

Controls on Pollution Ozone Production Measurable from Surface, Aircraft, and Satellite Monitors

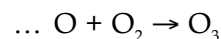
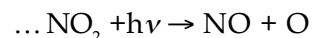
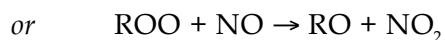
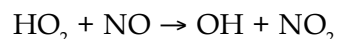
Robert B. Chatfield¹, Xinrong Ren², William H. Brune³, James H. Crawford⁴, Robert Esswein^{5,1}, Alan Fried⁶, Jennifer R. Olson⁴, Richard E. Shetter⁶

Widespread airborne sampling of HO₂ and more broadly measurable species over North America suggests empirical measures of regional ozone production. Two variables are important: the well-known $\nu = [\text{NO}]$, and $\phi = j_{\text{HCHO}} [\text{HCHO}]$, a gauge of organic oxidation rate. Ozone production functions as a semi-parametric fit $F_2(\phi, \nu)$, or a more theoretically based $Q_2(\phi, \nu)$, serve in different contexts. F_2 and Q_2 both give traditional empirical kinetic model assessment (EKMA) insights a measurable, local expression, and help define control strategies. At low $\phi \bullet \nu$, O₃ production is described as $f_1(\phi \bullet \nu)$, starting linear and dropping to sub-linear. We expect that similar ϕ, ν analysis may help separate O₃ transport from photochemistry, providing a language to express generalizations about complex kinetic systems and observations. We suggest: (i) a network of moderate-technology rural sampling stations; (ii) more broadly, analysis derived from remote sensing, mapping near-surface j_{HCHO} , HCHO and NO₂ (related to NO).

1. Dissecting Subcontinental Smog Ozone Production and Control

NASA's investigation of air pollution over the Eastern North America and its outflow into the Atlantic provided a remarkably broad sample of the polluted atmosphere during a moderate-pollution summer, July–August, 2004. Measurements made aboard the DC-8 flying laboratory during the INTEX-NA studies within ICARTT [Singh et al., 2007, Ren et al., 2008] yielded relationships utilizing measurements of only HCHO, NO, and UV that describe the chemical production rate of smog ozone with useful accuracy. Such formulas deserve further testing and application since the abatement of smog ozone has substantial economic value: health and crop productivity concerns must be related to the costs of restricting emission of ozone precursors from a wide variety of sources which may be kilometers or thousands of kilometers upwind. [NAAMS Draft, 2005, Solomon, Seinfeld, 1998, Jacobson, 1998, McKeen et al., 1991] For both forecasting and control, local ozone production must be distinguished from transported ozone [NAAMS Draft, 2005, Chatfield and Delany, 1990, McKeen et al., 1991, Sillman et al., 2002]. It is highly desirable to have broad maps of such indicators for the ozone production process. Space-borne observations could give broad and informative maps.

Gross production of tropospheric ozone P_{O_3} , according to established theory, can be written the consequence of either of two peroxy radical reactions [Seinfeld and Pandis, 1998, Jacobson, 1998], contributing to a broader class of chemically interchanging oxidants (NO₂, NO, O₃) whose predominant species is ozone:



43 and the rate is determined by $P_{O_3} = k_H[HO_2][NO] + k_R \sum_i [ROO]_i [NO]$, omitting very minor
44 pathways. Here, ROO stands for one of many organic radicals with a C-O-O peroxy
45 structure. We call P_{O_3} directly computed, but admit, in fact, there are implicit model
46 assumptions about reaction rates and products. (E.g., the rate k_R describes reactions
47 producing not peroxy nitrates but NO_2 ; relative yields may have uncertainties.) As we shall
48 see, the HO_2 reaction usually contributes 50–70% of P_{O_3} , depending somewhat on H/C ratio
49 and structure of the smoggy volatile organic compound (VOC) mix. Consequently we refer to
50 $P_{O_3}^* = k_H[HO_2][NO]$ as the “principal ozone production rate,” a rate unambiguously
51 measured. The residual ($P_{O_3} - P_{O_3}^*$) due to ROO radicals is intrinsically highly correlated to $P_{O_3}^*$,
52 since reactions producing ROO’s involve the creation of HO_2 ’s and vice versa [Seinfeld and
53 Pandis, 1998, Jacobson, 1998, Saunders et al., 2003]. The creation of ROO’s and HO_2 ’s
54 involves complex smog chemistry extending to tens of thousands of reactions; rate coefficients
55 for these must be estimated from limited data, as must the yield when more than one product
56 is possible. Consequently, measures of $P_{O_3}^*$ may provide useful analytical tools and guidance.
57 (While chemical loss and deposition of O_3 can pose difficulties in certain important situations
58 [Carroll et al., 2001], they are separable and beyond the current analysis.)

