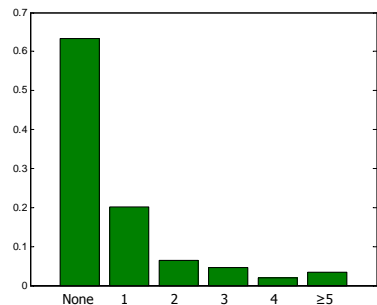


# Integrating toxicology and ecology in population-level risk assessment for wildlife: what data does your modeler really need?

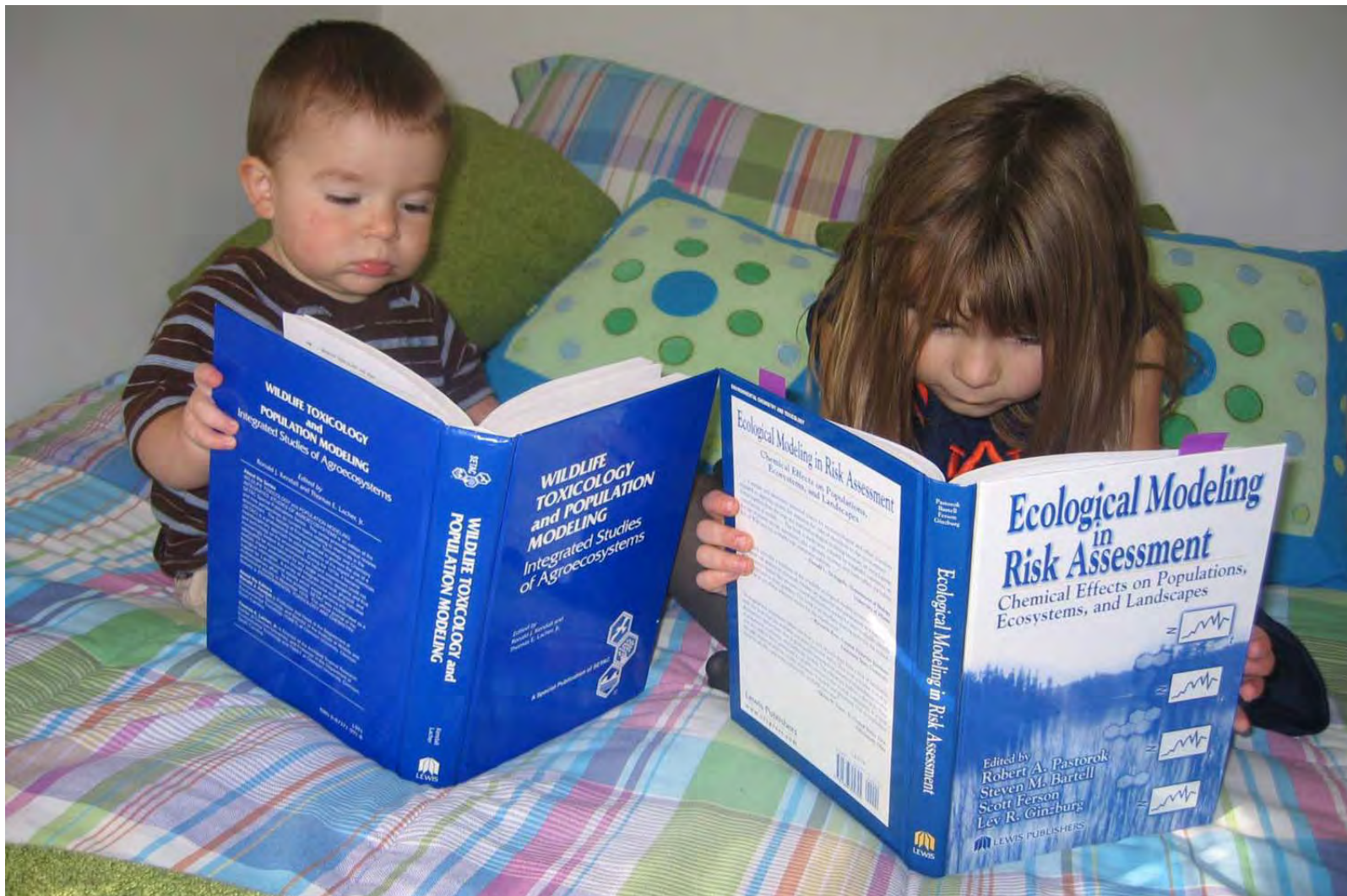
*Matthew A. Etterson and Richard S. Bennett*  
*USEPA/ORD/NHEERL*  
*Mid-Continent Ecology Division, Duluth, MN*



$$\begin{bmatrix}
 1 & 0 & 0 & 0 \\
 q_s & 0 & 1 - q_s & 0 \\
 0 & (1 - e)S & 0 & 1 - (1 - e)S \\
 q_f & 0 & 1 - q_f & 0
 \end{bmatrix}$$



