M$ for rougheye rockfish (0.025), and a large CV of 1.0 was used for the variance. The natural mortality rate for rougheye rockfish was catch curve analysis (Heifetz and Clausen 1991). The rationale for expecting $a$ to approximate $M$ is because the $a$ parameter in the Gompertz-Fox model is equivalent to $F_{m s y}$, and $M$ is often used as an approximation of $F_{\text {msy }}$ (Gulland 1970).

Finally, a Monte Carlo Markov Chain (MCMC) algorithm was used to obtain estimates of parameter uncertainty (Gelman et al. 1995). One million MCMC simulations were conducted, with every 1,000th sample saved for the sample from the posterior distribution. For this assessment, the posterior marginal distribution of $a$ for shortraker rockfish is presented.

## RESULTS

## Biomass trends

For rougheye rockfish, the differences between the high cooperative survey biomass estimates in the 1980s and the lower U.S. survey biomass estimates since 1991 resulted in a decline of predicted stock biomass (Figure 13.1a). The differences in methodology between these two portions of the time series should be considered in interpreting this predicted decline, although the cooperative survey estimates are the only data available from the 1980s. The biomass estimates for the beginning of the year decline from $26,227 \mathrm{t}$ in 1980 to $10,379 \mathrm{t}$ for 2004.

Shortraker rockfish has also shown a decline in predicted beginning year stock biomass, from 38,299 t in 1980 to 23,379 t in for 2004. An increase in predicted biomass is observed during the late 1990s that corresponds to the increase in observed survey biomass estimates in 1997 and 2000, but the lower biomass estimate for 2002 results in a decline of predicted biomass from 2001 to 2004.

## Fishing mortality

The time series of estimated fishing mortality are shown in Figure 13.2, and show higher fishing mortality rates for rougheye rockfish than shortraker rockfish. The higher fishing mortality rates for rougheye rockfish in the 1990s are consistent with the analysis presented in Table 13.6 showing occasionally disproportionate catches relative to survey biomass estimates. The fishing mortality rates for rougheye rockfish since 2000 are lower (except for 2001) than those estimated for much of the 1990s. The catches of rougheye rockfish in the 1990s must be viewed in the context of the existing management a two-species complex with OFL based upon uncertain observed survey biomass estimates.

## Annual Surplus Production

Considerable uncertainty in the parameter estimates of $a$ in the Gompertz-Fox model exist for the rougheye and shortraker stocks. The lack of data regarding this parameter can be seen in plots that express the observed survey biomass and estimated catch data in unit of annual surplus production (ASP), which is the change in biomass over a period plus the catch during the period, expressed on an annual basis. Plots of ASP as a function of mean biomass are shown in Figure 13.3, and indicate little information on the $a$ parameter for either rougheye or shortraker rockfish. The $a$ parameter is related to the slope of the production curve at low stock sizes, and one could imagine alternate production curves with high levels of $a$ providing suitable fits to ASP data. Given the longevity of rougheye and shortraker rockfish, one would not expect observed surplus production to deviate far from zero, and this was the motivation for constraining $a$ by information on the natural morality rate. The observation of some levels of surplus production substantially different from zero reflects large fluctuations in estimated survey biomass that are generally inconsistent with perceived rougheye and shortraker life-history characteristics.

## Projections and harvest alternatives

Rougheye and shortraker rockfish are currently managed under Tier 5 of Amendment 56 of the NPFMC BSAI Groundfish FMP, which requires a reliable estimate of stock biomass and natural mortality rate. Estimates of $M$ for rougheye and shortraker rockfish were obtained from Heifetz and Clausen (1991), and the $F_{a b c}$ is defined as $75 \%$ of $M$. The acceptable biological catch is obtained by multiplying $F_{a b c}$ by the estimated biomass. This procedure results in the following BSAI ABCs and OFLs :

|  | 2004 biomass | M | ABC | OFL |
| :--- | :--- | ---: | :--- | :--- |
| Rougheye rockfish | 10,379 | 0.025 | 195 t | 259 t |
| Shortraker rockfish | 23,379 | 0.03 | 526 t | 701 t |