59 Some previous empirical analyses on air mass ozone production, emphasizing
60 P_{O_3} and $P_{O_3}^*$ chemical production have utilized coordinate production of other oxidation
61 products such as HNO_3 and H_2O_2 , while admitting that inferences are influenced by rapid and
62 perhaps sporadic dry and wet removal of these species [Sillman et al., 2002]. In any case
63 these usefully emphasize the entire history of chemical ozone net production from a major
64 emission event (e.g., an urban or regional plume) rather than a local view more appropriate to
65 the formulation of more localized control measures. Thornton et al. [2002] make a local
66 analysis employing a full suite of instrumentation. Kleinman’s [Kleinman, 2005a, Kleinman et
67 al., 2005b] analyses employ kinetic simulations of VOC/ NO_x mixtures; nevertheless the
68 themes and analyses are ones that this work emulates. The twin complexities of complex
69 reaction pathways and difficult-to-simulate transport argue for broad space-borne
70 observations augmented by widespread, if scattered, rural observations.

71 Frequent measurements on the DC-8 of boundary layer HO_2 , NO , and other variables
72 allowed this principal ozone production rate to be calculated. The rich variety of
73 measurements made on board [Singh et al., 2007] also allowed an approximate estimation of
74 the ROO radicals through chemical simulations [Olson et al., 2001]. Significantly, we find that
75 ozone production estimated with high-technology radical measurements can be
76 approximated well with less demanding, expensive instruments measuring just two smog
77 precursors. Nearly a thousand smoggy boundary-layer mixtures were measured below 1300
78 m and extended from Texas to Ontario to Boston to Atlanta, and into the Western Atlantic
79 (see Supplemental Material).

80 We begin by highlighting two important controlling dimensions: first, $\nu = [NO]$, the
81 easily measured component of NO_x which which is well known for its role in converting
82 peroxies to ozone; then, second $\phi = j_{HCHO} [HCHO]$. We associate ϕ with the reactive flux of
83 organic species that simultaneously create and destroy formaldehyde while producing peroxy

84 radicals; the crux of this work is to initiate study of the degree of useful guidance which ϕ
85 provides. The abstract names ϕ and ν and are meant to suggest a decomposition of the smog
86 production problem into a peroxy-determining reaction-flux of VOC and an efficiency of
87 conversion of the peroxies to ozone. ϕ is thought important not just because HCHO
88 photolysis is an important source of radicals itself, even though as much as 25–30% of the
89 radicals originate from HCHO reactions in one urban study [Lee et al., 1998]. Rather ϕ
90 embodies many correlations with the HO₂-producing activity depending on the action of hard
91 UV on organics which are more difficult to measure. HCHO is common at the end of known
92 VOC oxidation chains. Glyoxal and methyl glyoxal should be lesser rivals as end-products
93 and indicators, mostly for mixtures rich in lower H/C ratio species [Saunders et al., 2003 and
94 work referenced there]. Extended work employing glyoxal and photolysis rates similarly
95 cannot be explored here. The j_{HCHO} [HCHO] indicator may also usefully parallel other radical
96 production pathways involving, e.g., O₃ photolysis producing OH and other VOC reactions.
97 At the simplest level, it converts HCHO measurements to one VOC-processing rate;
98 counterbalancing sources and sinks can leave substantial quantities of HCHO in low-radical
99 situations, e.g. at night.