SCIENCE



## What data does a population modeler really need?

1. Estimated vital rates (survival, fecundity) in the absence of the stressor
2. Estimated effects of the chemical stressor on vital rates

## How much data are there?

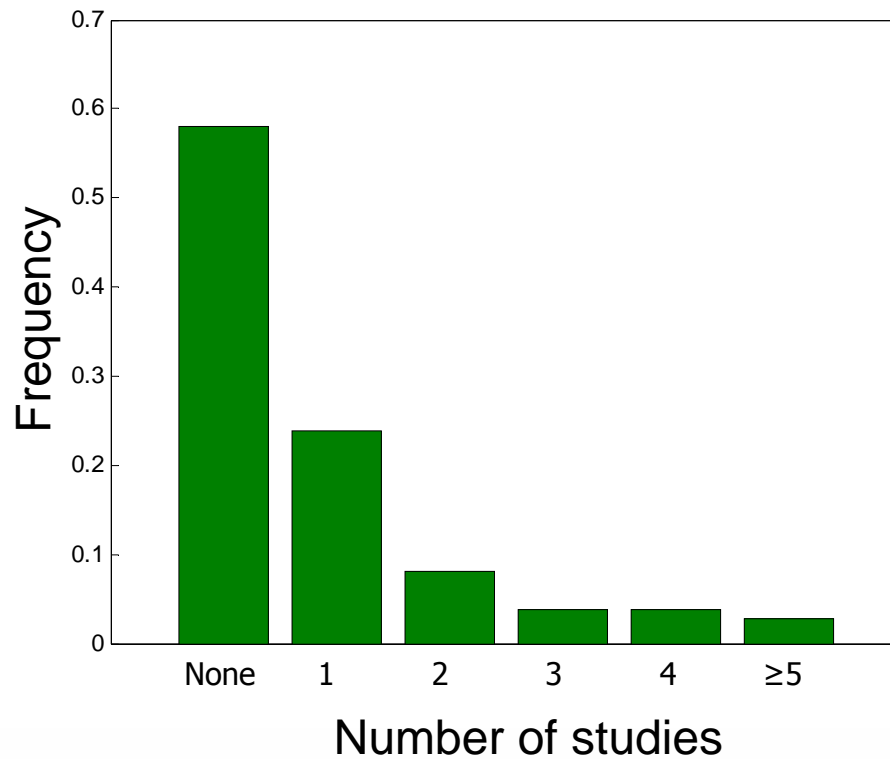
### **B**irds of **N**orth **A**merica

*Life Histories for the 21st Century*

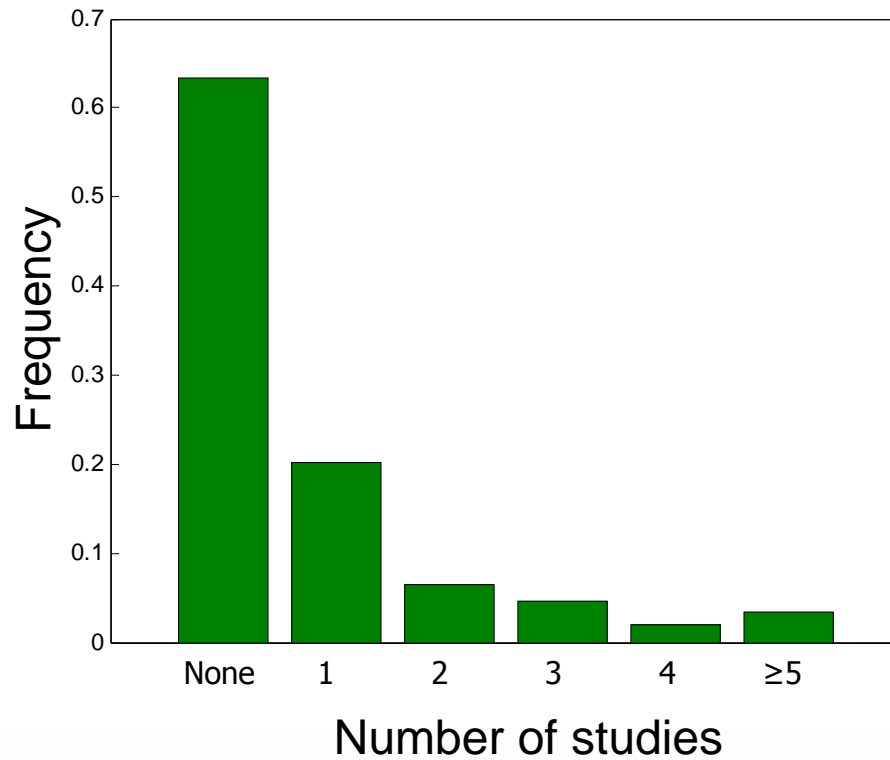


**Cornell Laboratory of Ornithology**  
**American Ornithologists' Union**

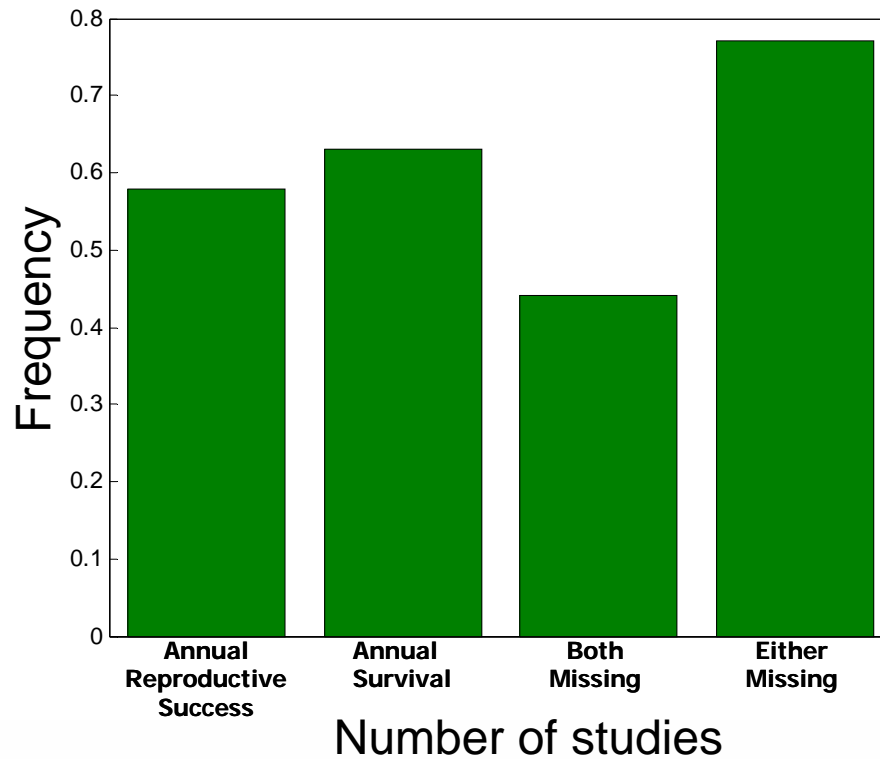