In previous assessments, the ABCs for rougheye and shortraker rockfish were partitioned between the EBS and AI management areas as a precautionary measure. Because the AI trawl survey spans the two management areas, one option is to use the proportional survey biomass from the two areas to partition the ABCs. For rougheye rockfish, the average biomass from 1991-2002 in the AI management area is $11,480 \mathrm{t}$, whereas the average from the southern Bering Sea is 950 t ; thus $92 \%$ of the estimated Aleutians Islands survey biomass for rougheye occurs in the Aleutian Islands management area. A similar calculation indicates that $94 \%$ of the shortraker rockfish AI survey biomass is found in the AI management area. Because the Aleutian Islands survey does not cover the EBS slope, it may be useful to consider the 2002 EBS slope survey biomass of 553 t and 4851 t for rougheye and shortraker, respectively. For rougheye rockfish, the combined biomass in
the EBS management area ( $950 \mathrm{t}+553 \mathrm{t}=1,503 \mathrm{t}$ ) is $11 \%$ of the combined BSAI biomass from both surveys of $13,380 \mathrm{t}$. For shortraker rockfish, the combined biomass in the EBS management area $(1,684 t+4,851 t=6,535 t)$ is $19 \%$ of the combined BSAI biomass from both surveys of $33,852 t$. Thus, it is recommended that $11 \%$ of the rougheye ABC, or 21 t , be allocated to the EBS region and 174 t be allocated to the AI region. For shortraker rockfish, it is recommended that $16 \%$ of the ABC, or $84 t$, be allocated to the EBS region and 442 t be allocated to the AI region. These results are summarized below

|  | AI ABC | EBS ABC | OFL |
| :--- | :--- | :---: | :---: |
| Rougheye rockfish | 174 t | 21 t | 259 t |
| Shortraker rockfish | 442 t | 84 t | 701 t |

In previous years, the current biomass of rougheye and shortraker rockfish was determined from an average of recent survey biomass estimates, and it is useful to calculate what the harvest quotas for 2004 under this methodology. Biomass estimates from the Aleutian Islands were obtained from averaging the 1991-2002 estimates for that portion of the AI survey within the AI management area. The biomass estimate for the EBS management area has two surveyed components-the EBS slope component and the portion of the AI survey in the EBS management region. The 2002 EBS slope survey represents the initiation of a new survey time series distinct from the previous surveys in 1988 and 1991. Biomass estimates for the EBS management area were obtained by averaging the 1991-2002 estimates for that portion of the AI survey within the EBS management area, and adding the estimate for the 2002 EBS slope survey. These two portions are considered additive because they survey different portions of the EBS management area. This procedure results in the following biomass estimates:

## Eastern Bering Sea

| Rougheye rockfish | $1,503 \mathrm{t}$ | $11,480 \mathrm{t}$ |
| :--- | :--- | :--- |
| Shortraker rockfish | $6,535 \mathrm{t}$ | $27,317 \mathrm{t}$ |

Application of the FABC and FOFL exploitation rates results in the following ABC and OFL levels:

|  | AI ABC | EBS ABC | OFL |
| :--- | :--- | :---: | :---: |
| Rougheye rockfish | 210 t | 28 t | 318 t |
| Shortraker rockfish | 664 t | 147 t | 1081 t |

Thus, the new methodology produces comparable but somewhat lower estimates of ABC and OFL, and is recommended because it uses more information than the methodology in previous assessments.

The recommendations above for single-species management of BSAI rougheye and shortraker rockfish are consistent with recent SAFE documents, in which the BSAI Plan Team has
recommended that a single BSAI-wide ABC be applied for each species, partitioned by management area in proportion to recent survey biomass estimates. Implementation was hindered by the large amount of shortraker and rougheye rockfish not identified to species by fishery observers on longline vessels. In 2003, the observer program implemented a number of changes aimed at increasing identification of shortraker and rougheye rockfish on longline vessels. With these changes in the observer program, NOAA Fisheries Alaska regional office staff thus expected to have the catch data required to implement the above single-species management recommendations for 2004.