100 2. Interpreting Interactions of Radical-Producing and Radical-Converting Variables

101 Figure 1a provides a contour plot of an empirically estimated function F_2 , allowing
102 independent fitting to many combinations of variables ϕ and $\nu = [\text{NO}]$. The supplement
103 describes the estimation; variance explained was ~82%. When F_2 is plotted on linear axes
104 (Figure 1a), the shape of the contours is similar to the “EKMA” Empirical Kinetic Modeling
105 Approach method used to address a similar O₃ problem: EKMA related morningtime [NO_x]
106 and VOC concentrations to O₃ produced 6-8 hours later, ignoring transport. The method
107 improved over previous “linear rollback” abatement strategies since it recognized complex
108 VOC/NO_x interaction and the sub-linearity of ozone production; however, it was eclipsed by
109 three-dimensional transport models [Shafer and Seinfeld, 1985, U.S. EPA, 1987]. F_2 is by
110 contrast a local measure, approximately defined by a distance scale that recognizes a 1-3
111 hour middle-of-the-day turnover time of HCHO. It lacks features of EKMA, most notably a
112 large triangular “notch” of low ozone in the upper left hand corner of EKMA plots. High
113 morning-time NO primarily lowered later O₃ by the need for NO+O₃ to react to NO₂; other
114 multi-hour effects (e.g., the OH+NO₂ reaction eventually lowering NO_x) also contribute
115 somewhat. We believe that F_2 builds on the analyses of Kleinman for nearer-source
116 descriptions of VOC and NO effects based on measurable VOC's, and also of Martin, using
117 satellite analyses of boundary layer NO₂ and HCHO to suggest somewhat more
118 quantitatively broad-area VOC vs NO_x control [Martin et al., 2004]. The greatest reduction in
119 ozone production is achieved by moving downwards perpendicular to the contour lines; of
120 course, tonnage-restriction of NO and tonnage removal of VOC's that reduce ϕ may have
121 different costs and imply a different strategy.

122 Figure 1a suggests that F_2 contours are nearly simple contours of $\phi \cdot \nu = \text{constant}$. An F_2
123 plot using logarithmic scales in Figure 1b shows this near-hyperbolic behavior more cleanly,
124 and emphasizes the strong role of a single variable, $a = \phi \cdot \nu$. Departures from this behavior

125 (e.g., additional VOC effects) are indicated by deviations from lines such as the dashed red
126 lines in Figure 1a and Figure 1b. Control strategies implied by the log-log plots are slightly
127 different from the linear plot: to move straight downwards (most effective control),
128 *proportional* reductions are implied, not equal-tonnage reductions in emissions.

129 These observations naturally suggested that an even simpler representation of $P_{O_3}^*$ was
130 available as an $f_1(\phi \cdot \nu)$, that is $f_1(j_{HCHO} [HCHO] [NO])$. This was determined by a spline fit
131 emphasizing simplicity; f_1 is shown in Figure 2a; it also emphasizes the less-than-linear
132 production of O_3 at high $\phi \cdot \nu$ in Figure 1a. This is a major feature resembling the EKMA
133 description. We present the relationship as *statistically* justified [Venables and Ripley, 2004]
134 over a very large diverse sample of smog ozone production rates of less than ~ 4 ppb hr^{-1} ,
135 and, suggest further validation in more varied circumstances.

136 The f_1 relationship is simply a more complete quantification of earlier understandings
137 emphasizing the roles of NO and UV radiation [Haagen-Smit, 1952, Leighton, 1961, Lin et al.
138 1988, Liu et al. 1987]. We can quantify the advance of understanding: When we estimated
139 with similar procedures a the sequence of best estimates using progressively more
140 information, which we will label $f_0([NO])$, $f_{00}(j_{HCHO} [NO])$ and $f_1(j_{HCHO} [HCHO] [NO])$
141 the variance explained rose sharply: <50–60%, <63–69% and 79%. $f_0([NO])$, $f_{00}(j_{HCHO} [NO])$
142 also suggest even more complex functional behavior at values of $\phi \cdot \nu$ higher than 0.25
143 ppb²/hr. Kleinman [2005a] describes a power law based on those VOC's which can be
144 measured (see Supplement).