## Annual reproductive success



## Annual survival



## Frequency of “no data”



## A fundamental conundrum

- Estimating wildlife demographic data is costly

- Extrapolating demographic data is ~~totally bogus~~ ~~ill-advised~~ ~~somewhat questionable~~

*a bit tricky.*

~~dangerous~~

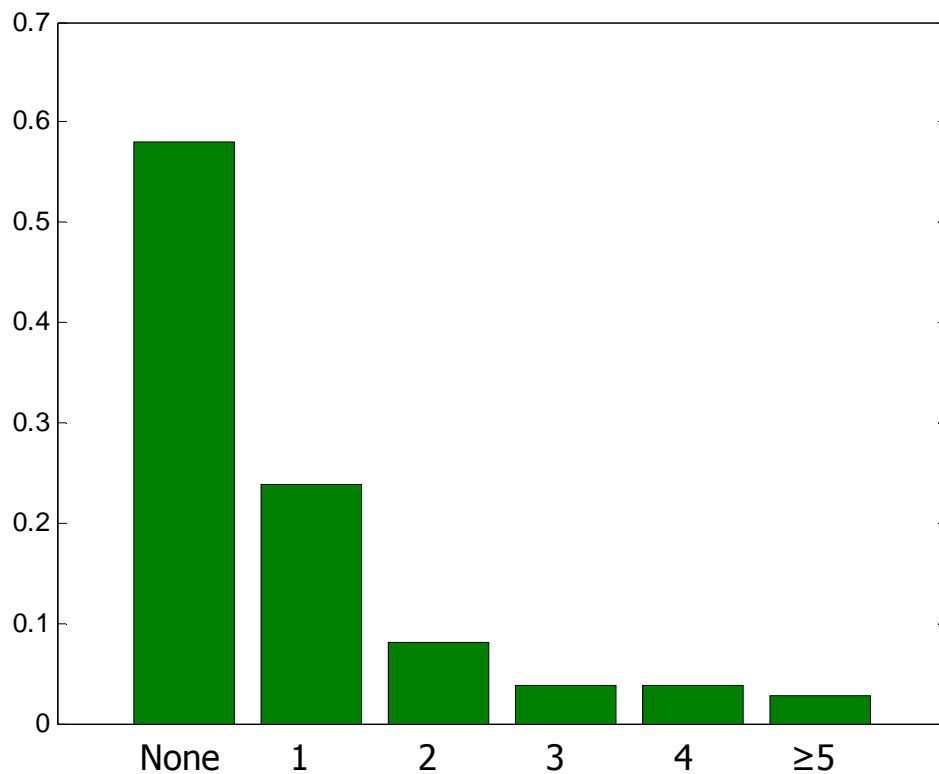


## Even if data exist, they may not be good enough...



## Eastern Meadowlark, a well-studied species?

Reference	Estimate
Kershner et al. 2004	$f$
Granfors et. al 1996	$f$
Granfors et. al 1996	$f$
Lanyon 1957	$f$
Michel et al. 2005	$\varphi$