## Summary

In summary, several quantities pertinent to the management of the shortraker and rougheye rockfish are listed below.

| Quantity | Value |
| :--- | :--- |
| $M$ (Shortraker) | 0.03 |
| $M$ (Rougheye) | 0.025 |
| Tier | 5 |
| Year 2004 Total Biomass |  |
| $\quad$ Shortraker | $23,379 \mathrm{t}$ |
| Rougheye | $10,379 \mathrm{t}$ |
| $F_{O F L}$ (Shortraker) | 0.03 |
| $F_{\text {OFL }}$ (Rougheye) | 0.025 |
| Maximum $F_{A B C}$ (Shortraker) | 0.0225 |
| Maximum $F_{A B C}$ (Rougheye) | 0.0188 |
| Recommended $F_{A B C}$ (Shortraker) | 0.0225 |
| Recommended $F_{A B C}$ (Rougheye) | 0.0188 |
| OFL (Shortraker) | 701 t |
| OFL (Rougheye) | 259 t |
| Maximum allowable ABC (Shortraker) | 526 t |
| Maximum allowable ABC (Rougheye) | 195 t |
| Recommended ABC (Shortraker) | 526 t |
| Recommended ABC (Rougheye) | 195 t |

## REFERENCES

Fox, W.W. 1970. An exponential surplus-yield model for optimizing exploited fish populations. Trans. Am Fish. Soc. 99:80-88.

Gelman, A., J.B. Carlin, H.S. Stern, and D.A. Rubin. 1995. Bayesian data analysis. Chapman and Hall, New York. 552 pp.

Gharrett, A..J. 2003. Population structure of rougheye, shortraker, and northern rockfish based on analysis of mitochondrial DNA variation and microsatellites: completion. Juneau Center of Fisheries and Ocean Sciences, University of Alaska-Fairbanks. 136 pp.

Gulland, J.A. 1970. The fish resources of the ocean. FAO Fish. Tech. Pap. 97. 425 pp.
Guttormsen, M., R. Narita, J. Gharrett, G. Tromble, and J. Berger. 1992. Summary of observer sampling of domestic groundfish fisheries in the northeast Pacific ocean and eastern Bering Sea, 1990. NOAA Tech. Memo NMFS-AFSC-5. 281 pp.

Harvey, A.C. 1990. Forcasting, structural time series models, and the Kalman Filter. Cambridge: Cambridge University Press. 554 pp.

Heifetz, J. and D. Clausen. 1991. Slope rockfish. In Stock assessment and fishery evaluation report for groundfish report for the 1992 Gulf of Alaska groundfish fishery. North Pacific Fishery Management Council, P.O. Box 103136, Anchorage, AK.

Kendall, A.W. Jr. 1991. Systematics and identification of larvae and juveniles of the genus Sebastes. Env. Biol. Fish. 30:173-190.

Matala, A.P., A.K. Gray, J. Heifetz, and A.J. Gharrett. In press. Population structure of Alaskan shortraker rockfish, Sebastes borealis, inferred from microsatellite variation. Env. Biol. Fish.

Meinhold, R.J. and N.D. Singurwalla. 1983. Understanding the Kalman Filter. Am. Stat. 37(2):123-127.

Pella, J.J. 1993. Utility of structural time series models and the Kalman filter on for predicting consequences of fishery actions. In Proceedings of the international symposium on management strategies for exploited fish populations, G. Kruse, D.M. Eggers, R.J. Marasco, C. Pautzke, and T.J. Quinn II (eds), 571-593. Alaska Sea Grant College Program, Fairbanks, AK.

Pella, J.J. and P.K. Tomlinson. 1969. A generalized stock production model. Bulletin of the InterAmerican Tropical Tuna Commission 13:419-496.

Rosenberg, B. 1973. Random coefficient models: the analysis of a cross-section of time-series by stochastically convergent parameter regression. Annals of Economic and Social Measurement 2:399-428.

Seber, G.A.F. 1982. The estimation of animal abundance, $2^{\text {nd }}$ ed. Macmillian, New York. 654 pp.
Thompson, G.G. 1996. Application of the Kalman Filter to a stochastic differential equation model of population dynamics. In Statistics in Ecology and Environmental Monitoring 2: Decision Making and Risk Assessment in Biology, D.J. Fletcher, L. Kavalieris, and B.J. Manly (eds.), 181-203. Otago Conference Series No. 6. University of Otago Press, Dunedin, New Zealand.

