EPA/ROD/R04-99/069 1999

EPA Superfund Record of Decision:

HOMESTEAD AIR FORCE BASE EPA ID: FL7570024037 OU 05 HOMESTEAD AIR FORCE BASE, FL 09/29/1999

Homestead Air Reserve Base, Florida

Final

Record of Decision for Operable Unit No. 5, Site WP-1 Electroplating Waste Disposal Area

June 1997



UNITED STATES ENVIRONMENTAL PROTECTION AGENCY REGION 4

ATLANTA FEDERAL CENTER 61 FORSYTH STREET ATLANTA. GEORGIA 30303-8960

Sep 29, 1999

<u>CERTIFIED MAIL</u> <u>RETURN RECEIPT REQUESTED</u>

4WD-FFB

Maj Gen. David R Smith Vice Commander, AFRC/CV 155 Second Street Robins AFB, GA 31098-1635

SUBJ Record Of Decision - Operable Unit 5 Homestead Air Force Base NPL Site Homestead. Florida

Dear Maj Gen. Smith,

The U.S. Environmental Protection Agency (EPA) Region IV has reviewed the subject decision document and concurs with the selected remedy for the remedial action at Operable Unit (OU) 5 at the former Homestead Air Force Base (HAFB). This remedy is supported by the previously completed Remedial Investigation, Feasibility Study, and Baseline Risk Assessment Reports. The selected remedy consists Land Use Controls which include:

- Posting of signs
- Restriction on construction
- Notify workers before they excavate
- Obtain FDEP and EPA approval prior to construction design
- Restrict groundwater access
- No water supply wells within restricted area
- No residential usage
- "Dig permit" required prior to construction
- Conduct inspections and correct discrepancies

The determination to implement this course of action at this site is consistent with the Comprehensive Environmental Response, Compensation, and Liability Act (CERCLA) as amended by the Superfund Amendments and Reauthorization Act (SARA) and the National Contingency Plan (40 CFR 300).

Concurrence with the Record of Decision (ROD) is conditioned on the express understanding that the Air Force is committed to the agreement reached with EPA Region IV and the Florida Department of Environmental Protection (FDEP) that complies with EPA's April 21, 1998 Memorandum titled "Assuring Land Use Controls at Federal Facilities" We reiterate as we advised Air Force Regional Environmental Office representatives in our meeting on May 21, 1998, our concurrence with this particular ROD is based on the understanding that the Air Force is committed to the Memorandum of Agreement (MOA) consistent with the above-referenced Land Use Control (LUC) Policy. Furthermore, the Homestead Air Force Base BRAC Cleanup Team (BCT) will be expected to craft specific provisions for Land Use Controls as part of the resulting Land Use Control Implementation Plan for OU- 5, that will prohibit residential land use.

EPA appreciates the level of effort that was put forth in the documents leading to this decision. EPA looks forward to working with HAFB as we move towards final cleanup of the National Priorities List (NPL) site.

If you have any questions, please call me at (404) 562-8651, or Doyle T. Brittain at (404) 562-8549.

Sincerely,

Richard D. Green, Director Waste Management Division

cc: Thomas J. Bartol, HAFB/AFBCA John Mitchell, HAFB/AFRES Jim Woolford, EPA/FFRO Jorge Caspary, FDEP

DEPARTMENT OF THE AIR FORCE

AIR FORCE RESERVE COMMAND



September 8, 1999

MEMORANDUM FOR:

SEE DISTRIBUTION

FROM: 482d SPTG/CEV 29050 Coral Sea Blvd, Bldg. 232 Homestead ARS, FL 33039-1299

SUBJECT: Insertion of Institutional Control language into the Record of Decision for OU-5

Enclosed please find a copy of a paragraph to be inserted into the Record of Decision dated April, 1997 for OU-5. This paragraph incorporates language committing to institutional controls as included in the Land Use Control Implementation Plan (LUCIP) for this site.

If you have any questions, please do not hesitate to contact me at (305) 224-7163.

John B. Mitchell, Chief Environmental Engineering Flight

Attachment: ROD Insertion

Cc: HQ AFRC/CEVV, Mr. Philippe Montaigne AFBCE/DD Homestead, Mr. Tom Bartol Gannett Fleming, Hugh Vick

DISTRIBUTION: U.S. EPA, Doyle T. Brittain FDEP, Jorge R. Caspary DERM, James A. Carter



DEPARTMENT OF THE AIR FORCE

AIR FORCE RESERVE

MEMORANDUM FOR: SEE DISTRIBUTION

November 3, 1997

FROM: 482d SPTG/CEV 29050 Coral Sea Blvd. Building 232 Homestead ARS, Florida 33039-1299

SUBJECT: Record of Decision (ROD): Operable Unit #5

Attached for your approval and records is the final ROD signed by the Homestead Air Force Base Installation Commander, Air force Reserves for Operable Unit #5. Please contact Mr. Enrique Escalera at (305) 224-7324 or myself if you have any questions.

John B. Mitchell, Chief Environmental Engineering Flight

cc: w/o attach AFBCA/DD DERM, James A. Carter HQ AFRC, Toni B. Thorne

DISTRIBUTION: U.S. EPA, Doyle T. Brittain FDEP, Jorge R. Caspary

RECORD OF DECISION

Operable Unit 5 Site WP-1 Electroplating Waste Disposal Area Homestead Air Reserve Base Homestead, Florida FDEP Facility No. 138521996

June 1997

Montgomery Watson appreciates the opportunity to work for the U.S. Army Corps of Engineers, at the Homestead Air Reserve Base facility in Homestead, Florida. If you have any questions or comments concerning this report, please contact Mr. John B. Mitchell, Remedial Program Manager, Homestead Air Reserve Base.

Respectfully submitted,

MONTGOMERY WATSON

Jerry D. Gaccetta, P.G. Project Manager

9

Freddie Moreton, P.G. Project Geologist

FINAL

RECORD OF DECISION

FOR

OPERABLE UNIT 5 SITE WP-1, ELECTROPLATING WASTE DISPOSAL AREA

Homestead Air Reserve Base, Florida

June 1997

Prepared for:

U. S . Army Corps of Engineers Missouri River Division Omaha District Omaha, Nebraska

Prepared by:

Montgomery Watson 107 Mallard Street, Suite D St. Rose, Louisiana 70087

RECORD OF DECISION OPERABLE UNIT FIVE MOA INCORPORATION LANGUAGE

By separate Memorandum of Agreement (MOA) dated 15 March, 1999, with U.S. Environmental Protection Agency (U.S. EPA) and the Florida Department of Environmental Protection (FDEP), HARS, on behalf of the Department of the Air Force, agreed to implement base-wide, certain periodic site inspection, condition certification and agency notification procedures designed to ensure the maintenance by Installation personnel of any site-specific Land Use Controls (LUCs) deemed necessary for future protection of human health and the environment. A fundamental premise underlying execution of that agreement was that through the Air Force's substantial good-faith compliance with the procedures called for therein, reasonable assurances would be provided to U.S. EPA and FDEP as to the permanency of those remedies which included the use of specific LUCs.

Although the terms and conditions of the MOA are not specifically incorporated or made enforceable herein by reference, it is understood and agreed by the Air Force, U.S. EPA and FDEP that the contemplated permanence of the remedy reflected herein shall be dependent upon the Installation's substantial good-faith compliance with the specific LUC maintenance commitments reflected therein. Should such compliance not occur or should the MOA be terminated, it is understood that the protectiveness of the remedy concurred in may be reconsidered and that additional measures may need to be taken to adequately ensure necessary future protection of human health and the environment.

Land Use Controls Implemented:

Homestead ARS Installation Restoration Manager coordinates inspections and forwards discrepancies for correction.

Maintenance of signage to prevent unauthorized access.

Restrict construction. Workers must be notified that contamination exists and OSHA regulations apply if excavation activities are proposed on the site. Obtain concurrence from USEPA and FDEP prior to design. No residential usage allowed. Restrict groundwater access. No water supply wells allowed within the restricted area. Prior to all construction activities, a dig permit is required which also restricts groundwater access for this site.

Objective:

Prevent direct contact with contaminated media. Prevent trespasser and residential use.

Homestead Air Reserve Base, Florida Operable Unit 5, Site WP-1 Electroplating Waste Disposal Area

Declaration for the Record of Decision

DECLARATION STATEMENT

FOR THE

RECORD OF DECISION FOR

OPERABLE UNIT NO. 5 HOMESTEAD AIR RESERVE BASE SUPERFUND SITE

SITE NAME AND LOCATION

Homestead Air Reserve Base Homestead, Dade County, Florida Operable Unit No. 5, Site WP-1, Electroplating Waste Disposal Area (Former Site SP-1)

STATEMENT OF BASIS AND PURPOSE

This decision document presents the selected remedial action for the former Electroplating Waste Disposal Area, Operable Unit No. 5 (OU-5), Site WP-1, at the Homestead Air Reserve Base (ARB) (formerly Homestead Air Force Base), in Homestead, Florida. The selected remedial action is chosen in accordance with CERCLA, as amended by SARA, and, to the extent practicable, the National Oil and Hazardous Substances Pollution Contingency Plan (NCP). This decision is based on the administrative record for this site.

The State of Florida, the U.S. Environmental Protection Agency (USEPA), and the U.S. Air Force (USAF) concur with the selected remedy presented in the Record of Decision (ROD).

DESCRIPTION OF THE SELECTED REMEDY

No Further Investigation (NFI) with Land Use Controls.

DECLARATION STATEMENT

The selected remedy is protective of human health and the environment, complies with Federal and State requirements that are legally applicable or relevant and appropriate to the remedial action, and is cost-effective. This remedy utilizes permanent solutions and alternative treatment technologies, to the maximun extent practicable for the site. Because the previous removal/remedial action at this site left contaminants at industrial levels, institutional controls are required to prevent unacceptable exposures from hazardous substances that remain above health-based levels. This site will require a five-year review to assure there has been no migration of contaminants off site and that the institutional controls are effective in safeguarding human health and the environment. United States Air Force Homestead Air Reserve Base

By achard Justac _____

Date:_____

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Homestead Air Reserve Base, Florida Operable Unit 5, Site WP-1 Electroplating Waste Disposal Area

Decision Summary for the Record of Decision

DECISION SUMMARY

FOR THE

RECORD OF DECISION

1.0 SITE NAME, LOCATION, AND HISTORICAL DESCRIPTION

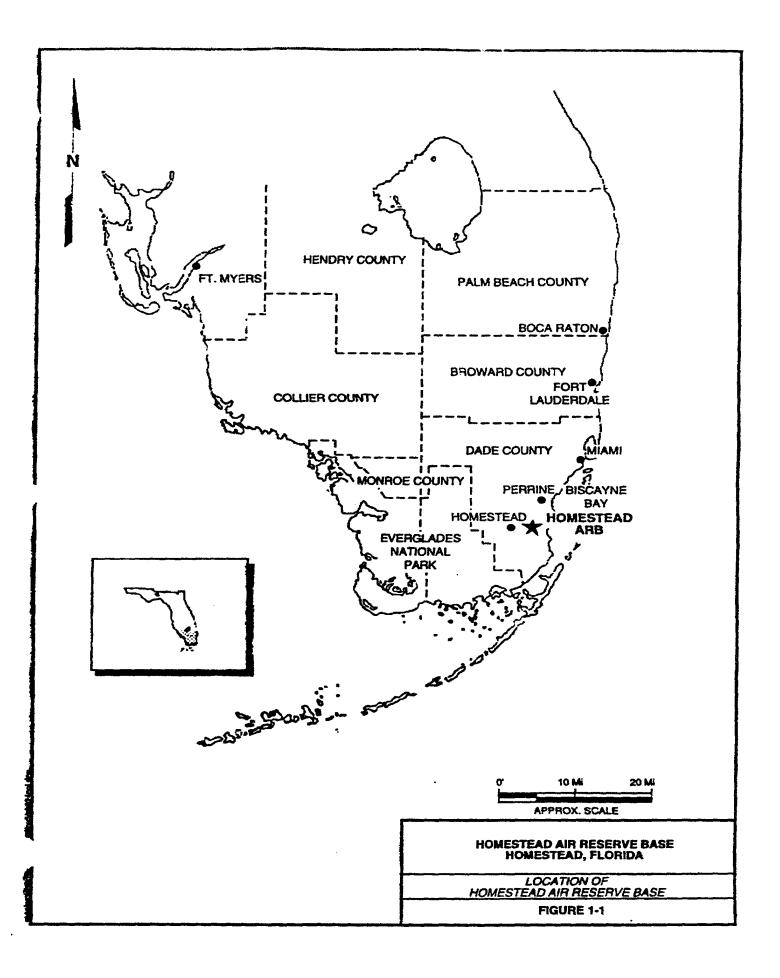
Homestead Air Reserve Base (ARB) is located approximately 25 miles southwest of Miami and 7 miles east of Homestead in Dade County, Florida (Figure 1-1). The main Installation covers approximately 2,916 acres while the surrounding areas are serni-rural. The majority of the Base is surrounded by agricultural land. The land surface at Homestead ARB is relatively flat, with elevations ranging from approximately 5 to 10 feet above mean sea level (msl). The Base is surrounded by a canal (Boundary Canal) that discharges to Outfall Canal and ultimately into Biscayne Bay approximately 2 miles east.