## Sources of error in demographic data

- Sampling error & error propagation
- Currency
- Study design (“best habitat bias”)
- Assumptions underlying estimators
- Explanatory variables

## What can we learn from population-level risk assessment using these data?

- Predict future population size (forecasting)
- Project future population size (projection)
- Minimize conservative bias

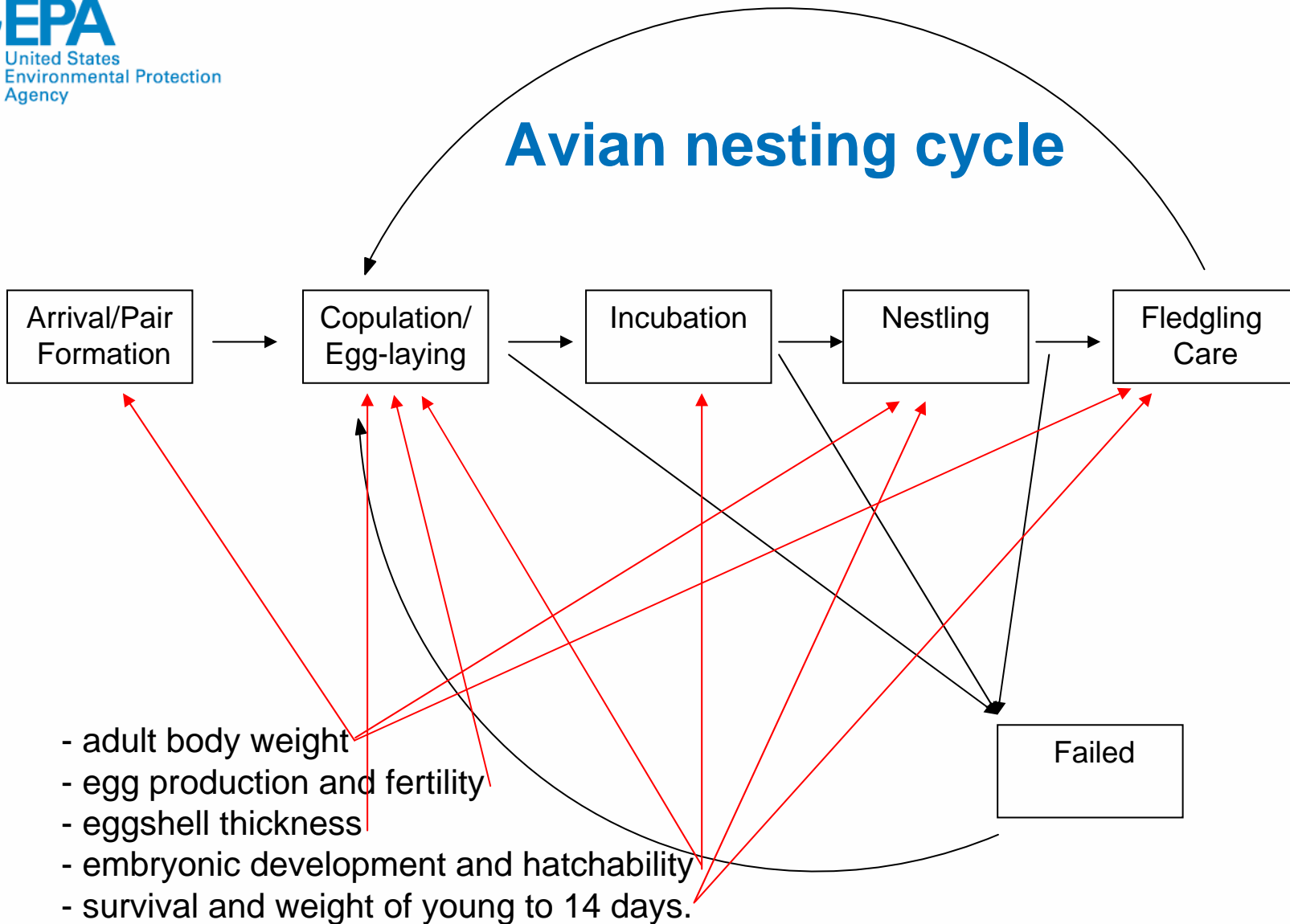
## Example: avian reproduction test

- Designed experiment using Mallards (*Anas platyrhynchos*) or Bobwhite (*Colinus virginianus*) that vary in dietary concentration of pesticide.
- Analysis of Variance used to determine highest dietary concentration at which no adverse effects are observed (NOAEC).
- Specific endpoints include:
  - adult body weight
  - egg production and fertility
  - eggshell thickness
  - embryonic development and hatchability
  - survival and weight of young to 14 days.
- NOAECs compared to estimated exposure (risk quotients)
- Core requirement for pesticide registration nationally and internationally