145 We would expect the HCHO-based relationships to become applicable over time
146 periods of 1–3 hours corresponding to its midday lifetime, as VOC oxidation and NO_x effects
147 adjust. There needs be time for any fresh NO to come into the Leighton [1961] quasi-steady
148 state with O₃ and NO₂, at least ~ 300 seconds from mid-morning to mid-afternoon. More
149 stringently, HCHO needs to come into a quasi-steady state with its source and product
150 compounds, typically ~ 2 hours around midday, but ~ 12 hours at dawn or dusk.
151 Additionally, strong sources of freshly emitted VOC and HCHO mixtures may need to adjust
152 to build HCHO from VOC's or consume primary HCHO.)
153

154 3. Comparison to More Theoretically Based Methods

155 A somewhat more theoretical description of ozone production informing these
156 statistical results is available by equating the production and loss rates of the transient HO₂
157 radical. The quasi-steady-state approximation for HO₂ suggests that

$$158 \quad 2\eta k_{HH} [HO_2]^2 + (1-\rho)k_{NH}[NO] [HO_2] = \gamma j_{HCHO}[HCHO]$$

159 The supplement describes the kinetic rate coefficients and some additional parameters
160 (η, ρ , and γ), parameters necessary to address some complexities contributed by statistically
161 correlated reactions directly affecting HO₂. Briefly, γ summarizes related HO₂ (new HO_x)
162 production pathways, η summarizes peroxy-peroxy destruction reactions strongly correlated
163 to HO₂+ HO₂, and ρ summarizes radical recycling producing HO₂ as some fraction of the
164 HO₂+NO destruction reaction. Each parameter was necessary to produce reasonable graphs
165 of HO₂ as a function of NO [Ren et al., 2008] although the set (η, ρ , and γ) clearly exhibits one

166 linear dependency: production and destruction rates of HO₂ can offset each other and still
 167 yield the same solution. This is a two-parameter fit which is limited to the same observable
 168 Quantities. Solving the quadratic to obtain HO₂ and thus the ozone production rate, we have
 169
$$P_{O_3}^{\circ} = Q_2(\phi, \nu) = k_{NH} \gamma (j_{\text{rads}} [\text{HCHO}] [\text{NO}]) / \left[(1 - \rho) k_{NH} [\text{NO}] - \sqrt{(1 - \rho)^2 k_{NH}^2 [\text{NO}]^2 + 4 k_{NH} \gamma j_{\text{rads}} [\text{HCHO}]} \right]$$

 170 The fundamental role of $\phi \cdot \nu$ is clear from the numerator, and so is a degree of less-than-linear
 171 response indicated by f_1 as both radical production ϕ and ν increase in progressively more
 172 polluted situations. The form of the response behavior at higher $\phi \cdot \nu$ is less clear, as is the
 173 approximate $\alpha^{0.4}$ behavior at values of $\phi \cdot \nu$ higher than 0.05 in Figure 2a. Figure 3a shows that
 174 ozone production using this formula with observed data displays much the same form as f_1 .
 175 This two-parameter fit exhibits a lower r^2 , 0.75, vs. 0.79 for the completely empirical plots. f_1
 176 and F_2 appear more successful than Q_2 for low-VOC, high NO situations, reflecting observed
 177 HO₂ behavior that is not fully understood [Ren et al., 2008]. Why do the empirical
 178 relationships emphasize higher $P_{O_3}^{\circ}$ at low ν / high ϕ and even more strongly at high ν / low ϕ ?
 179 Are there actually variations in ρ or η ? The quadratic method may deserve elaboration with
 180 broader datasets. Analyses of the family HO_x = HO₂ + OH could address these questions
 181 [Thornton et al., 2002], but we found no way to describe HO_x with just a few simple
 182 measurements.

183 In order to complete the description of ozone production, some estimate of the
 184 relationship of total production P_{O_3} to $P_{O_3}^{\circ}$ is required. While measures of total ROO+HO₂
 185 radicals are steadily being improved, their interpretation goes beyond this brief report.
 186 Mechanisms of photochemical reaction based on laboratory measurements can be used to
 187 relate these quantities; two methods are available. One is the use of complete three-
 188 dimensional models with carefully checked emissions of organics. While useful, errors in
 189 transport and emissions as well as necessarily limited chemical mechanisms can affect results.
 190 "Point" models as employed are closely tied to the observations made on the DC-8 but must
 191 approximate complex chemical history with a simple diurnally repeating steady state [Olson
 192 et al., 2001. Limited VOC measurements also restrict the number of cases to ~280, suggesting
 193 point-model effects are more approximate. We found for this dataset that $P_{O_3} = 1.6 \times P_{O_3}^{\circ}$, as
 194 shown in Figure 2b. This ratio is expected to vary somewhat, between 1.3 and 2, with the
 195 chemical composition of the reacting VOC mixture; most basically high C/H ratios in the
 196 VOC favor a higher ratio. In Figure 3b, contour plots based on the point-model calculations of
 197 ϕ and ν appear roughly intermediate between the Q_2 and F_2 contour plots, suggesting
 198 methods to critique each.