Table 13.1. Estimated removals (t) from 1992-2003 of other red rockfish (the sum of northern rockfish, sharpchin rockfish, shortraker rockfish, and rougheye rockfish) and the shortraker/rougheye (SRRE) complex from the eastern Bering Sea and Aleutian Islands regions, with estimates of species-specific catches. Prior to 2001, harvests in the eastern Bering Sea were managed with the ORR complex. Unless otherwise noted, catch data were obtained from summaries produced by the NMFS Alaska regional office.

| Year | Eastern Bering Sea |  |  |  | Aleutian Islands |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | ORR | SRRE | Est RE | Est SR | SRRE | Est RE | Est SR |
| 1992 | 467 |  | 65 | 72 | 1466 | 1174 | 292 |
| 1993 | 1226 |  | 82 | 184 | 1130 | 873 | 257 |
| 1994 | 129 |  | 27 | 55 | 925 | 751 | 174 |
| 1995 | 343 |  | 13 | 43 | 559 | 381 | 178 |
| 1996 | 207 |  | 23 | 68 | 959 | 850 | 109 |
| 1997 | 230 |  | 33 | 79 | 1043 | 958 | 85 |
| 1998 | 97 |  | 11 | 39 | 661 | 524 | 137 |
| 1999 | 227 |  | 11 | 69 | 485 | 383 | 102 |
| 2000 | 245 |  | 19 | 112 | 443 | 256 | 187 |
| 2001 |  | 42 | 10 | 32 | 704 | 615 | 89 |
| 2002 |  | 104 | 7 | 97 | 463 | 252 | 211 |
| 2003 |  | 90 | 6 | 84 | 256 | 154 | 102 |

*Estimated removals through September 27, 2003

Table 13.2. Estimated retained, discarded, and percent discarded of other red rockfish (ORR) and shortraker/rougheye (SR/RE) from the eastern Bering Sea (EBS) and Aleutian Islands (AI) regions. Prior to 2001, ORR in the eastern Bering Sea were managed as a single complex.

| Species |  |  |  |  |  |  |
| :--- | :--- | :--- | :--- | ---: | ---: | ---: |
| Area | Group | Year | Catch <br> Retained | Discard | Total | Percentage |
| EBS | ORR | $1993^{* *}$ | 916 | 310 | 1226 | $25.2 \%$ |
|  |  | 1994 | 28 | 101 | 129 | $78.3 \%$ |
|  |  | 1995 | 273 | 71 | 344 | $20.6 \%$ |
|  |  | 1996 | 58 | 149 | 207 | $72.0 \%$ |
|  |  | 1997 | 57 | 173 | 230 | $75.2 \%$ |
|  |  | 1998 | 41 | 71 | 112 | $63.4 \%$ |
|  |  | 1999 | 67 | 161 | 228 | $70.6 \%$ |
|  |  | 2000 | 107 | 139 | 246 | $56.5 \%$ |
| EBS | RE/SR | 2001 | 26 |  |  |  |
|  |  | 2002 | 49 | 16 | 42 | $38.1 \%$ |
|  |  |  |  | 55 | 104 | $52.9 \%$ |
| AI | SR/RE | 1993 | 733 | 397 | 1,130 | $35.1 \%$ |
|  |  | 1994 | 700 | 224 | 924 | $24.2 \%$ |
|  |  | 1995 | 455 | 103 | 558 | $18.5 \%$ |
|  |  | 1996 | 752 | 208 | 960 | $21.7 \%$ |
|  |  | 1997 | 732 | 310 | 1,042 | $29.8 \%$ |
|  |  | 1998 | 449 | 235 | 684 | $34.4 \%$ |
|  |  | 1999 | 293 | 191 | 484 | $39.5 \%$ |
|  |  | 2000 | 258 | 183 | 441 | $41.5 \%$ |
|  |  | 2001 | 457 | 246 | 703 | $35.0 \%$ |
|  |  | 2002 | 318 | 145 | 463 | $31.3 \%$ |
|  |  |  |  |  |  |  |

Table 13.3. Catches of shortraker rockfish in the BSAI area, obtained from the North Pacific Groundfish Observer Program, NMFS Alaska Regional Office, and PACFIN.