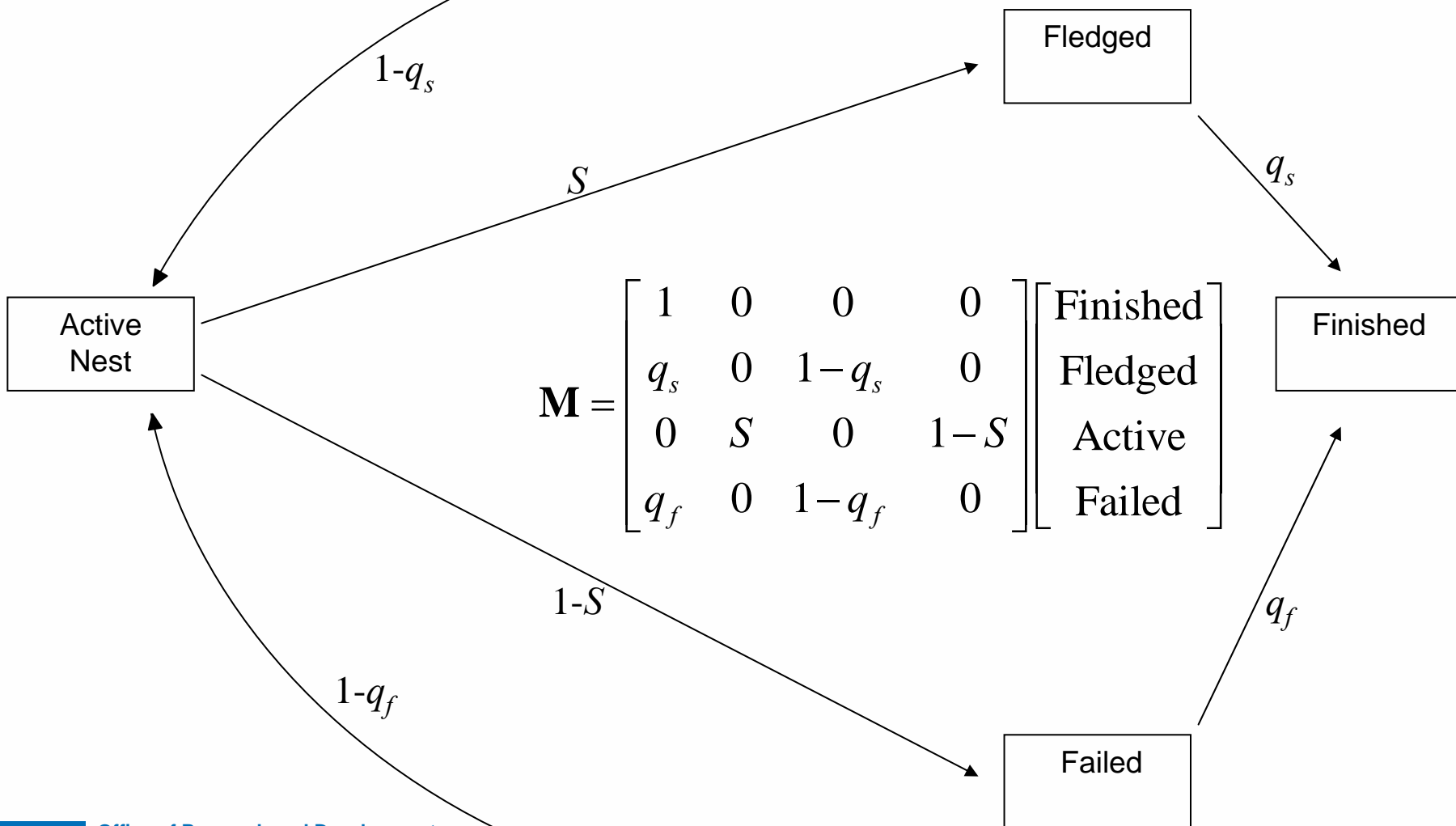
## Problems with avian reproduction test

- Test endpoints do not provide direct information on reduced annual reproductive success.
- Birds can renest after failure and/or success
- Avian reproduction test does not provide dose-response information