The Biscayne Aquifer underlies the Base and is the sole source aquifer for potable water in Dade County. Within 3 miles of Homestead ARB over 4,000 area residents obtain drinking water from the Biscayne Aquifer while 18,000 acres of farmland are irrigated from aquifer wells (USEPA, 1990). All recharge to the aquifer is through rainfall.

Homestead Army Air Field, a predecessor of Homestead Air Reserve Base, was activated in September 1942, when the Caribbean Wing Headquarters took over the air field previously used by Pan American Air Ferries, Inc. The airline had developed the site a few years earlier for pilot training. Prior to that time, the site was undeveloped. Initially operated as a staging facility, the field mission was changed in 1943 to training transport pilots and crews.

In September 1945, a severe hurricane caused extensive damage to the air field. The Base property was then turned over to Dade County and was managed by the Dade County Port Authority for the next eight years. During this period, the runways were used by crop dusters and the buildings housed a few small industrial and commercial operations.

In 1953, the federal government again acquired the airfield, together with some surrounding property, and rebuilt the Site as a Strategic Air Command (SAC) Base. The Base operated under SAC until July 1968 when it was changed to the Tactical Air Command (TAC and the

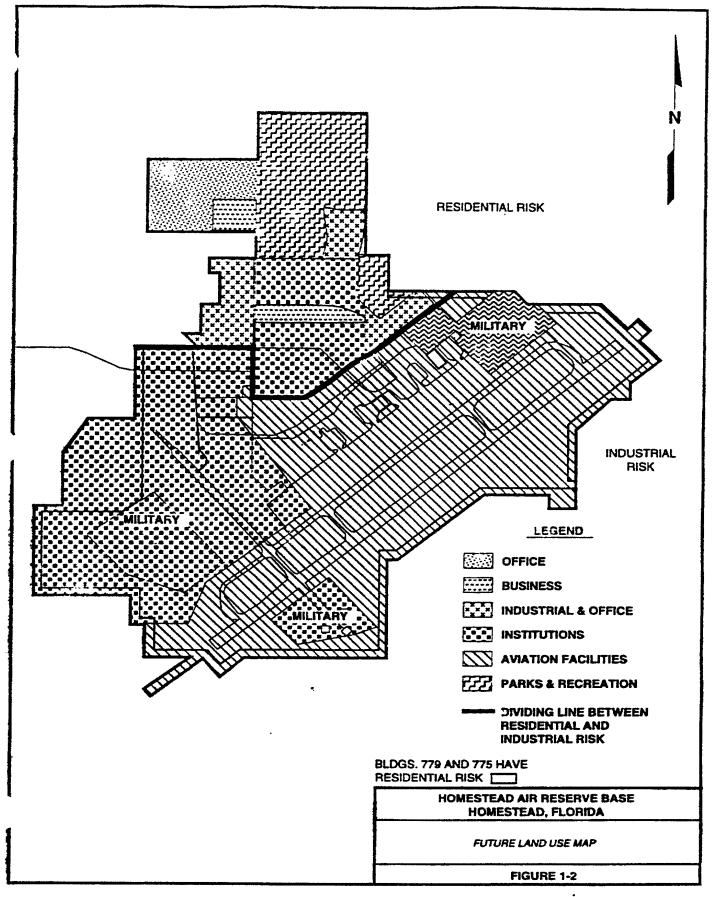


4531st Tactical Fighterwing became the new host. The Base was transferred to Headquarters Air Combat Command on June 1, 1992.

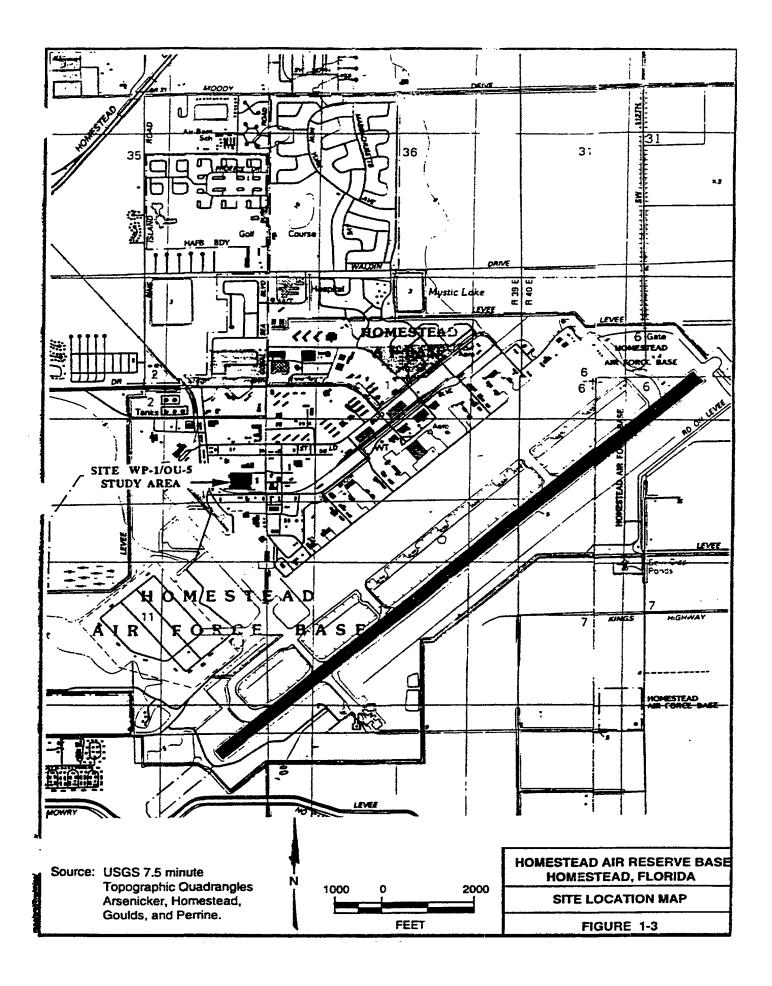
In August 1992, Hurricane Andrew struck south Florida causing extensive damage to the Base. The Base was placed on the 1993 Base Realignment and Closure (BRAC) list and slated for realignment with a reduced mission. Air Combat Command departed the Base on March 31, 1994 with Air Force Reservists activated at the Base on April 1, 1994. The 482nd Reserve Fighter Wing now occupies approximately 1/3 of the Base with the remaining 2/3 slated for use and oversight by Dade County. Figure 1-2 depicts the proposed future land use for the Base.

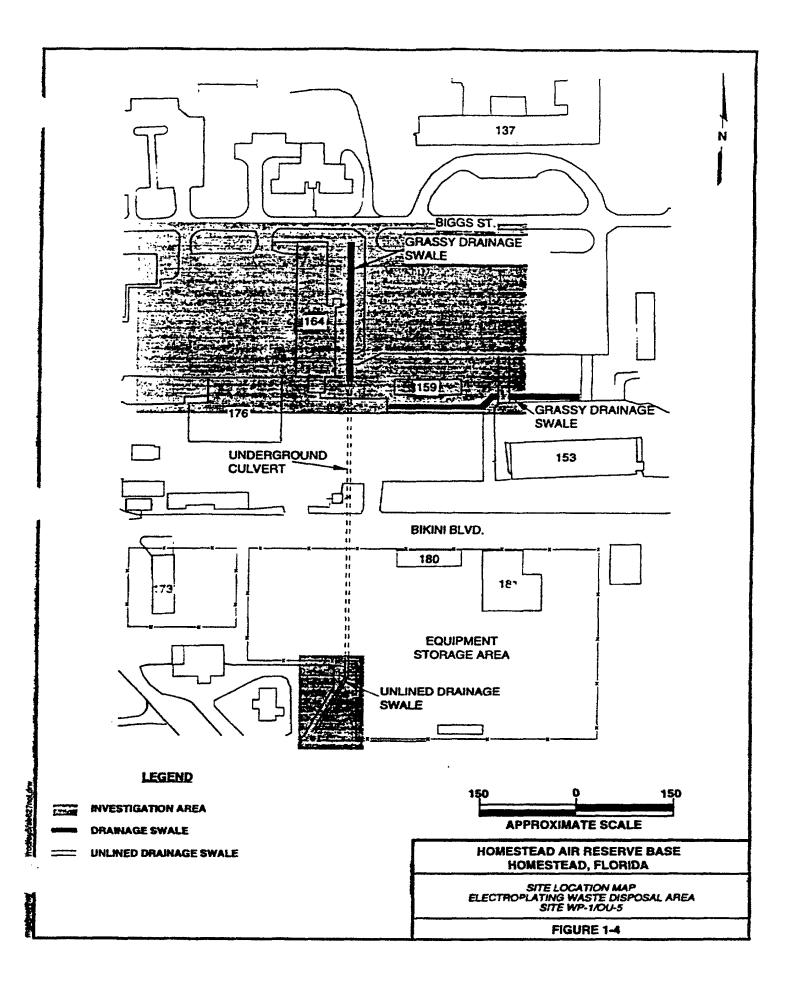
1.1 SITE DESCRIPTION

The Homestead ARB location map is depicted in Figure 1-3. The former Electroplating Waste Disposal Area (OU-5/Site WP-1) is located in the west-central portion of the Facility (Figure 1-4). The site is located in the portion of the Base remaining under control of the 482nd Reserve Fighter Wing. The site is in the cantonment area, east of Building 164 and consists of a grass lawn adjacent to Building 164, which is approximately 50 feet wide, and approximately the western third of the asphalt parking lot east of the grass lawn (Figure 1-4). Biggs Street is located north of the of the site and Buildings 163, 159, and 176 are located west, southeast, and southwest, respectively, of the site. A grassy drainage swale (low lying depressed stretch of land), with a north to south drainage flow direction, transects the grass lawn located adjacent to Building 164. Asphalt parking areas are located east of the grassy drainage swale and west of Building 164. The drainage swale east of Building 164, discharges into an underground culvert located south of Building 164. The underground culvert extends southward under Bikini Blvd. and an equipment storage area for approximately 500 feet and discharges into an unlined drainage swale just south of the equipment storage area. The equipment storage area also serves as a motor pool for large vehicles and contains an equipment/vehicle washrack. The unlined drainage swale is approximately five feet wide and three feet deep and has sparse terrestrial vegetation growing in the bottom of it and grass along the sides. The underground culvert collects runoff from OU-5/Site WP-1, Bikini Road, the equipment storage area, and other surrounding areas. South of the equipment storage area the drainage swale flows southwest into a drainage canal which travels for approximately one-half mile before entering the Boundary Canal. A second drainage swale, which flows from west to east into the culvert, is located southeast of Building 164 and south of Building 159. The two drainage swales located near Building 164



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and the unlined drainage Swale located south of the equipment storage area contain surface water only during heavy rain storms.

1.2 **REGIONAL LAND USE**

The area adjacent to Homestead ARB including OU-5/Site WP-1, to the west, east, and south within a half-mile radius is primarily composed of farmland and plant nurseries. Residential areas are located within a half-mile to the north and southwest of the Base. Woodlands are located approximately one-half-mile east of the facility and mangroves and marsh occur adjacent to Biscayne Bay. The Biscayne National Park is located 2 miles east of Homestead ARB; the Everglades National park is located 8 miles west-southwest of the Base; and the Atlantic Ocean is approximately 8 miles east of the Base.

1.3 SURFACE HYDROLOGY

Surface hydrology at Homestead ARB, including OU-5/Site WP-1 is controlled by five main factors: 1) relatively impermeable areas covered by runways, buildings and roads; 2) generally high infiltration rates through the relatively thin layer of soil cover; 3) flat topography; 4) generally high infiltration rates through the outcrop locations of the Miami Oölite Formation; and 5) relatively high precipitation rate compared to evapotranspiration rate. Infiltration is considered to be rapid through surfaces of oölite outcrop and areas with a thin soil layer. Infiltration rates are accelerated by fractures within the oölite, as well as naturally occurring solution channels. Precipitation percolates through the relatively thin vadose zone to locally recharge the unconfined aquifer.