| Year | Eastern Bering Sea |  |  | Aleutian Islands |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Foreign | Joint Venture | Domestic | Foreign | Joint Venture | Domestic |  |
| 1977 | 0 |  |  | 26 |  |  | 27 |
| 1978 | 713 |  |  | 131 |  |  | 844 |
| 1979 | 372 |  |  | 977 |  |  | 1,349 |
| 1980 | 380 | 0 |  | 74 |  |  | 455 |
| 1981 | 258 | 0 |  | 315 |  |  | 573 |
| 1982 | 242 | 0 |  | 379 | 0 |  | 621 |
| 1983 | 145 | 0 |  | 89 | 1 |  | 235 |
| 1984 | 54 | 0 |  | 28 | 0 |  | 83 |
| 1985 | 19 | 0 |  | 1 | 0 |  | 21 |
| 1986 | 2 | 2 | 14 | 0 | 0 | 12 | 30 |
| 1987 | 0 | 0 | 28 |  | 0 | 36 | 64 |
| 1988 |  | 0 | 31 |  | 0 | 37 | 69 |
| 1989 |  | 0 | 58 |  | 0 | 130 | 188 |
| 1990 |  |  | 116 |  |  | 546 | 662 |
| 1991 |  |  | 211 |  |  | 250 | 461 |
| 1992 |  |  | 72 |  |  | 292 | 364 |
| 1993 |  |  | 184 |  |  | 257 | 440 |
| 1994 |  |  | 55 |  |  | 174 | 230 |
| 1995 |  |  | 43 |  |  | 178 | 222 |
| 1996 |  |  | 68 |  |  | 109 | 177 |
| 1997 |  |  | 79 |  |  | 85 | 164 |
| 1998 |  |  | 39 |  |  | 137 | 176 |
| 1999 |  |  | 69 |  |  | 102 | 171 |
| 2000 |  |  | 112 |  |  | 187 | 300 |
| 2001 |  |  | 32 |  |  | 89 | 122 |
| 2002 |  |  | 97 |  |  | 211 | 308 |
| 2003* |  |  | 84 |  |  | 102 | 186 |

[^0]Table 13.4. Catches of rougheye rockfish in the BSAI area, obtained from the North Pacific Groundfish Observer Program, NMFS Alaska Regional Office, and PACFIN.

| Eastern Bering Sea |  |  |  |  |  |  |  |  |  | Aleutian Islands |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Foreign | Joint Venture | Domestic | Foreign | Joint Venture | Domestic |  |  |  |  |  |  |

* Estimated removals through September 27, 2003.

Table 13.5. Catch of rougheye and shortraker rockfish in the Aleutian Islands from 2001 to 2003, with reported species ABC and OFL levels. In 2002, sharpchin rockfish were dropped from the other red rockfish group. The SR/RE species code includes both shortraker and rougheye rockfish.

| Species | Aleutian Islands |  |  | Eastern Bering Sea |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Observed Catch | Total Catch | ABC | Observed Catch | Total Catch | ABC | BSAI |
| 2003* Rougheye | 86.13 | 154.16 | 215 | 4.48 | 5.61 | 32 | 330 |
| Shortraker | 56.90 | 101.84 | 615 | 67.36 | 84.39 | 104 | 959 |
| SR/RE | 14.97 |  |  | 15.87 |  |  |  |
| 2002 Rougheye | 174.94 | 252.11 | 230 | 2.48 | 7.31 | 32 | 350 |
| Shortraker | 146.33 | 210.89 | 682 | 32.74 | 96.69 | 84 | 1021 |
| SR/RE | 40.44 |  |  | 11.50 |  |  |  |
| 2001 Rougheye | 362.59 | 614.67 | 230 | 6.53 | 9.63 | 32 | 350 |
| Shortraker | 52.69 | 89.33 | 682 | 21.95 | 32.37 | 84 | 1021 |
| SR/RE | 68.47 |  |  | 9.38 |  |  |  |

[^1]Table 13.6. Catch of rougheye and shortraker rockfish in the Aleutian Islands from 1994 to 2000, with potential single-species ABC and OFL levels. The SR/RE species code includes both shortraker and rougheye rockfish.