# Avian nesting cycle



# A Markov chain nest productivity model (MCnest)

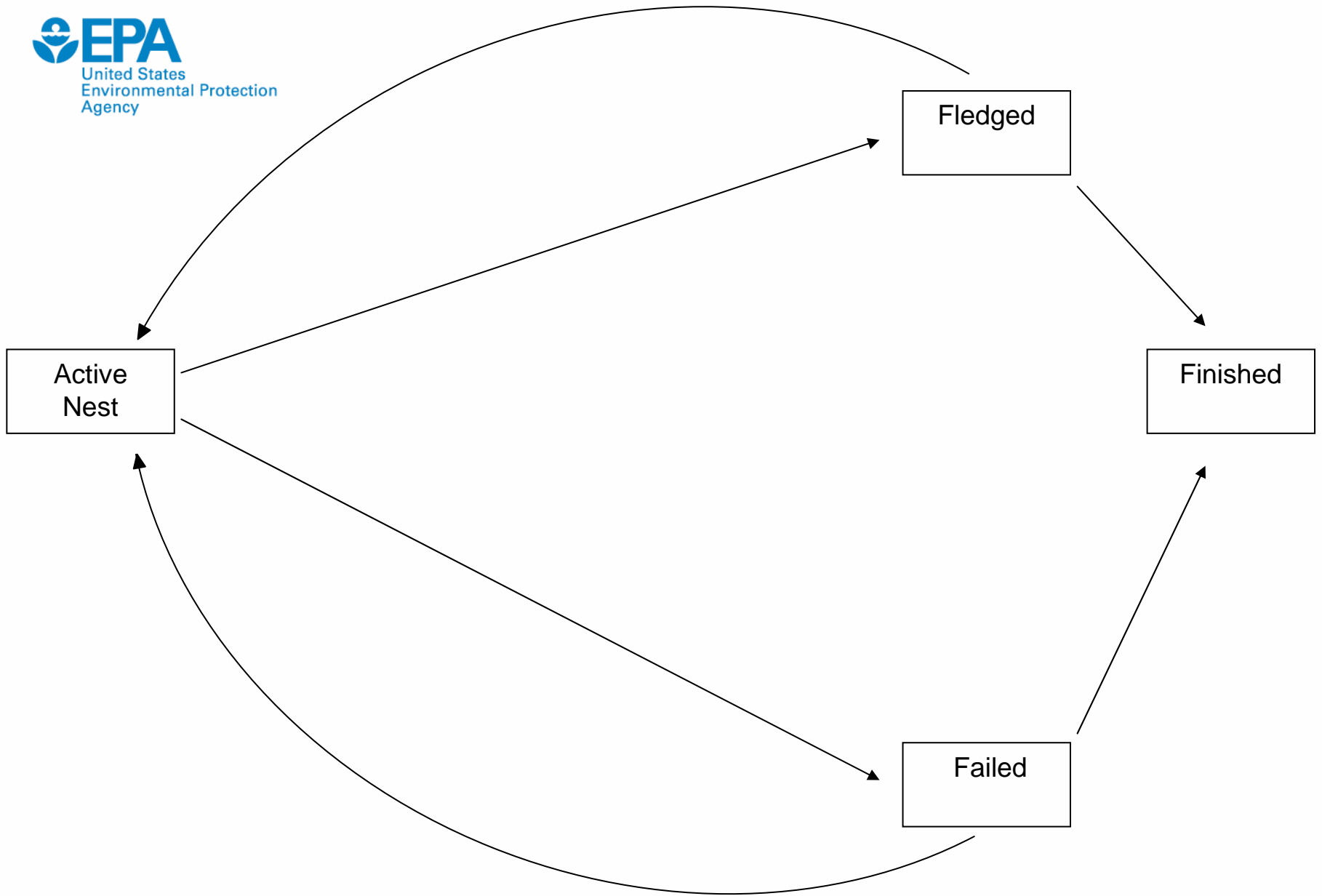




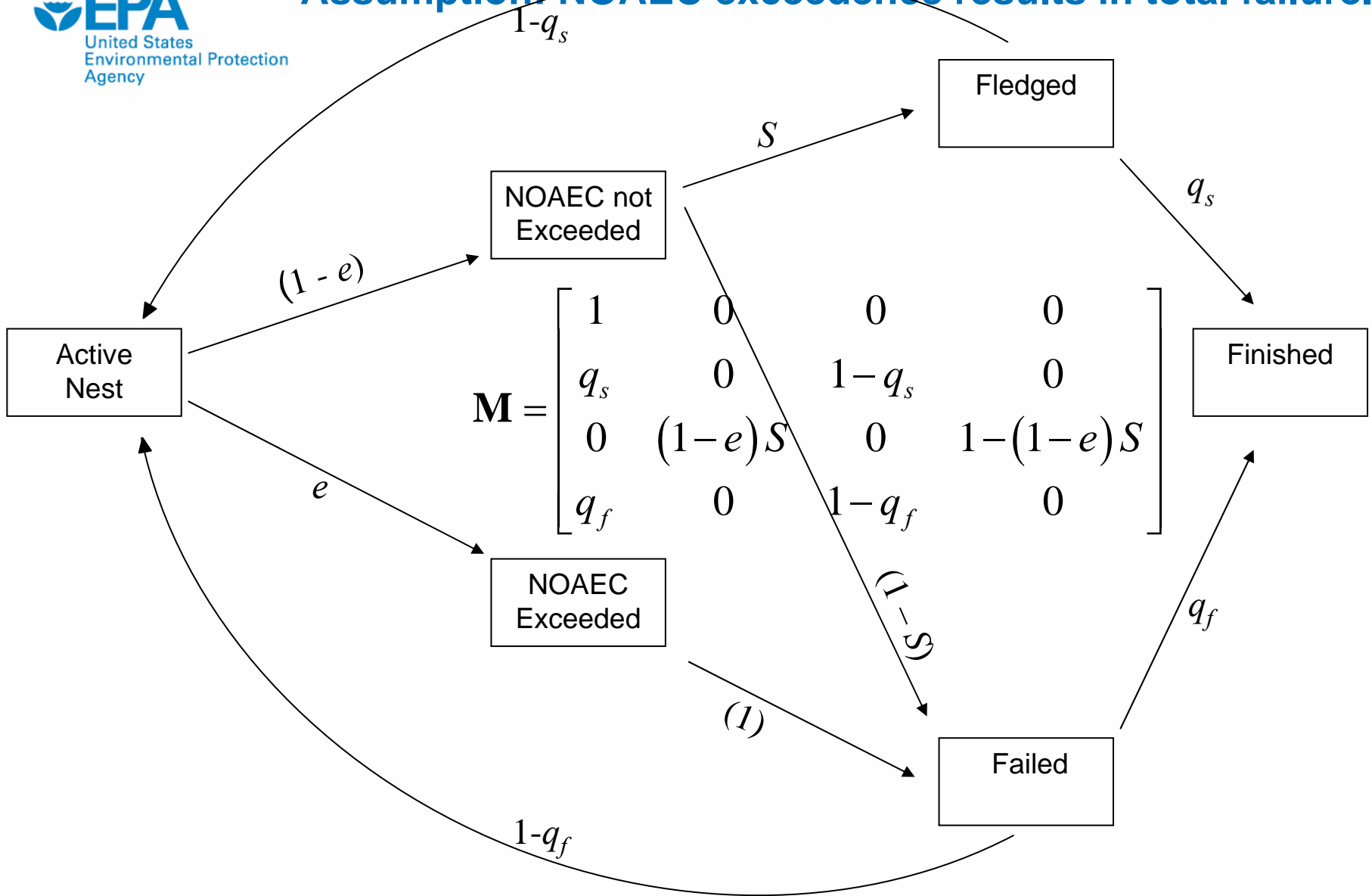
## Validation of the MC approach

- Is it really conservative to assume that an exceeded NOAEC results in total nest loss?
- When estimating effects on reproductive success can we ignore potential effects on survival?