Natural drainage is limited because the water table occurs at or near land surface. The construction of numerous drainage canals on Homestead ARB has improved surface water drainage and lowered the water table in some areas. Rainfall runoff from within Homestead ARB boundaries is drained via diversion canals to the Boundary Canal.

A drainage divide occurs within the Homestead ARB facility property, running from the northern end of the facility, toward the center. Water in the Boundary Canal flows generally south and east along the western boundary of the property, and south along the eastern boundary, converging at a storm-water reservoir located at the southeastern corner of the Base. Flow out of the stormwater reservoir flows into Outfall Canal, which, in turn, flows east into Biscayne Bay, approximately 2 miles east of the Base. Water movement is typically not visible in the canals in dry weather due to the lowered water table and the very low surface gradient (0.3 feet per mile) that exists at the Base.

1.3.1 Regional Hydrogeologic Setting

The regional hydrogeology in the southeast Florida area consists of two distinct aquifers: the surficial aquifer system, which consists of the Biscayne Aquifer and the Grey Limestone Aquifer, and the lower aquifer, the Floridan Aquifer.

Biscayne Aquifer. The Biscayne Aquifer at Homestead ARB consists of the Miami Oölite, Fort Thompson Formation, and the uppermost part of the Tamiami Formation. In general, the most permeable parts of the aquifer lie within the Miami Oölite and the Fort Thompson Formation.

The Biscayne Aquifer underlies all of Dade, Broward, and southeastern Palm Beach Counties. The Biscayne Aquifer is the sole source of potable water in Dade County and is a federally-designated sole-source aquifer pursuant to Section 1425 of the Safe Drinking Water Act (SDWA). The Biscayne Aquifer supplies drinking water to approximately 2.5 million people within local communities. All recharge to the aquifer is derived from local rainfall, part of which is lost to evaporation, transpiration, and runoff.

The Biscayne Aquifer has reported transmissivities ranging from approximately 4 to 8 million gallons per day per foot (mgd/ft) (Allman et al., 1979).

Water-table contours indicate that under natural conditions, groundwater flows southeasterly toward Biscayne Bay. The hydraulic gradient is approximately 0.3 ft/mile. The water table at Homestead ARB generally is encountered within 5 to 6 feet of land surface, but may occur at or near land surface during the wet season (May to October). Fluctuations of groundwater levels and local variations in the direction of groundwater flow are due to several factors: (1) differences in infiltration potential, (2) runoff from paved areas, (3) water-level drawdown near pumping wells, (4) significant but localized differences in lithology (e.g., silt-filled cavities) and (5) drainage effects of canals and water-level control structures.

Floridan Aquifer. Underlying the low-permeability sediments of the Tamiami Formation and Hawthorn Group are the formations which constitute the Floridan Aquifer.

The Floridan Aquifer is made up of limestones and dolomites. It is under artesian pressure and water levels in deep wells may rise 30 to 40 ft above ground surface. Groundwater within these Miocene and Eocene age formations tends to contain dissolved constituents at levels significantly above those recommended for drinking water. In view of the poor water quality and the depth of water yielding zones (800 to 900 feet bgs), the Floridan Aquifer is of limited usefulness as a source of potable water supply in the study area.

1.4 SITE GEOLOGY AND HYDROGEOLOGY

The stratigraphy of the shallow aquifer system as determined from soil borings performed during site investigations by Geraghty & Miller (G&M) and Montgomery Watson consists of a surficial weathered Miami Oölite ranging in depth from 2 to 6 feet below ground surface (bgs). The weathered limestone consists of a white to brown semi-consolidated oölitic limestone. This strata is underlain by consolidated to semi-consolidated oölitic and coral limestone interbedded with coarse to fine sand and clayey sand layers and lenses down to the total depth of borings (approximately 40 feet bgs).

The Biscayne Aquifer is one of the most transmissive aquifers in the world. It underlies Homestead ARB. A thin vadose zone, nominally less than 5 feet deep, overlays the groundwater table at the site. As previously stated, the aquifer structure is a calcium carbonate matrix. This lithology is known to have natural concentrations of target analyte list (TAL) metals. In descending order by concentration, calcium, aluminum, iron magnesium, sodium, and potassium can be considered the primary metals of carbonate rock. The other TAL metals occur in trace concentrations, less than 50 milligrams per kilogram (mg/kg). The range and the standard deviations are not provided at this time. It should be expected that, as precipitation infiltrates and recharge takes place, leaching of metal ions from the weathered vadose zone and shallow unsaturated zone occurs. Regional data collected suggest that concentrations of trace metals can be expected to be the greatest in the shallow portion of the aquifer because of the proximity to the source (i.e., the weathering vadose structure) and the decreasing retention time with decreasing depth of the saturated zone. These observations support a hydrogeologic model in which the shallow portion of the aquifer has a greater horizontal transmissivity than the vertical component during recharge events. However, it is not possible, from the available data at the site, to quantitatively differentiate horizontal and vertical components of the aquifer's hydrologic conductivity. The possible presence of vertical solution zones is well documented in literature. The site-specific effects have not been fully investigated. Nevertheless, the available data does not lead to the immediate conclusion that this is a necessary task. The conceptual model that the

shallow groundwater is discharging to the ditches and canals provides sufficient detail for the purpose of discussing Site OU-5/WP- 1.

2.0 SITE HISTORY AND ENFORCEMENT ACTIVITIES

2.1 OPERABLE UNIT NO. 6 HISTORY

2.1.1 Past Site Usage

OU-5/Site WP- 1 was formerly used as a disposal area for spent plating baths and rinses from a plating shop located in Building 164. During the period between 1946 and 1953, when Homestead ARB was inactive and ownership of the property was transferred to Dade County, a small electroplating operation was located in Building 164. Spent plating solutions containing chromium, nickel, copper, and sulfuric and hydrochloric acid were routinely disposed of by discharging them on the ground in an area just east of Building 164 (Engineering-Science, 1993). Wastes were generated at a rate of approximately 250 gallons per month, and the electroplating operation continued for about two years (Engineering-Science, 1983). According to 1958, 1962, and 1973 aerial photographs, the asphalt parking lot located east of Building 164 did not exist when the electroplating company was operating and the area was covered with trees and grass. The parking lot was constructed between 1962 and 1973. No visible evidence of waste residue is observed on the ground surfaces that are presently exposed (not covered by parking lot). The high amount of rainfall typical of the area is suspected to have dissipated the waste residues.

2.1.2 Future Site Usage

OU-5/Site WP-1 is within the cantonment area of the 482nd Air Force Reserve unit. The cantonment area has restricted access and is fenced off from other areas of the Base.

2.2. ENFORCEMENT HISTORY

2.2.1 CERCLA Regulatory History

The Comprehensive Environmental Response. Compensation and Liability Act of 1980 (CERCLA) established a national program for responding to releases of hazardous substances into the environment. In anticipation of CERCLA, the Department of Defense (DOD) developed the Installation Restoration Program (IRP) for response actions for

potential releases of toxic or hazardous substances at DOD facilities. Like the U.S. Environmental Protection Agency's (USEPA's) Superfund Program, the IRP follows the procedures of the National Oil and Hazardous Substances Pollution Contingency Plan (NCP). Homestead ARB was already engaged in the IRP Program when it was placed on the National Priorities List (NPL) on August 30, 1990. Cleanup of DOD facilities is paid for by the Defense Environmental Restoration Account (DERA), which is DOD's version of Superfund.

The Superfund Amendment and Reauthorization Act (SARA), enacted in 1986, requires federal facilities to follow NCP guidelines. The NCP was amended in 1990 (see 40 CFR 300 et seq.) to implement CERCLA under SARA. In addition, SARA requires greater USEPA involvement and oversight of Federal Facility Cleanups. On March 1, 1991, a Federal Facility Agreement (FFA) was signed by Homestead ARB (formerly Homestead AFB), the USEPA, and the Florida Department of Environmental Protection (FDEP). The FFA guides the remedial design/remedial action (RD/RA) process.

The purpose of the FFA was to establish a procedural framework and schedule for developing, implementing, and monitoring appropriate response actions at Homestead ARB in accordance with existing regulations. The FFA requires the submittal of several primary and secondary documents for each of the operable units at Homestead ARB. This ROD concludes all of the remedial investigation/feasibility study (RI/FS) requirements for OU5/Site WP- 1.

As part of the RI/FS process, Homestead ARB has been actively involved in the Installation Restoration Program (IRP). From 1983 to 1992, 27 Potential Sources of Contamination (PSCs) were identified at Homestead ARB. Ten sites have been investigated in the PA/SI stage of CERCLA, with four sites warranting no further investigation and six sites requiring further investigation. One of the PSCs sites has been closed under the Resource Conservation and Recovery Act (RCRA) guidelines and seven sites were investigated under the FDEP petroleum contaminated sites criteria (Florida Administrative Code (FAC) 62-770). Additionally, a RCRA Facility Investigation (RFI) has been conducted to evaluate numerous solid waste management units (SWMUs) identified during the RCRA Facility Assessment (RFA). A cleanup effort was initiated after Hurricane Andrew to prepare the base for realignment. This included the removal of fuel storage tanks and oil1water separators. Additional PSC have been identified subsequent to 1992 as a result of investigations and/or remediation of the base. The following PSC sites are currently in various stages of reporting under the CERCLA RI/FS guidelines.

	<u>Operable</u>
PSC Name	<u>Unit No.</u>
Fire Protection Training Area 2	1
Residual Pesticide Disposal Area	2
Oil Leakage Behind the Motor Pool	4
Electroplating Waste Disposal Area	5
Aircraft Washrack Area	6
Entomology Storage Area	7
Fire Protection Training Area 3	8
Boundary Canal	9
Landfill LF-12	10
Sewage Treatment Plant/Incinerator Ash Disposal Area	11
Entomology Shop	12
Landfill SS-22	13
Drum Storage Area	14
Hazardous Storage Bldg.	15
Missile Site	16
Hangar 793	17
Construction Debris Landfill	18
Bldg. 208	19
Bldg. 618 Parking Lot	20
# 32,Bldg. 619 Parking Lot	21
Bldg. 761/764	22
Bldg. 814	25
Bldg. 745	26
Bldg. 268 & 268 A	27
Bldg. 750	28
Bldg. 760	29

Operable Unit No. 3 PCB Spill, C.E. Storage Compound has been closed out with a No Further Action Record of Decision (ROD) in June 1994. Operable Units 1, 2, 4, and 6 have been completed through the ROD stage requiring various levels of remedial action/remedial design. OU-8 has been closed out under CERCLA with a No Further Investigation Decision Document and is being transfered to the FAC 62-770 program. Two Solid Waste Management Units, OU-23 and OU-24, have been closed out while three areas of concern (AOC-1, AOC-3, and AOC-5) are in the preliminary assessment phase of investigation.

Figure 2-1 depicts the above-listed CERCLA sites, as well as the FAC 62-770 fuel contaminated sites currently under investigation.

The Base Realignment and Closure (BRAC) Cleanup Plan currently incorporates both the IRP and associated environmental compliance programs to support full restoration of the base.