199

200 **4. Conclusions: Extend Sampling to Rural Sites and to Space**

201 Simulations with organic reaction mechanisms for various compounds suggest that
 202 there should be variations for (a) different organic mixes, some making for higher j_{HCHO}
 203 [HCHO], some making for less, and that (b) HCHO may not be exactly in an evolving "quasi
 204 steady state" so that j_{HCHO} [HCHO] uniformly represents peroxy radical levels (HCHO may
 205 be increasing or decreasing significantly enough to affect the analysis). Both of these situations
 206 will occur closer to pollution sources, which are more likely to have idiosyncratic compositions

207 (e.g., refinery emissions, aromatic emissions, ...) and also to be highly reactive and transient.
208 The widespread rural relationship described by our samples may not apply with the very
209 same parameters. Still, the relationships might be broadened to allow insights and methods
210 which separate HCHO-to-radical relationships from broader effects describing the intensity
211 and the NO_x-sensitivity of VOC oxidation activity, as suggested by forms like F_2 , Q_2 , and f_1 .

212 The results we analyze from widespread extra-urban sampling seem to suggest that the
213 formaldehyde oxidation rate (j_{HCHO} [HCHO]) is particularly able to define the oxidation rate
214 of all organics (more exactly, the concentrations of peroxy and organic-peroxy radicals). We
215 suggest study of the usefulness of this technique in at least two directions. One is to make
216 widespread characterizations of rural sites so as to describe regional smog, describing
217 unsuspected ozone production hotspot situations and their VOC or NO_x sensitivity. Step-by-
218 step, measurements and analysis can approach nearer-source air chemistry. The approach can
219 also be used for a three-dimensional model improvement tool, validating the calculation of
220 [HO₂] and consequently [ROO]. The general description of ozone chemical production rate
221 can be used to highlight and define situations in which such models (e.g., the Community
222 Model for Air Quality) may overestimate, or possibly miss important peroxy radical sources.
223 We have initiated analyses in situations in which [HO₂] was measured, and find it useful;
224 however sites and situations away from strong sources and sinks (NO emissions, deposition,
225 etc) are preferred. Rural, tall-tower measurements should be most informative about regional
226 chemistry, even plume chemistry.

227 Finally, components defining ϕ and ν , lowermost-troposphere HCHO, NO₂ and UV
228 radiation are currently measured with satellite technology. It will be helpful to create a very
229 empirical description of ozone production, using these variables, and deducing NO from NO₂
230 and lower-tropospheric O₃. We do not elaborate, but note that this builds on current work
231 [Martin et al., 2004]. With modest investment, capabilities for lowermost troposphere HCHO
232 and O₃ measurements may steadily improve, for example with new technology exploiting the
233 3- μm region; boundary layer remains the most difficult observation [Kumer et al., 2006]. More
234 quantitatively, the maps of integrated-column j_{HCHO} [HCHO] and NO₂ variables can closely
235 constrain three-dimensional simulation models. Maps of f_1 , F_2 , or Q_2 , can specify regions
236 most deserving of improvement in emissions or simulation. Remote sensing measurements
237 can then constrain both the NO availability and the HO₂ production rates in studies using
238 complex three-dimensional models. Because the information we have added to prior
239 knowledge refers to organic radical reactivity and especially peroxy radicals, we have nick-
240 named the method "POGO-FAN," Production of Ozone by Gauging of (organic) Oxidation -
241 Formaldehyde Activity and Nitric oxide. The name recognizes, in an American idiom, a
242 collective responsibility to limit ozone pollution for our own good.