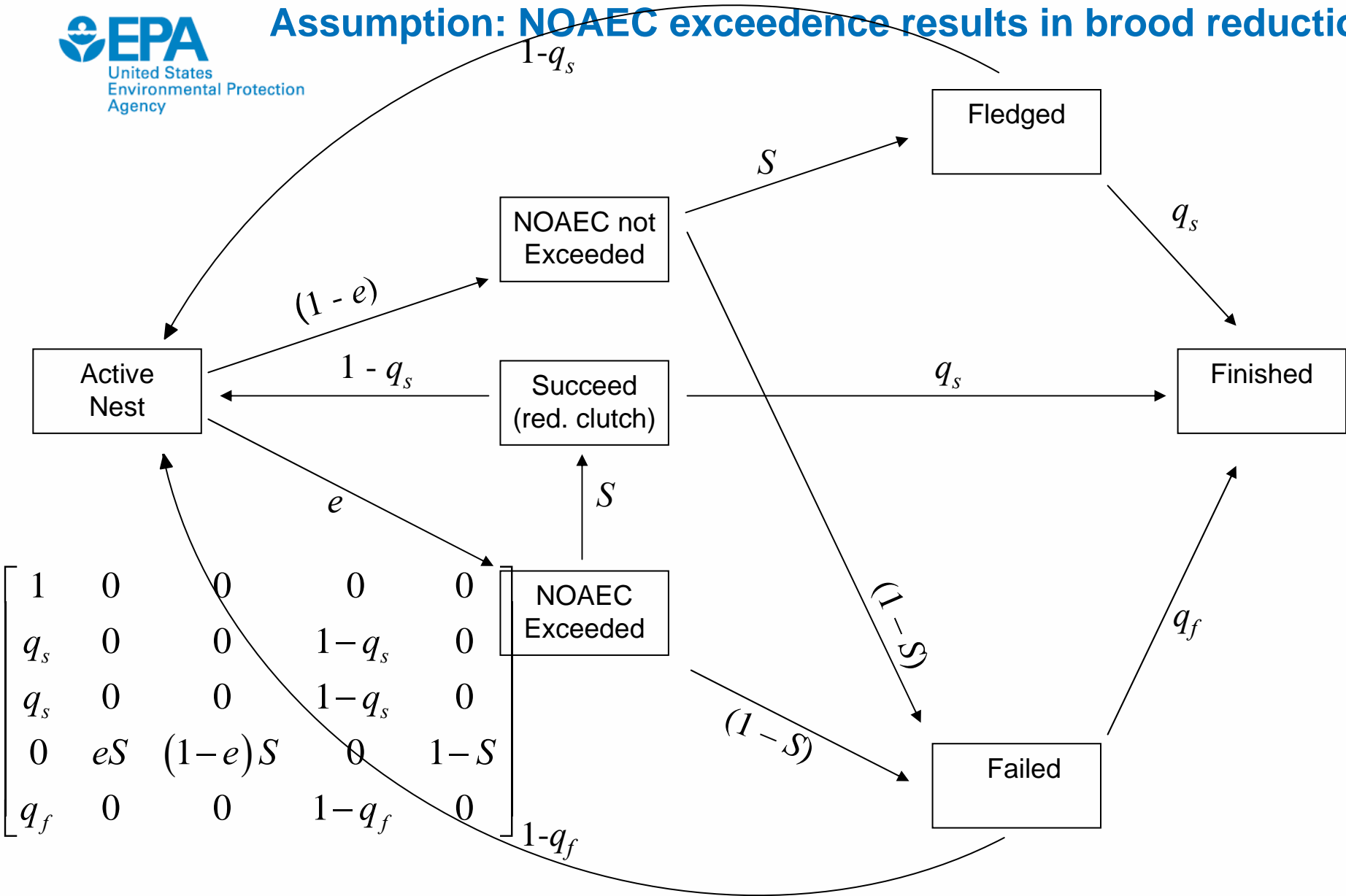
For more background on the Markov chain models see:  
Etterson and Bennett. 2005 Ecology  
Etterson and Bennett. 2006. Ecological Modelling  
Etterson et al. 2007 Studies in Avian Biology  
Etterson et al. 2007 Auk.



# Assumption: NOAEC exceedence results in total failure.



# Assumption: NOAEC exceedance results in brood reduction



1	0	0	0	0
$q_s$	0	0	$1 - q_s$	0
$q_s$	0	0	$1 - q_s$	0
0	$eS$	$(1 - e)S$	0	$1 - S$
$q_f$	0	0	$1 - q_f$	0

The total failure model is protective when:  $S > \frac{p}{(1-p)} \frac{1}{(1-e)} \frac{q_f}{(q_s - q_f)}$

What we can learn from this inequality?

1. We are more likely to be protective when:

- proportional brood reduction  $(1 - p)$  is large,
- exposure exceeding the NOAEC is *less* likely,
- $q_f$  is small,
- the difference between  $q_s$  and  $q_f$  is large and positive.

2. We can use it to classify life histories:

- $q_s > \approx q_f \rightarrow$  assuming total failure is never protective
- $q_s < q_f \rightarrow$  assuming total failure is always protective
- $q_s > q_f \rightarrow$  assuming total failure might be protective
- $q_s \gg q_f \rightarrow$  assuming total failure is probably protective

## Validation of the MC approach

- Is it really conservative to assume that an exceeded NOAEC results in total nest loss?
  - In most cases, yes, but it depends on life-history characteristics of the nest cycle
- When estimating effects on reproductive success can we ignore potential effects on survival?
  - No!** similar analyses suggest we cannot ignore survival effects, even when we only wish to estimate reproductive effects.

# Summary and conclusions

- Our ability to integrate effects estimates from molecular, cellular, and even individual levels to effects on populations is hampered by a chronic paucity of demographic data on most species of birds (wildlife in general?).
- Even when data exist, they are likely to be imprecisely estimated (and probably biased too).
- And even if you overcome the first two hurdles, data collected in lab or pen are likely to be expressed in a currency that disagrees with what is needed to parameterize a population model.
- Nevertheless, it is possible to make meaningful inference about chemical effects on populations using standard population modeling tools.
- Discussion about life-history traits and reconciliation of data across hierarchical scales is a critical component of model design for wildlife population level risk assessment.
- We believe the approach we have described may be useful for similar problems with other taxa for which laboratory-collected data must be used to make population-level inference