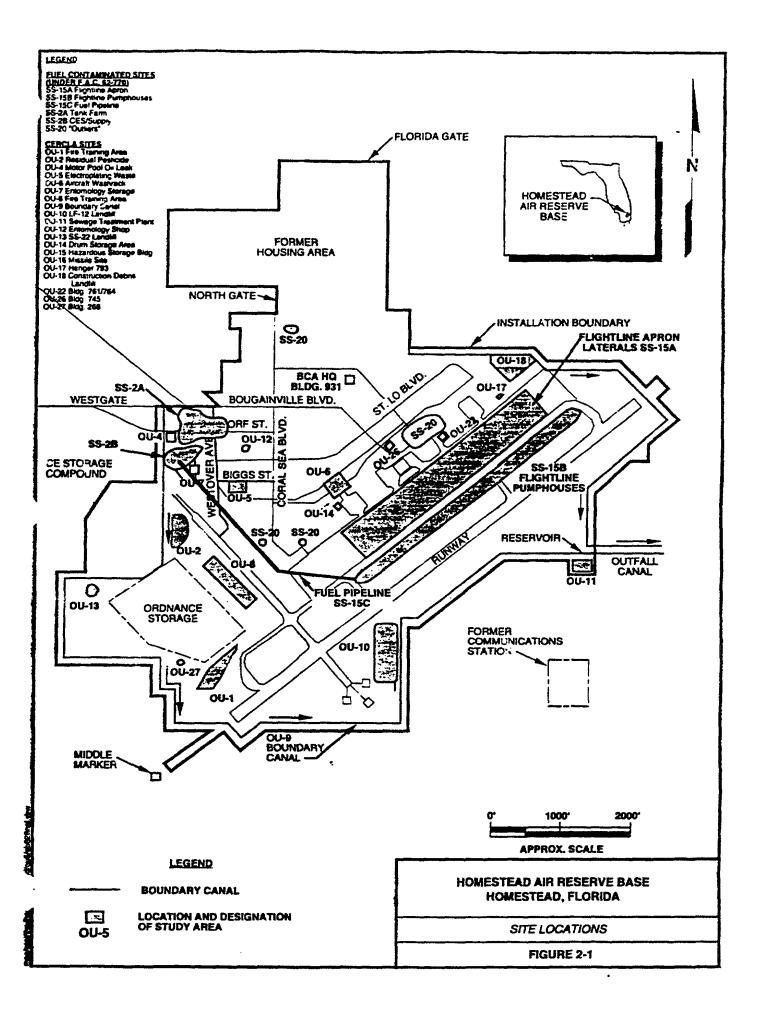
2.3 INVESTIGATION HISTORY

2.3.1 IRP Phase I - Record Search

An IRP Phase I - Records Search was performed by Engineering Science, and is summarized in their report, dated August 1983 (Engineering-Science, 1983). During the Phase I study, sites with the potential for environmental contamination resulting from past waste disposal practices were identified. Thirteen sites of potential concern were identified by reviewing available installation records, interviewing past and present Base employees, inventorying wastes generated and handling practices for these wastes, conducting field inspections, and reviewing geologic and hydrogeologic data. In general, Phase I studies are used to determine if a site requires further investigation.

The thirteen identified sites were ranked using the Hazard Assessment Rating Methodology (HARM) developed by JRB Associates of McLean, Virginia, for the U.S. Environmental Protection Agency (USEPA). HARM was later modified for application to the Air Force IRP. The following factors are considered in HARM: (1) the possible receptors of the contaminants; (2) the characteristics of the waste; (3) potential pathways for contaminant migration; and (4) waste management practices. HARM scores for the sites ranked at Homestead ARB ranged from a high of 72 to a low of 7 out of 100. Eight of the thirteen sites were determined to have a moderate-to-high contamination potential, one of which was the Electroplating Waste Disposal Area. These sites were recommended for additional monitoring. The remaining five sites were determined to have a low potential for environmental contamination.

According to the IRP Phase I Report, Site OU-5/WP-1 received a moderate to high HARM score of 72 due to the nature of the waste (persistent metals) that were disposed of at the site (Engineering-Science, 1983). In addition, Site OU-5/WP-1 scored high as a potential migration pathway because of the extremely permeable nature of the underlying rock, shallow groundwater, and the proximity of the drainage swale to the site. This score,



however, did not reflect the site conditions which have probably caused dissipation of the wastes; including heavy rain infiltration and the amount of time (30 years) since the disposal activities ceased. The Phase I report recommended sampling the existing Base supply well, the one currently or most recently in service, from Well Field No. 1 and analyzing for pH, total dissolved solids (TDS), chromium, nickel, copper, and sulfate.

2.3.2 IRP Phase II - Confirmation/Quantification

An IRP Phase II study was performed by Science Applications International Corporation in 1984 and a report was completed in March 1986 (SAIC, 1986). The objectives of a Phase II study are to confirm the presence or absence of contamination, to quantify the extent and degree of contamination, and to determine if remedial actions are necessary. During the Phase II study, additional investigations were performed at the eight sites recommended for monitoring in the Phase I report, as well as two of the other thirteen originally-identified sites. The Electroplating Waste Disposal Area was included in this investigation.

The Phase II - Confirmation/Quantification investigation included installation of three monitoring wells (I-01 through I-03) and four soil borings (SL-1 through SL-4), collection of two sediment samples (SD-1 and SD-2), and groundwater sampling. The monitoring wells were sited for the purpose of confirming and quantifying suspected contaminants. The locations of these monitoring wells and the sampling points are shown in Figure 2-2. The soil, sediment, and groundwater samples were analyzed for total metals (cadmium, chromium, copper, lead, nickel, zinc) and cyanide. The groundwater samples were also analyzed for hexavalent chromium.

The groundwater analyses indicated the presence of metals; however, the concentrations detected were below applicable Federal and Florida drinking water standards. The detected levels of metals in the soil and sediment samples are comparable to common background levels for those metal constituents. Elevated levels of cyanide, however, were detected in the soil and sediment samples. Table 2-1 presents the analytical results for the groundwater samples and Table 2-2 presents the analytical results for soil and sediment samples collected during the 1984 IRP investigations.

The Phase II report contained the following alternatives for additional investigation at this site: (1) resample existing wells for inorganics; (2) install an additional monitoring well located southeast of Building 159; (3) collect surface-water (and sediment) samples from the drainage canal and runoff from the site and other areas to assess the contaminant contribution

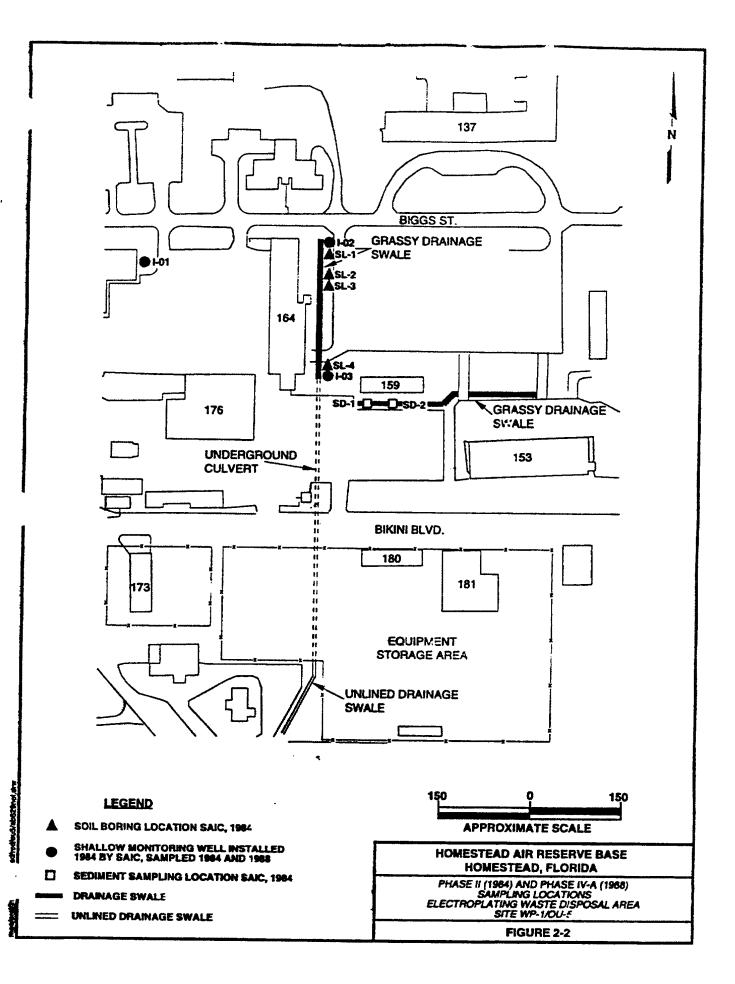


TABLE 2-1

ANALYTICAL RESULTS FOR PHASE II GROUNDWATER SAMPLES SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA SAIC, 1984 INVESTIGATION

Homestead Air Reserve Base, Florida

		REPLICATE				
ANALYTE	LOCATION	I-01	I-02	I-02	I-03	
Field Parameters						
pH (Std. Units)		6.8	6.8	6.8	NA	
Temp (°C)		29	24	24	29	
Specific Conductance		420	420	420	430	
(umhos/cm)						
Metals: ug/L:						
Cadmium		0.2	0.2	<0.2	0.4	
Total Chromium		19.7	<0.5	<0.5	<0.5	
Hexavalent Chromium		0.3	< 0.1	<0.1	1.7	
Copper		4.2	5.3	4.8	7.0	
Lead		7.8	6.8	9.0	5.7	
Nickel		10.6	10.8	9.2	16.9	
Zinc		15.1	16.3	15.7	16.3	
Cyanide ug/L		<5.0	<5.0	<5.0	<5.0	

Source: Geraghty & Miller, Inc., 1992

NA - Not Available

ANALYTICAL RESULTS FOR PHASE II SOIL AND SEDIMENT SAMPLES COLLECTED SITE WP-1/OU.5, ELECTROPLATING WASTE DISPOSAL AREA

SAIC, 1984 INVESTIGATION

			SOI	L			SEDIMENT	ſS
					REPLICATE		F	REPLICATE
ANALYTE (mg/Kg)	SL-1	SL-2	SL-3	SL-4	SL-4	SD-1	SD-2	SD-21
								_
Cadmium	0	0	0	0	0	0	0.01	0.01
Total Chromium	0.07	0.02	0.03	0.04	0.04	0.02	0.09	0.07
Copper	0.08	0.02	0.03	0.11	0.12	0.01	0.04	0.05
Lead	0.17	0.09	0.11	0.21	0.20	0.11	1.18	1.17
Nickel	0.02	0.01	0.04	0.05	0.04	0	0.01	0.01
Zinc	0.72	0.12	0.15	0.28	0.24	0.18	0.92	0.80
Cyanide	1.3	0.60	0.70	0.80	0.80	0.60	3.00	3.90

Homestead Air Reserve Base, Florida

Source: Geraghty & Miller, Inc., 1992

Shading indicates sediments have been removed during the 1995 Interim Removal Action. mg/kg - milligrams per kilogram

of the site area and whether significant levels of inorganics are migrating from the site via surface water, (4) collect additional soil samples to delineate soil contamination; and (5) a combination of alternatives one and three. The recommendations of the Phase II report included: sampling the canal water and sediments southeast of Building 159 at two locations during the wet season to determine the role of surface water as a pathway; resampling existing monitoring wells for comparison with Stage I results to determine the statistical significance of the elevated inorganics indicated by Stage I analysis, and installation and sampling of two additional monitoring wells. A complete discussion of the methods and results of this study are presented in the Phase II - *Confirmation/Quantification Report* (SAIC, 1986).

2.3.3 IRP Phase III - Technology Base Development

The IRP Phase III is a research phase and involves technology development for an assessment of environmental impacts. There have been no Phase III tasks conducted at the Base to date.

2.3.4 IRP Phase IV - Additional Investigations

The IRP Phase IV investigations congist of two areas of work activity. Phase IV-A involves additional site investigations necessary to meet the Phase II objectives, a review of all management methods and technologies that could possibly remedy site problems, and preparation of a baseline risk assessment to address the potential hazards to human health and the environment associated with the constituents detected at the site. Detailed alternatives are developed and evaluated and a preferred alternative is selected. The preferred alternative is then described in sufficient detail to serve as a baseline document for initiation of Phase IV-B.

An IRP Phase IV-A investigation was performed at Site OU-5/WP-1 by Geraghty & Miller in 1988. The results of this investigation are included in the report entitled *Draft Remedial Investigation Report for the Building 207 Underground Storage Tank Area, Residual Pesticide DisposalArea, and the Electroplating Waste DisposalArea, HomesteadAir Force Base, Florida,* October 1988.

The Phase IV-A - Remedial Actions investigations included additional groundwater analysis, water-level measurements, and a topographic survey. Groundwater samples from the existing wells were collected and analyzed for total metals (arsenic, barium, cadmium,

chromium, copper, lead, mercury, nickel, selenium, silver, sodium) and cyanide. Arsenic and sodium were the only constituents detected at concentrations exceeding the quantitation limit (Table 2-3). All detected concentrations were below applicable Federal and Florida standards. The locations of the permanent monitoring wells, as well as significant topographic features of the site, have been indicated on Figure 2-2.

No significant risks to public health or the environment were identified in the baseline risk assessment; therefore, the *Draft RI Report* (Geraghty & Miller, 1988) recommended no further action at this site. A "no further action" recommendation specifies that additional investigations and/or remedial action are not warranted at the site.