243
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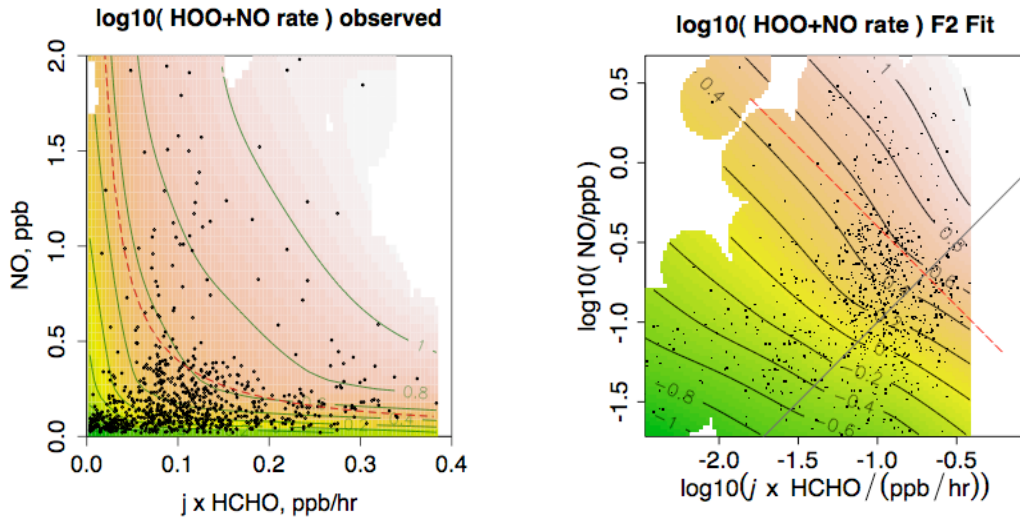
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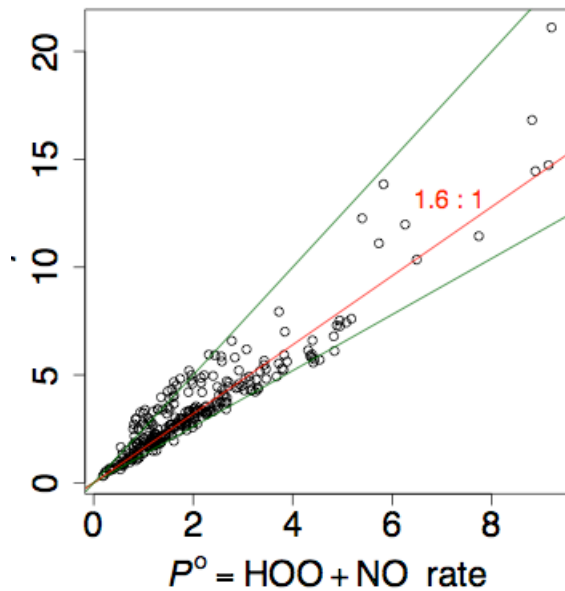
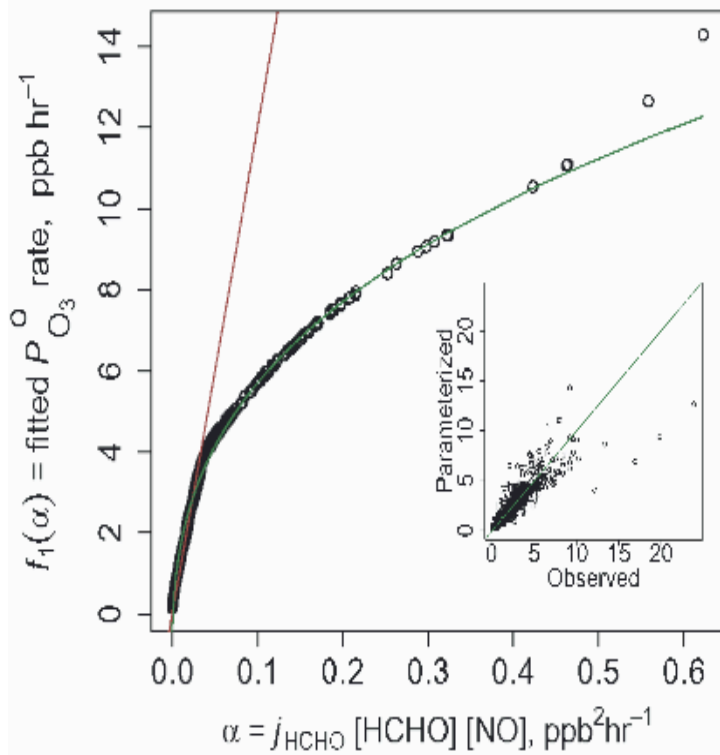


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A

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Fig. 1. (a) Statistical-fit F_2 of principal production rate of ozone P_{O_3} as a function of the product (j_{HCHO} [HCHO] and of [NO] with datapoints shown. Large variations of ozone production are shown as contours of $\log(P_{O_3} / (\text{ppb}/\text{day}))$. (b) Logarithmic axes are employed for the same information in second panel. 1:1 line shown. Red hyperbola and 1:1 and 1:-1 (slope) straight lines help define isopleths of $\phi \cdot v$. Most effective control strategies for O_3 for any particular mix at any data point are downward normals to the contours shown.