| Species | Observed Catch | Proportion of Sp. Group | Estimated total catch | ABC | OFL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 Rougheye | 141.91 | 0.5768 | 255.54 | 239 | 319 |
| Shortraker | 104.11 | 0.4232 | 187.46 | 646 | 861 |
| SR/RE | 83.77 |  |  |  |  |
| 1999 Rougheye | 285.04 | 0.7893 | 382.82 | 405 | 540 |
| Shortraker | 76.08 | 0.2107 | 102.18 | 560 | 747 |
| SR/RE | 39.28 |  |  |  |  |
| 1998 Rougheye Shortraker SR/RE | 347.62 | 0.7926 | 523.90 | 405 | 540 |
|  | 90.97 | 0.2074 | 137.10 | 560 | 747 |
|  | 73.48 |  |  |  |  |
| 1997 Rougheye Shortraker SR/RE | 723.73 | 0.9185 | 957.99 | 440 | 587 |
|  | 64.23 | 0.0815 | 85.01 | 498 | 664 |
|  | 6.49 |  |  |  |  |
| 1996 Rougheye Shortraker SR/RE | 519.52 | 0.8866 | 850.27 | 587 | 587 |
|  | 66.44 | 0.1134 | 108.73 | 664 | 664 |
|  | 8.79 |  |  |  |  |
| 1995 Rougheye Shortraker SR/RE | 195.61 | 0.6808 | 380.56 | 632 | 632 |
|  | 91.72 | 0.3192 | 178.44 | 590 | 590 |
|  | 1.58 |  |  |  |  |
| 1994 Rougheye Shortraker SR/RE | 465.96 | 0.8116 | 750.71 | 632 | 632 |
|  | 108.18 | 0.1884 | 174.29 | 590 | 590 |
|  | 0.79 |  |  |  |  |

Table 13.7. Catch of in the eastern Bering Sea from 1994 to 2000, with potential single-species ABC and OFL levels. The SR/RE species code includes both shortraker and rougheye rockfish.

| Species | Observed Catch | Proportion of Sp. Group | Estimated total catch | ABC | OFL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 Northern | 64.14 | 0.4647 | 113.86 | 34 | 45 |
| Sharpchin | 0.11 | 0.0008 | 0.19 |  |  |
| Rougheye | 10.55 | 0.0764 | 18.72 | 35 | 47 |
| Shortraker | 63.23 | 0.4581 | 112.23 | 125 | 167 |
| SR/RE | 15.95 |  |  |  |  |
| 1999 Northern | 86.84 | 0.6353 | 144.22 | 537 | 716 |
| Sharpchin | 1.83 | 0.0134 | 3.04 |  |  |
| Rougheye | 6.46 | 0.0473 | 10.73 | 51 | 68 |
| Shortraker | 41.56 | 0.3040 | 69.02 | 185 | 247 |
| SR/RE | 5.05 |  |  |  |  |
| 1998 Northern | 28.77 | 0.4841 | 46.96 | 537 | 716 |
| Sharpchin | 0.05 | 0.0009 | 0.09 |  |  |
| Rougheye | 6.91 | 0.1163 | 11.28 | 51 | 68 |
| Shortraker | 23.69 | 0.3987 | 38.67 | 185 | 247 |
| SR/RE | 8.55 |  |  |  |  |
| 1997 Northern | 24.95 | 0.5107 | 117.46 | 788 | 1051 |
| Sharpchin | 0.12 | 0.0025 | 0.58 |  |  |
| Rougheye | 6.97 | 0.1426 | 32.80 | 56 | 75 |
| Shortraker | 16.81 | 0.3442 | 79.15 | 207 | 276 |
| SR/RE | 4.66 |  |  |  |  |
| 1996 Northern | 61.27 | 0.5606 | 116.04 | 1051 | 1051 |
| Sharpchin | 0.01 | 0.0001 | 0.01 |  |  |
| Rougheye | 12.05 | 0.1103 | 22.82 | 75 | 75 |
| Shortraker | 35.97 | 0.3291 | 68.13 | 276 | 276 |
| SR/RE | 0.93 |  |  |  |  |
| 1995 Northern | 159.10 | 0.8352 | 286.48 | 1051 | 1051 |
| Sharpchin | 0.00 | 0.0000 | 0.00 |  |  |
| Rougheye | 7.33 | 0.0385 | 13.20 | 75 | 75 |
| Shortraker | 24.05 | 0.1263 | 43.31 | 276 | 276 |
| SR/RE | 0.93 |  |  |  |  |
| 1994 Northern | 20.08 | 0.3617 | 46.66 | 1051 | 1051 |
| Sharpchin | 0.02 | 0.0004 | 0.05 |  |  |
| Rougheye | 11.63 | 0.2095 | 27.02 | 75 | 75 |
| Shortraker | 23.79 | 0.4285 | 55.27 | 276 | 276 |
| SR/RE | 0.00 |  |  |  |  |