2.3.5 1991 Remedial Investigation

In 1991, G&M conducted a remedial investigation at OU-5/Site WP-1. This investigation included the installation of one additional monitoring well (SP1-MW-0001), sampling of the three existing monitoring wells plus the new well, collection of four shallow soil/weathered rock samples, two sediment, and two surface water samples. The 1991 sampling locations are shown on Figure 2-3. All samples were analyzed for the USEPA Target Compound List (TCL) VOCs and BNAs, and the Target Analyte List (TAL) metals and cyanide using USEPA SW-846 the methods. The results of these analyses are presented in Sections 2.6.2, 2.6.3, 2.6.4, and 2.6.5. Complete results of the 1991 RI are presented in G&Ms report titled *Remedial Investigation Report for Site WP-1 Electroplating Waste Disposal Area*, June 1992. Geraghty & Miller's conclusion presented in the RI Report was that no additional study was recommended.

2.3.6 1993 Remedial Investigation Addendum

In 1993, Montgomery Watson Americas, Inc. performed supplemental RI activities to evaluate the soil and groundwater quality with respect to the USEPA target compound list/target analyte list (TCL/TAL), to fill data gaps from the previous field investigations, and to evaluate any impacts due to Hurripane Andrew. The 1993 investigation included the drilling of five soil borings, groundwater sampling of seven shallow and one deep monitoring well, and the collection of three sediment and surface water samples. Sampling locations are illustrated on Figure 2-4. All samples were analyzed for TCL Organochlorine Pesticides/PCBs and cyanide. All sediments, one soil, and one groundwater sample were also analyzed for the presence of TCL VOCs, BNAs, and TAL metals. Dissolved (filtered)

ANALYTICAL RESULTS OF PHASE IV-A GROUNDWATER SAMPLES SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1988 INVESTIGATION Homestead Air Reserve Base, Florida

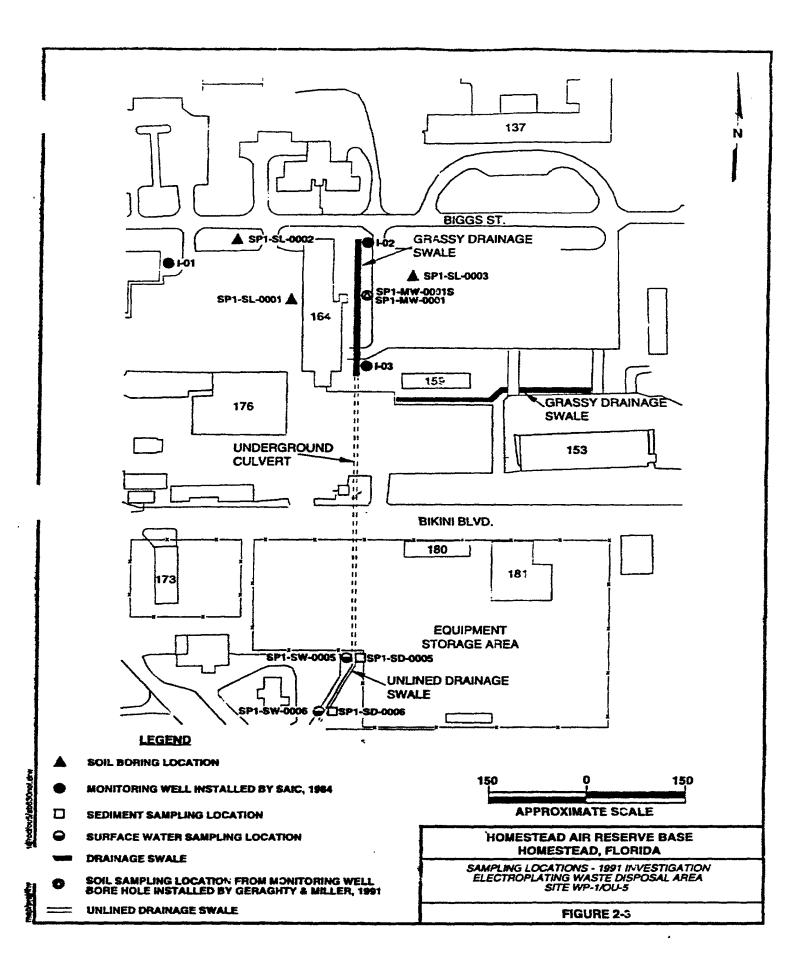
	Homesteau Air Ke	serve base, Florida	
		LOCATION	
Analytes	I-01	I-02	I-03
Metals (ug/L)			
Total arsenic	24	13	[8.1]
Total barium	[7.6]	[5.7]	[5.8]
Total cadmium	<0.12	[0.48]	<0.12
Total chromium	[8.9]	[9.4]	[9.1]
Total copper	<7.8	<7.8	[8.3]
Total lead	[1.4]	[2.6]	[1.6]
Total mercury	<0.13	[0.15]	[0.16]
Total nickel	[14]	<11	<11
Total selenium	[1.0]	[0.54]	[0.99]
Total sodium	33,500	7710	31,400
Total cyanide	[7.4]	<2.6	24

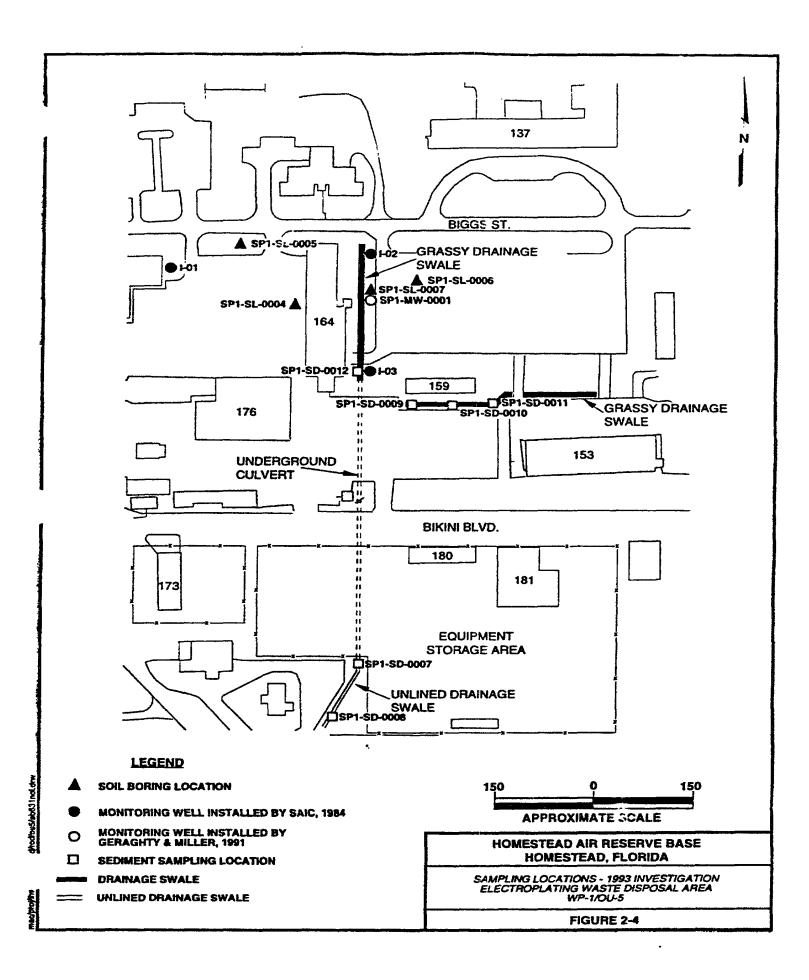
Source: Geraghty & Miller, Inc., 1992

ug/L - micrograms per liter

< Less than stated quantitation limit

[] Value is between level of quantition and instrument detection limit.





TAL metals were also analyzed for in groundwater samples. The results of these analyses are presented in Sections 2.6.2, 2.6.3, 2.6.4, and 2.6.5.

2.3.7 1994 and 1995 Investigations

Confirmatory samples were collected during the summer of 1994 from 28 shallow borings at the locations shown in Figure 2-5. This work was completed under the direction of the USACE-Mobile District. For the purposes of this evaluation, samples collected from the two borings advanced west of Building 164 and the five borings in the grassy swale east of Building 164 were considered soil samples due to their location in a asphalt paved parking area and the absence of a well defined drainage system in the swale. The remaining samples were considered sediments. The two samples collected from the borings located west of Building 164 were analyzed for semi-volatile organic compounds (USEPA SW-846 Method 8270) and TAL metals (USEPA SW-846 Method 6010, 7060, 7421, 7740, and 7471). The remaining 26 soil/sediment samples were analyzed for semi-volatile organics, TAL metals, and pesticides (USEPA Method 8080). All analyses were performed in accordance with USEPA SW-846 protocol. The results of these analyses are presented in Sections 2.6.2 and 2.6.4. Further information regarding the collection of the 28 confirmatory soil samples collected during the 1994 sampling event is provided in the IT report entitled *Confirmation Sampling Results: Electroplating Waste Disposal Area* (IT Corporation, August 29, 1994).

In 1995, IT conducted an Interim Action (IA) at OU-5/Site WP-1. This work was completed under the direction of the USACE-Mobile District. The remedial activities included delineation and profiling of contaminated soils/sediments, excavation and disposal of contaminated soils/sediments, and analysis of confirmation samples collected from within the excavation limits. Further details on the activities conducted during the 1995 investigation can be found in the IT report entitled *Interim Action Report: Electroplating Waste Disposal Area* (OU-5) (IT Corporation, November 30, 1995).

After excavation, three confirmation samples were collected from the base of the excavated areas. These confirmation samples were analyzed for semi volatile organics, TAL metals, and pesticides using USEPA SW-846 methods. Confirmation sample locations are depicted on Figure 2-6. The results of these analyses are presented in Section 2.6.2.

During the investigations and excavation activities, 14 analytical samples were collected from soil/sediments that were later excavated during the Interim Action. Of those samples, 2 were collected during the 1984 Science Applications International Corporation (SAIC)

investigation, 4 were samples collected during the 1993 Montgomery Watson Investigation, and 14 were samples collected during the 1994 IT Corporation projects. The results from these analysis are no longer representative of current site conditions. A complete list of the soil/sediment samples which were subsequently excavated is presented in Table 2-4.

2.3.8 1996 Confirmation Groundwater Sampling

On January 24 and 25, 1996, OHM, under contract with the Air Force Center for Environnental Excellence (AFCEE), conducted as confirmation groundwater sampling event at OU-5/Site WP-1. During this event groundwater samples were obtained from each of the sites 4 groundwater monitoring wells (SPI-MW-0001, I-01, I02, and I-03) (Figure 2-7). Groundwater samples were collected with USEPA and State of Florida approved methods. The 4 groundwater samples were analyzed by Analytical Technologies, Inc., in Pensacola Florida, for the target compound list (TCL) volatile organic compounds (VOC's), TCL base neutral and acid extractable (BNA's), TCL organochlorine pesticides/PCB's, and target analyte list (TAL) metals and cyanide. The results of these analytes are presented in Section 2.6.3. All samples were analyzed in accordance with USEPA Contract Laboratory Protocol (CLP) requirements.

The results of the site characterization activities conducted during the 1991 and 1993 investigations, as well as the results from the 1994 and 1995 IT Corporation Interim Action and the 1996 OHM confirmation sampling event are presented in Sections 2.6.2, 2.6.3, 2.6.4, and 2.6.5 of this ROD.