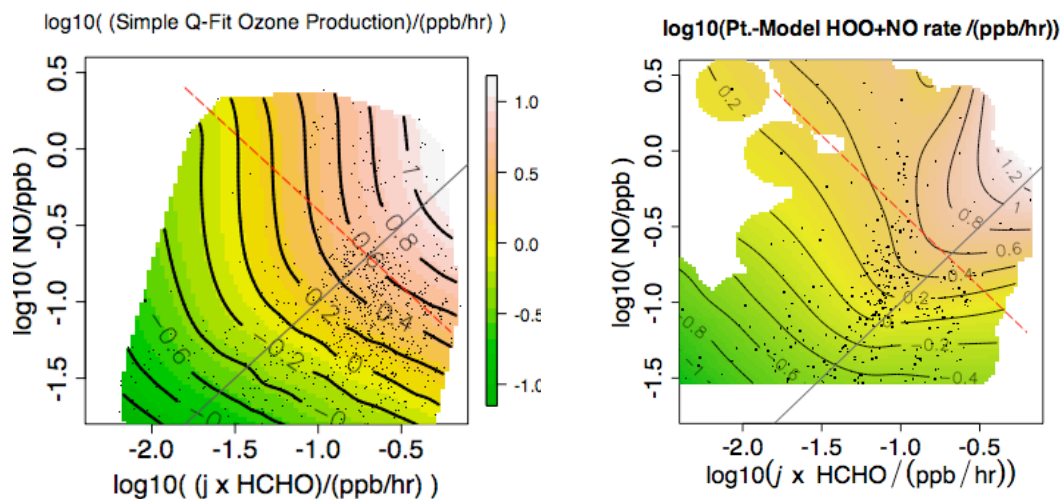


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Fig. 2. Statistical-fit f_1 of principal production rate of ozone P_3^o as a function of the product $\alpha = \phi \cdot v = j_{\text{HCHO}} [\text{HCHO}] [\text{NO}]$ which summarizes several smog processes. Small circles show individual data instances with modeled response; line shows interpolation and extrapolation of the fit. Inset shows correlation of the fit with direct computations of P_3^o using measurements made in INTEX-A flights over much of North America. Approximating red line has slopes of 120 and the green curve is the approximation $16 \alpha^{0.37-1.3}$. Correlation plot is inset. Correlation is 0.88, but rises to 0.91 if one high and four low points are dropped. (b) Relationship of total and primary (HO_2) ozone production rates, using box-model results (including measured VOC's) for each data point sampled on the DC-8.

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Fig. 3. (a) Two-parameter fit for ozone production Q_2 based on an empirical reaction-kinetic model of $[\text{HO}_2]$. 1:1 and 1:-1 lines are as in Figure 1b. Behavior for high NO and low ϕ differs significantly from F_2 . (b) Similar two-dimensional relationship derived from point modeling of kinetics, with behavior somewhat intermediate but differing from both F_2 and Q_2 at high ϕ , low ν . Note: Color shadings are not uniform from one contour graph to another.

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314 Affiliations

315 (1) Robert B. Chatfield: NASA Ames, Moffett Field, CA, USA Robert.B.Chatfield@nasa.gov

316 (2) Xinrong Ren, Rosenstiel School, U. Miami, Miami, FL, USA

317 (3) William H. Brune, Pennsylvania State U., State College, PA, USA

318 (4) James H. Crawford, NASA Langley, Hampton, VA, USA

319 (5,1) Robert Esswein, BAER Institute and NASA Ames, Moffett Field CA, USA

320 (6) Alan Fried, National Center for Atmospheric Research, Boulder, CO, USA

321 (4) Jennifer R. Olson, NASA Langley, Hampton, VA, USA

322 (6) Richard E. Shetter, National Center for Atmospheric Research, Boulder, CO, USA