Table 13.8. Estimated biomass (t) of rougheye, shortraker, and northern rockfishes from the NMFS bottom trawl surveys. For the Aleutian Islands surveys since 1991 and the eastern Bering Sea surveys since 1988, the coefficient of variation (CV) is shown in parentheses.

| Year | Shortraker | AI survey <br> Rougheye | EBS Slope survey |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Shortraker | Rougheye |
| 1979 |  |  | 1391 | 1053 |
| 1980 | 16,983 (0.20) | 22,807 (0.79) |  |  |
| 1981 |  |  | 3571 | 816 |
| 1982 |  |  | 5176 | 605 |
| 1983 | 40,992 (0.69) | 23,412 (0.37) |  |  |
| 1984 |  |  |  |  |
| 1985 |  |  | 4010 | 1716 |
| 1986 | 25,823 (0.28) | 52,354 (0.62) |  |  |
| 1987 |  |  |  |  |
| 1988 |  |  | 1260 (0.43) | 876 (0.32) |
| 1989 |  |  |  |  |
| 1990 |  |  |  |  |
| 1991 | 23,703 (0.64) | 11,131 (0.45) | 2758 (0.38) | 884 (0.30) |
| 1992 |  |  |  |  |
| 1993 |  |  |  |  |
| 1994 | 28,190 (0.21) | 14,552 (0.26) |  |  |
| 1995 |  |  |  |  |
| 1996 |  |  |  |  |
| 1997 | 38,487 (0.26) | 11,596 (0.21) |  |  |
| 1998 |  |  |  |  |
| 1999 |  |  |  |  |
| 2000 | 37,781 (0.44) | 15,259 (0.21) |  |  |
| 2001 |  |  |  |  |
| 2002 | 16,845 (0.19) | 9,613 (0.19) | 4851 (0.44) | 553 (0.20) |
| 2003 |  |  |  |  |

Table 13.9. Estimated beginning year biomass and fishing mortality rates for BSAI rougheye and shortraker rockfish.

|  | Rougheye |  | Shortraker |  |
| :---: | ---: | :--- | ---: | :--- |
| Year | Biomass | F | Biomass | F |
| 1980 | 26227 | 0.014 | 38299 | 0.012 |
| 1981 | 25498 | 0.026 | 37156 | 0.015 |
| 1982 | 24513 | 0.013 | 36092 | 0.017 |
| 1983 | 23898 | 0.005 | 35120 | 0.007 |
| 1984 | 23499 | 0.005 | 34902 | 0.002 |
| 1985 | 23137 | 0.001 | 34356 | 0.001 |
| 1986 | 22860 | 0.003 | 33895 | 0.001 |
| 1987 | 23183 | 0.008 | 33172 | 0.002 |
| 1988 | 22565 | 0.009 | 32738 | 0.002 |
| 1989 | 21964 | 0.028 | 32329 | 0.006 |
| 1990 | 20950 | 0.105 | 31812 | 0.021 |
| 1991 | 18390 | 0.014 | 31145 | 0.015 |
| 1992 | 17386 | 0.073 | 30084 | 0.012 |
| 1993 | 16123 | 0.061 | 29639 | 0.014 |
| 1994 | 15157 | 0.052 | 29225 | 0.008 |
| 1995 | 14341 | 0.027 | 28813 | 0.007 |
| 1996 | 13949 | 0.064 | 28700 | 0.006 |
| 1997 | 13128 | 0.077 | 28559 | 0.005 |
| 1998 | 12113 | 0.044 | 30165 | 0.006 |
| 1999 | 11757 | 0.033 | 29348 | 0.006 |
| 2000 | 11541 | 0.023 | 28496 | 0.011 |
| 2001 | 12029 | 0.056 | 28929 | 0.005 |
| 2002 | 11243 | 0.024 | 27901 | 0.013 |
| 2003 | 10510 | 0.015 | 23748 | 0.008 |
| 2004 | 10379 |  | 23379 |  |



Figure 13.1. Survey biomass and estimated biomass of BSAI rougheye (a) and shortraker (b) rockfish.


Figure 13.2. Estimated fishing mortality rate of BSAI rougheye (solid line) and shortraker (dashed line) rockfish.


Figure 13.3. Annual surplus production and production model fits of BSAI rougheye (a) and shortraker rockfish (b).


[^0]:    Estimated removals through September 27, 2003.

[^1]:    * Observer data through October 17, 2003; total catch estimate through September 27, 2003