2.4 COMMUNITY RELATIONS HISTORY

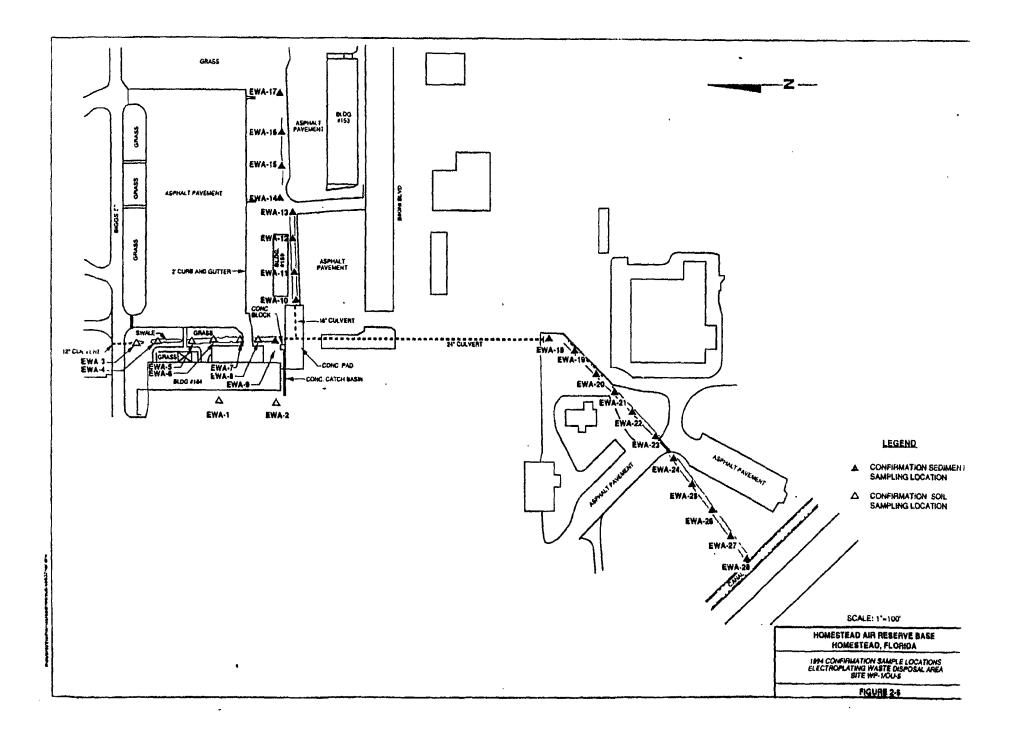
The Remedial Investigation/Baseline Risk Assessment Report and the Proposed Plan (PP) for Homestead ARB, OU-5/Site WP-1 were released to the public in October and December 1996, respectively. These documents were made available to the public in both the administrative record and an informal repository maintained at the Air Force Base Conversion Agency OL-Y office. A public comment period will be held from March 16, 1997 to April 14, 1997 as part of the comunity relations plan for OU-5/Site WP-1. Additionally, a public meeting was be held on Thursday, March 13, 1997 at 7:00 pm at the South Dade Senior High School. A Public Notice was published in the Miami Herald and the South Dade News Leader on February 21, 1997. At this meeting, the USAF, in coordination with USEPA Region 4, FDEP and Dade County Environmental Resources Management (DERM), will be prepared to discuss the Remedial Investigation, the

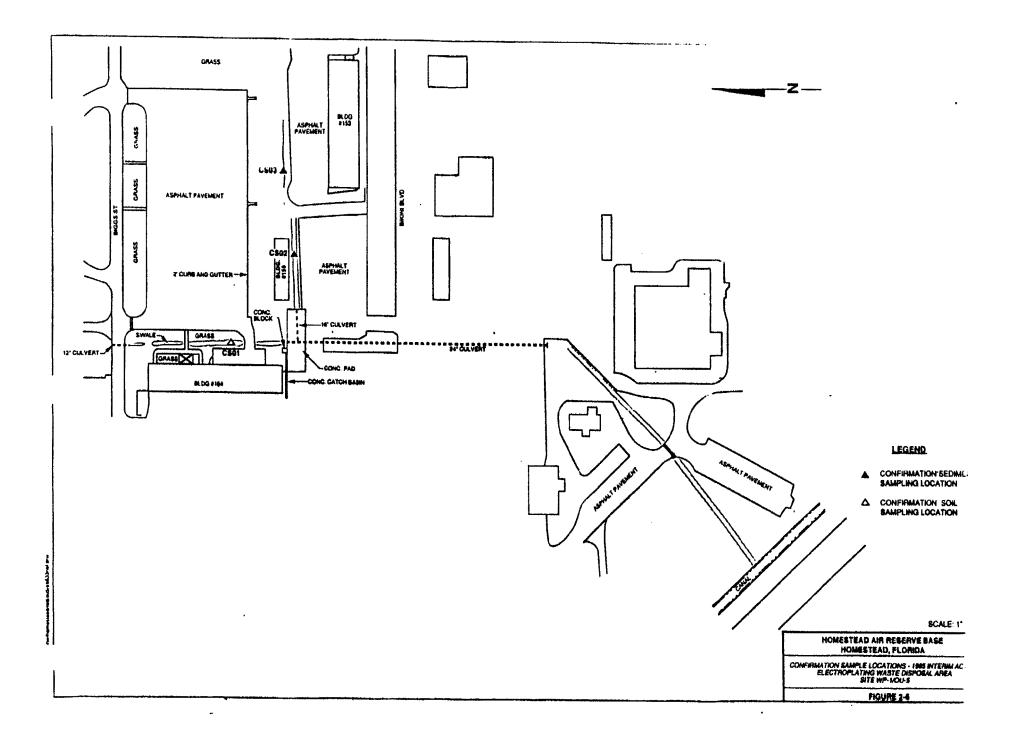
SUMMARY OF EXCAVATED SAMPLES Site WP-1/OU-5, Electroplating Waste Disposal Area IT Corporation, 1995 Interim Removal Action (IRA) Homestead Air Reserve Base, Florida

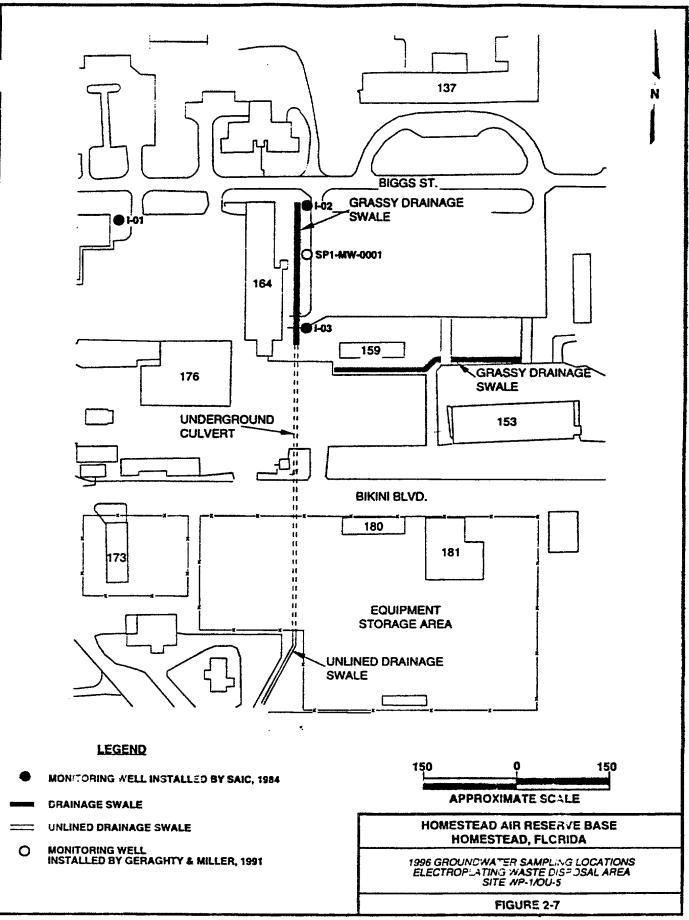
Sample Identifier	Sampling Depth Interval	IRA Excavation Depth	Investigation	Comments
NORTH AREA (a)				
SD-1 (sediment)	(a)	1'	1984 (SAIC)	Excavated
SD-2 (sediment)	(a)	1'	1984 (SAIC)	Excavated
SP1-SD-0009 (sediment)	0'-1'	1'	1993 (Montgomery Watson)	Excavated
SP1-SD-0010 (sediment)	0'-1'	1'	1993 (Montgomery Watson)	Excavated
SP1-SD-0011 (sediment)	0'-1'	1'	1993 (Montgomery Watson)	Excavated
SP1-SD-0012 (sediment)	0'-1'	1'	1993 (Montgomery Watson)	Excavated
EWA-6 (soil)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-7 (soil)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-8 (sediment)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-10 (sediment) EWA-11 (sediment)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-12 (sediment)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-12 (sediment)	0'-0.5'	1'	1994 (IT Corporation)	Excavated
EWA-15 (sediment)			-	

NOTES:

(a) Sample interval unknown







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Baseline Risk Assessment, and the Preferred Alternative as described in the Proposed Plan. A court reporter will prepare a transcript of the meeting. A copy of the transcript and all written comments received during the comment period will be placed in the Administrative Record. A response to the comments received during this period will be included in the Responsiveness Summary section of a later draft of this ROD. This decision document presents the selected remedial action for OU-5/Site WP-1 at Homestead ARB, chosen in accordance with CERCLA, as amended by SARA and, to the extent practicable, the National Contingency Plan. The decision for this site is based on the administrative record.

2.5 SCOPE AND ROLE OF RESPONSE ACTION

Homestead ARB, Florida with concurrence from the FDEP and USEPA, has elected to define Operable Unit 5 as the former Electroplating Waste Disposal Area and associated potential soil and groundwater contamination only. The remedial actions planned at each of the Operable Units at Homestead ARB are, to the extent practicable, independent of one another. However, with respect to OU-5 and OU-9 (Boundary Canal), the definition of these two operable units has resulted in the necesity to assign different physical media to each operable unit. Consequently, all sediment and surface water samples collected in association with investigations conducted at OU-5 have been evaluated in the OU-9, Boundary Canal RI through ROD.

2.6 SUMMARY OF SITE CHARACTERISTICS

OU-5/Site WP-1 was formerly used as a disposal area for spent plating baths and rinses from a plating shop located in Building 164. During the period between 1946 and 1953, when Homestead ARB was inactive and ownership of the property was transferred to Dade County, a small electroplating operation was located in Building 164. Spent plating solutions containing chromium, nickel, copper, and sulfuric and hydrochloric acid were routinely disposed of by discarding them on the ground in an area just east of Building 164 (Engineering-Science, 1983). Wastes were generated at a rate of approximately 250 gallons per month, and the electroplating operation continued for about two years (Engineering-Science, 1983). According to 1958, 1962, and 1973 aerial photographs, the asphalt parking lot located east of Building 164 did not exist when the electroplating company was operating and the area was covered with trees and grass. The parking lot was constructed between 1962 and 1973. No visible evidence of waste residue is observed on the ground surfaces that are presently exposed (not covered by parking lot). The high amount of rainfall typical of the area is suspected to have dissipated the waste residues.

2.6.1 Nature and Extent of Contamination

This section describes the nature and extent of contamination defined to-date at the Electroplating Waste Disposal Area, OU-5/Site WP-1. Subsurface investigations at the site were initiated by SAIC in 1984 (SAIC, 1986). Further field investigations were conducted by G&M in 1988 and 1991 following IRP and CERCLA directives. The results of the 1984 activities are reported in *"Installation Restoration Program Phase II - Confirmation/Quantification, Stage I, Homestead AFB, Florida"* prepared by SAIC. The 1991 investigative results are reported in *"Remedial Investigation Report for Site WP-1, Electroplating Waste Disposal Area, July 1992"* (G&M, 1992). Based on recommendations following the 1992 RI report by G&M, Montgomery Watson performed an additional field and sampling investigations in 1993. The Montgomery Watson investigation was conducted in accordance with the approved *Facility Work Plan and Work Plan Addenda* (G&M, 1991a,b,c).

In 1994, IT Corporation completed a soil and sediment confirmation sampling program at the Electroplating Waste Disposal Area, OU-5/Site WP-1. This confirmation sampling program was developed to further define the nature and extent of contamination at the OU-5/Site WP1 area. Based on the findings of this investigation, excavation of sediments from the northern swale was performed (Figure 2-6). Upon completion of the excavation, three confirmation soil/sediment samples were collected from the base of each of the three excavations in the North Area. Additionally, groundwater samples were collected from each of the 4 site monitoring wells in 1996 by OHM Corporation. This action was completed based on USEPA recommendations to confirm the groundwater quality of the site given there has been a three year delay between sampling and reporting.

A summary of the scope of previous investigations, including those associated with the 1994 and 1995 Interim Action, the 1996 Groundwater Sampling event, and a discussion of data collected to-date at OU-5/Site WP-1 are presented below.

An IRP Phase II investigation was completed by SAIC in 1984. The Phase II Confirmation Quantification investigation included installation of three monitoring wells (I-01 through I-03) and four soil borings (SL-1 through SL-4), collection of two sediment samples (SD-1 and SD-2), and sampling of groundwater. All of the samples were analyzed for total metals (cadmium, chromium, copper, lead, nickel, and zinc) and cyanide. The groundwater was also analyzed for hexavalent chromium. Locations of the wells and soil borings are shown in Figure 2-2. Analytical results for this sampling are discussed in Section 2.3.2.

Based on the presence of contamination at the site, a Phase IV-A investigation was performed in 1988 by G&M to further define the extent and degree of contamination. The 1988 work included groundwater analysis, water-level measurements, and a topographic survey. The three existing wells were sampled and analyzed for total metals (arsenic, barium, cadmium, chromium, copper, lead, mercury, nickel, selenium, silver, sodium) and cyanide. Results of the 1988 groundwater sampling are discussed in Section 2.3.4. Well locations are shown on Figure 2-2.

2.6.1.1 1991 Investigation. In 1991 additional CERCLA field investigations were performed by G&M to evaluate groundwater and soil quality with respect to the U.S. Environmental Protection Agency (USEPA) Contract Laboratory Program (CLP) analyte list for organics (TCL) and inorganics (TAL). This field effort included the installation of one shallow monitoring well (SP1 -MW-0001) and the collection of groundwater samples from this new well and from the three existing wells (I-01, I-02, and I-03) at OU-5/Site WP-1. Monitoring Well SP1-MW-0001 was completed to 13 ft bls at a location east of Building 164 approximately midway between existing monitoring wells I-02 and I-03. The well is located near the SL-3 sampling location of the 1984 Phase II investigation; this location is the suspected center of the waste disposal area. Three shallow soil/weathered rock samples (SP1-SL-0001, SP1-SL-0002, SP1-SL-0003; plus SP1-SL-9002, the duplicate of SP1-SL-0002) were collected. In addition, one soil sample (SP1-MW-0001-S) from the monitoring well boring was collected. All soil samples were collected from the 2-4 ft bls interval. Two sediment and two surface water samples were collected from the unlined drainage swale south of the equipment storage area. The sediment/surface water pairs were identified as SP1-SW/SD-0005 and SP1-SW/SD-0006. Locations of the water and soil/sediment sampling points are illustrated on Figure 2-3. All samples were analyzed for the Target Analyte List (TAL) and the Target Compound List (TCL) elements or compounds excluding pesticides/PCBs. USEPA SW846 methodologies were employed for all analyses performed in 1991. Analytical results for soil, groundwater, sediment, and surface water samples are discussed in Sections 2.6.2, 2.6.3, 2.6.4 and 2.6.5, respectively. Tables 2-5, 2-6, 2-7, and 2-8 provide a summary of analytical results for the 1991 investigation.

2.6.1.2 1993 Investigation. In 1993, Montgomery Watson performed additional investigations of soil, sediment, and groundwater at OU-5/Site WP-1. Surface water was not present in any of the drainage swales during the 1993 sampling activities; therefore, it could not be collected. The Montgomery Watson field effort included drilling four soil borings (SP1-SL-0004 through SP1-SL-0007). These borings were in the same relative locations as SP1-SL-0001 through SP1-SL-0003 and SP1-MW-0001-S, drilled during the 1991

SUMMARY OF ANALYTICAL CONSTITUENTS DETECTED IN SOILS SITE WP-1/OU-6, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homesteed Air Reserve Base, Florida

G Analyla	&M Sample LD. Laboratory LD. Sampling Data	Average Carbonate Composition(b)	Average Homestand AFB Background 2-4 ft ble		SP1-SL-0002-epik 36713-1 10/22/91	Composite Background Sample(a)	SP1-8L-0001 13950-001 10/22/91	SP1-8L-0003 13950-005 10/22/91	\$P1-MW-0001-4 13950-007A 10/22/91
VOLATILE ORGANIC COMPC	JUNDS (ug/Kg dw):								
Acelone		-	1028.97	1257.94	960 J	1,258	-6	- 5. 95	7301.33
BASE/NEUTRAL and ACID ED	TRACTABLE COM	POUNDS (ug/Kg d	wj:						
Benzo(a)anthracene		-	21.5	<625	[25]	[25]	<581	<595	-602
Benzo(a)pyrane		-	22	<1250	[30]	[30]	<1163	<1190	<1205
Benzo(b)fluoranthene		-	37	<2500	[37]	[37]	<2326	<2361	<2410
Benzo(g.h.i)perviene		-	8.1	<2500	[8.1]	[8.1]	<2326	- 2381	<2410
Benzo(k)Buoranthene		-	44	<2500	[44]	[44]	<2326	<2381	<2410
bis(2-Ethythexyl) phthalai	•	-	94	<375	[110]	[110]	<349	< 357	<361
Chrysene		-	45	<625	[46]	[46]	<581	<595	<602
2-Chlorophenol		-	14	⊲625	[10]	[1 0]	<581	<595	<602
Fluoranthene		-	41	-625	[55]	(55)	<581	<595	<602
Phenanthrone		-	10	<625	[10]	[10]	<581		<602
Pyrane		-	42	<375	[58]	[56]	<349	<357	<361
METALS (mg/Kg dw):									
Aluminum		8.970	1,187	1050	685	1,050	728	9270	1710
Arsenic		1.8	0.61	0.61	<1.1	0.61	0.26	1.5	0.83
Sanum		30	6.6	9.3	8.0 J	9.3	5.3	14.1	6.6
Barylium		0.18	ND	⊲0.12	<2.9	⊲0.12	<0.11	0.31	<0.12
Calcium		272,000	381,443	464239	420000	464,239	537638	490772	646718
Chromiura	······································	7.1	7.6	9.5	6.8	9.5	7.2	23.4	9.2
Cobalt		0.12	ND	⊲0.36	<1.2	<0.38	<0.34	1	⊲0.35
Copper		4.4	1.8	1.8	<2.9	1.8	1.7	2.2	2.6
tran		8,190	621	584	320	584	302	5090	946
Lead		16	3.3	4.3	4.7 J	4.75	0.64	5.5	2.4
Magnesium		45,300	1,140	1220	1200 J	1,220	1450	1500	1350
Manganese		842	17	14.2	12 J	14.2	5.1	139	14.8
Mercury		0.046	0.011	⊲0.05	0.011 J	0.011 J	<0.06	⊲0.06	-0.05
METALS (mg/Kg dw) cont'd:									
Nickel		13	1	1.1	<4.6	1.1	0.87	5.2	2.8
Potassium		2,390	33	33.4	<120	33.4	28	123	36.3
Sodium		393	657	741	620	741	689	605	652
Vanadium		13	6	8.2	<5.8	8.2	8.3	14.7	9.1
Zinc		16	3	3.3	<12	3.3	2.2	2.9	2.9
CYANIDE (mg/Kg dw)		-	0.5	<0.5	<1.2	<0.5	<0.5	-0.5	⊲0.\$

Source: Geraghty & Miller, Inc., 1992

ug/kg dw micrograms per kilogram dry weight

mg/kg dw milligrams per kilogram dry weight

Value is greater than instrument detection limit but less than practical-quantitation limit. []

Analyte was not detected at or above the indicated concentration. <

J Positive result has been classified as qualitative.

Background Samples are SP1-SL-3002 values and SP1-SL-0002-Split. These values are a combination of both samples. a

If both were detects, the lower measured analytical concentration was used. If only one result was positive that concentration was used. Hern, 1989 Could not be determined.

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SUMMARY OF CONSTITUENTS DETECTED IN GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homestead Air Reserve Base, Florida

Analyte	Florida Drinking Water Slandards	(EPA Maxium Contaminan Level	1	EPA Maximum Contaminant Level Goal		of CONS	ange TITUENT ected Max.		5P1-MW-000 37428-18 11/13/91)1		SP[1]-I-01 37460-4 11/14/91	ļ	SP1-1-01 37508-14 11/18/91	3	P1-1-02 7460-1 1/14/91		ļ	\$P1-1-03 37428-17 11/13/91	1		P1-1-9002 37460-2 11/14/91			P1-EB-0023 37460-3 11/14/91
VOLATILE ORGANIC COMPOUNDS (ug Methylene chloride	/L): 5		NS		NS		BDL	. [3.1]	< 4	5.0			(1.4)		NA		[3.1]		۲	5.0		<	5.0		۲	5.0
BASE/NEUTRAL and ACID EXTRACTA	BLE COMPOU	NDS	(ug/L):				001	200		10.71			10		NA		(0,9)	J		320	J	<	10		<	10
bis(2.Ethylhexyl) phthalate	6		6		0		BDL	- 320 J		[0.7]		¢	10				[v,a]									
METALS (ug/L):								- 24000		0000			5700	J	NA		24000			3000			6400		<	200
Aluminum	200	1	50-200	i,h	NS		01000	- 24000 - 92J		3000			92	J	NA	(60	1		19	J		42	3	<	10
Arsenic	50	k	50	9	NS	,	11 J	· 52.5 · 150		11	J		92 63	4	NA		150	v		25	·		89	•	<	10
Barium	2,000	k	2,000	l,g	2,000	1	25	- 150 - 80L		33			63 25	ູ	NA	<	25		~			¢	25		<	5.0
Cadmium	5	K	5	1	5	1	BDL	· 540000	^ <			۲	25 3100000		NA		400000	n	•	130000			2600000			370
Calcium	NS		NS		NS		130000	· 540000	v	130000			59	1.0	NA	5	130			34			62		٢	10
Chromium	100	k	100	i,g	100	I	-	40000	•	22			3600	J	NA		18000	J		1800	J		4900	J	<	50
Iron	300		300	h 	NS		1700 J 8.6 J		J	<u> </u>	<u></u>		11		NA	_	30	Ĵ			Ĵ	_	30	J	<	5.0
Lead	15	K	15	đ	0			1 1000			J		7300	J	NA		14000	•		3800	•		6600		<	50
Magnesium	NS		NS		NS		33 J	- 14000 - 200 J		3900 53			75	J	NA		200	J		33	J		110	J	¢	10
Manganese	50	ļ	50	ħ	NS		33 J 1600 J	- 3900		3900	J		1600	1	NA		2600	•		2100	·		2000		¢	1000
Potassium	NS		NS	•.	NS	ŧ		• BDL				<	50	UJ	NA	~	50	W	ć	50	W	۲	50	IJ	¢	10
Selenium	50	K	50	i,g	•	I			<	50 18000	03	•	23000	Q.	NA		26000	••	•	20000			20000		<	500
Sodium	160,000	k	NS	_	NS			·					50	UJ	NA		82		-	12		<	50	IJ	<	10
Vanadium	NS		NS	F	NS			•		12 25		< <	100	W	NÅ	<	100			26		۲	100		ĸ	20
Zinc	5,000	I	5,000	h	NS			•		23		۰,	IVV	~		-	•••									
CYANIDE	200	k	200	n	200	ņ			<	10			NA		10	<	10		<	10		<	10		¢	10
TOTAL DISSOLVED SOLIDS (mg/L)	500	1	500	h	NS					NA			NA		NA		300			NA			290		<	5.0

SUMMARY OF CONSTITUENTS DETECTED IN GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homestead Air Reserve Base, Florida

Analyte	Florida Drinking Water Standards	(EPA Maxluffi Contaminan Level	¢ 1	EPA Maximum Contaminani Level Goel		CONS	inge TITUENTS lected Max.		B-1114-SP1 37460-5 11/14/91	TB-1113-SP4 37428-6 11/14/91	
VOLATILE ORGANIC COMPOUNDS (ug Methylene chioride	y'L): 5		N9		NS				e	5.0	NA	
ASE/NEUTRAL and ACID EXTRACTA bis(2-Ethylhexyl) phthelele	BLE COMPOU	NDS ((ug/L): 6		0					NA	NA	
NETALS (Ug/L):												
Atuminum	200	- 1	50-200	i,h	NS					NA	NA	
Arsenic	50	k	50	9	NS					NA	NA	
Barium	2,000	k	2,000	i,g	2,000	1				NA	NA	
Cadmium	5	k	5	1	5	ł.				NA	NA	
Calcium	NS		NS		NS					NA	NA	
Chromium	100	k	100	I,g	100	1			ĸ	NA	NĂ	
tron	300	1	300	h	NS					<u>NA</u>	NA	
Lead	15	k	15	Б	0					NA	NA	
Magnesium	NS		NS		NS					NA	NA	
Manganese	50	1	50	h	NS					NA	NA	
Potassium	NS		NS		NS					NA	NA	
Selenium	50	k	50	I.g	50	ł				NA	NA	
Sodium	160,000	k	NS		NS					NA	NA	
Vanadium	NS		N\$		NS					NA	NA	
Zinc	5,000	1	5,000	h	NS					NA	NA	
YANIDE	200	k	200	n	200	n				NĂ	NA	
										NA	NA	
TOTAL DISSOLVED SOLIDS (mg/L) Source: Geraphty & Miller, Inc., 1992												

NOTES:

d Federal Action Level - the action level is exceeded if the level of copper/lead in more than 10% of targeted tap samples is greater than the action level.

I Numbers represent EPA's Proposed Primary MCL, Federal Register, Vol. 55, No. 143, July 1990.

g Numbers represent EPA's Primary MCL for Inorganics.

h Numbers represent FPA's Secondary MCL for inorganics which are non-enforceable taste, odor or appearance guidelines.

Numbers represent EPA's Final MCL effective July 1992, Federal Register, January 30, 1991.

J Positive result has been classified as qualitative.

k Florida Primary Drinking Water Standard.

I Fiorida Secondary Drinking Water Standard.

m Florida Ground-Water Buildance CONSTITUENTS for Minimum Criteria Requirements (Rule 62-3.402,FAC), Florida Department of Environmental Regulation, February 1989.

n Proposed Primary MCL or MCLG

NS No Standard Available

NA Not analyzed.

U Classified as undetected.

UJ Analyte was not detected at or above the indicated concentration and has not been classified as qualitative.

[] Value is greater than instrument detection limit but less than practical quantitation limit.

SUMMARY OF CONSTITUENTS DETECTED IN SEDIMENT SAMPLES COLLECTED IN 1991 SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homestead Air Reserve Base, Florida

alyte	G&M Sample I.D.	SQC (ug/Kg Ot ^{1/}	SQC at 1% OC	NOAA ER-L ^{3/}	NOAA ER-M ≮		MERCHAR		Trip Blank #2	
	Sampling Date		(ug/Kg)	ug/Kg	ug/Kg				08/28/91	Background Sample
OLATILE OR	SANIC COMPOUND	<u>S (ua/ka dw):</u>								
Acetone		NS	NS	N	N		U State in the second		[4.7]	NA
Methylene chio	ide	, NS	NS	N	N K				[1.5]	NA
Tetrachioroethe	ne	NS	NS	N	N	< 10.00 (A)			< 5.0	NA
Trichloroethene		NS	NS	N	N	< 10.00 Arts	< 10 million		< 5.0	NA
ASE/NEUTRA	L and ACID EXTRA	CTABLE COMP	OUNDS (ug/k	g dw):						
Acenaphthene		732,000	7,320	150	650		< Distant		NA <	1400
Anthracene		NS	NS	85	960				NA <	1400
Benzo(a)anthra	cene	1,317,000	13,170	230	1600				NA <	1400
Benzo(a)pyrane	i	1,063,000	10,630	400	2500		TATISANCE INC.	S FROM	NA <	
Benzo(b)fluorar	thene	NS	NS	N	N		B # 1420001 - B	12000 j. 1	NA <	
Benzo(g,h,i)per	/lena	NS	NS	N	N				NA <	1400
Benzo(k)lluoran	thene	NS	NS	N	N				NA <	1400
bis(2-Ethylhexy) phthalate	NS	NS	N	N				NA <	1400
Bulyibenzyiphih	elale	NS	NS	N	N				NA <	1400
Chrysene		NS	NŞ	400	2800				NA <	1400
Dibenzo(e,h)ani	hracene	NS	NS	60	260	1. 化价格	<		NA <	1400
Dibenzofuran		NS	NS	N	N				NA <	1400
Fluoranthene		1,883,000	18,830	600	3600				NA <	1400
Fluorene		NS	NS	35	640				NA <	1400
Indeno (1.2,3-c	l)pyrene	NS	NS	N	N	ST PRODUCT	Mar 11700714	ing and 1	NA <	1400
Phenanthrene		139,000	1,390	225	1380				NA «	1400
Pyrene		1,311,000	13,110	350	2200	1. 人民 秋谷			NA <	1400

SUMMARY OF CONSTITUENTS DETECTED IN SEDIMENT SAMPLES COLLECTED IN 1991 SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homestead Air Reserve Base, Florida

nalyte	G&M Sample I.D.	SQC (ug/Kg Oi 1/	SQC at 1% OC	NOAA ER-L ^{3/} ug/Kg	NOAA ER-M ^{4/} ug/Kg	(313) (513) (477) (513)			Trip Blank #2 06/26/91	BC-SD-010 Background Sample
. <u></u>	Sampling Date		(ug/Kg)	ugring					00/20/91	Sample
METALS (mg/	<u>kg dw):</u>						2007-00-00-00-00-00-00-00-00-00-00-00-00-	28 		
Aluminum	1	NS	NS	N	N	125312	S 1400		NA	2700
Arsenic		NS	NS	33	85				NA	2
Barium		NS	NS	N	N	이 문화 및 가 관		(-P) = (-1) + (-1)	NA	14
Cadmium		NS	NS	5	9	s. halles		EQ.331	<u>NA <</u>	2.1
Calcium		NS	NS	N	N		C 180000 81	1 000 TO 00	NA	310000
Chromium		NS	NS	80	145		「読む」である	a sa ing sa ing sa	NA	11
Cobalt		NS	NS	N	N	<			NA <	4.2
Copper		NS	NS	70	390 5	J CONTRACT			NA	16
Iron	.*	NS	NS	N	N		11.379	200,2200,414	NA	1700
Lead		NS	NS	35	110				NA	11
Magnesium		NS	NS	N	N	38 9 X 6 S			NA	1000
Manganese		NS	NS	N	N	L SAME I		HAR BURN J	NA <	29
Mercury		NS	NS	0.15	1.3		1411316) I	2 18 2 M S	NA	0.043
Sodium		NS	NS	N	N	11. 在19 · · · ·	144 A 14	all and all a	NA	290
Variadium		NS	NS	N	N	l k stern 3 n	PARTY AND		NA	5.7
Zinc		NS	NS	120	270			HANNING J	NA	27
YANIDE		NS	NS	N	N	< : 李 法的建议		< 41.5. 11.19.24	NA	NA

÷.

ug/kg dw micrograms per kilogram dry weight

mg/kg dw milligrams per kilogram dry weight

NA Not Analyzed

< Analyte was not detected at or above the indicated concentration.

[] Value is greater than instrument detection limit but less than pactical quantitation limit.

- J Positive result has been classified as qualitative.
- UJ Analyte was undetected. Classified as qualitative.
- U Classified as undetected.
- N Not Available
- NS No Standard
- 1/ Organic Carbon
- 2/ The sediment quality criteria (SQC) cannot be directly compared with the drainage swale data because the SQC are presented as normalized to organic carbon (i.e. presented on a per organic carbon weight basis). To allow a direct comparison between the sediment data and SQC, the SQC for an organic carbon content of 1.0% OC were calculated. The SQC (ug/Kg) at 1% OC were derived by multiplying the SQC (ug/Kg OC) by an OC content of 1% (.01 Kg of OC/Kg of sediment).
- 3/ National Oceanic and Atmospheric Administration (NOAA) Technical Memorandum NOS OMA 52; ER-L is effects Range Low (10th percnetile).
- 4/ Effects Range Median (50th percentile).
- Source sediments may have been removed during the 1995 Interim Removal Action. However, DATA MAY STILL REPRESENT SITE CONDITIONS.

SUMMARY OF CONSTITUENTS DETECTED IN SURFACE-WATER SAMPLES COLLECTED IN 1991 SITE WP-1/DU-5, ELECTROPLATING WASTE DISPOSAL AREA GERAGHTY & MILLER, 1991 INVESTIGATION Homestand Air Resorve Base, Floride

Analyle	G&M Sample I.D. Savannsh I.D. Sampling Date	Florida Surface-Water Quality Standard d/	Federal Water Ouelity Criterion	Trip Blank 35504-6 06/28/91	!	SP1-EB-0002 35504-2 08/28/91	5	SP1-SW-0 35504-: 06/26/9)	SP1-SW-0 35504-1 08/28/9		5	P1-5W-9 35504-/ 06/28/9	4
OLATILE ORGANIC	COMPOUNDS (ug/L):				_							···		_
Acetone		NS	NS	(4.7)		(3.7)		17.81	υ	(5.5)	U		[9.2]	U
1,1-Dichloroet	1010	NS	NS	[1.5]	<	5.0	*	5.0	•	< 5.0	•	٠	5.0	Ŭ
Methylene chic	vide	NS	NS	< 5.0		[2.0]		[1.7]	υ	[1,1]	U	-	[0.86]	U
ASE/NEUTRAL and	ACID EXTRACTABLE C	OMPOUNDS (ug/L):												
Anilvacene		NS	NS	NA	•	5.0	<	5.0		[0.96]		<	5.0	
Senzo(s)anthr	Icene	NS	NS	NA		5.0	Ì	5.0		15		, ,	5.0	
Benzo(a)pyren	•	NS	NS	NA		5.0	k	5.0		14		, ,	5.0	
Benzo(b)fluora	nthene	NS	NS	NA	<	5.0	`.	5.0		22	<u>. </u>	<u> </u>	[2.0]	
Benzo(k)fluora	nhene	NS	NS	NA	ć	5.0	è	5.0		17		<	(2.07 5.0	
bis(2-Ethylhexy	() phihalate	NS	NS	NA		5.0	, i	5.0		< 5.0			[3.0]	
Chrysene		NS	NS	NA		5.0	Ì	5.0		26		~	5.0	
Fluoranthene		NS	NS	NA		10	-	(3.6)	U	41	U	•	(3.6)	U
Phenanthrene		NS	6.3 .	:NA	<	5.0		5.0	<u> </u>	[6.5]			5.0	<u> </u>
Pyrene		NS	NS	NA	٠	5.0		[1.8]	5	27	J	•	[2.4]	J
ETALS (ug/L):			······	· · · · · · · · · · · · · · · · · · ·			<u> </u>						(=. •)	-
Aluminum		NS	NS	NA	۲.	200	۲	200		3000		•		
Bartum		NS	NS	NA		10	Ì	10		170			200 10	
Cadmium		1.2 b	2	NA		0.60	Ż	0.80	IJ	9.7	L	<	0.80	
Celcum		NS	NS	NA		50	•	14000	00	250000	3	<	13000	U
Chromium		50	11(IV), 424 (III) c	NA		10	٠	10		42		<	10	
Copper		30	25	NA	•	2.0		8.6		92		•	8.1	J
kron		1000	1000	NA	~	50		170		4100			190	
Leed		30	9.7 c	NA	٠	5.0		12	j	230	J		15	L
Magnesium		NS	NS	NA	<	50		270	•	2900	•		300	
Manganese		NS	NS	NA		10	۲	10		83	J	٠	10	
Mercury		0.2	0.012	NA	٠	0.10	<	0.10	IJ	0.54	j.	~	0.10	U.
Potassium		NS	NS	NA	; <	1000	~ ~	1000		2900		<u></u>	1000	
Silver		0.07	16	NA	\ <	0.070	<	0.070	υJ	0 47		è	0.070	
Sodium		NS	NS	NA	•	500		1300		21000		•	1100	
Vanadium		NS	NS	NA	٠	10	۲	10		17		<	10	
Zinc		30	223 c	NA		26		64	υ	820		-	63	υ
ANIDE (mg/L)		5	5.2	NA		10							<u> </u>	<u> </u>

Notes:

Source: Geraghty & Miller 1991

ug/L. micrograms per Ner

- mg/L milligrams per liter
- NA Not Analyzed
- [] Value is greater than instrument detection limit but less than pactical quantitation limit.
- Analyte was not detected at or above the indicated concentration
- J Positive result has been classified as qualitative.
- UJ Analyte was not detected. Classified as qualitative.
- U Classified as undetected.
- NS No Standard
- e/ Proposed Chronic Federal Water Quality Criterion; USEPA, 1988
- b/ For herdness greater than 150 mg/L.
- c/ Federal Water Quality Cnterion Calculated using a hardness value of 240 mp/L.
- d/ Florida Surface-water Quality Standards for Class III Surface Waters (Rule 17-302.560 & 17-302.510 FAC).

investigations by G&M. Soil/weathered rock samples were collected from each of the 1993 borings at the 0-1 or 0-2 ft bls intervals. Groundwater samples were collected from the four shallow monitoring wells at OU-5/Site WP-1. A total of six sediment samples were collected from the three drainage swales. Sediment thickness varied with sampling location and were encountered underlying several inches of benthic organic material such as algae in the drainage swales. Figure 2-4 illustrates the locations of all samples collected during this current investigation. Each matrix was analyzed using USEPA CLP protocols for TCL organics and TAL inorganics. All samples were analyzed for TCL organochlorine pesticides/PCBs and cyanide. All sediments, one soil and one groundwater sample were also analyzed for the presence of TCL VOCs, BNAs and TAL metals. Dissolved (filtered) TAL metals were also analyzed for in groundwater samples. All samples were analyzed by Savannah Laboratories, Tallahassee, Florida. Analytical results of the montgomery Watson sampling are discussed in Section 2.6.2 and 2.6.3, and 2.6.4. Tables 2-9, 2-10, and 2-11 provide a summary of analytical results for the 1993 investigation.

2.6.1.3 1994 And 1995 Investigation. The 1994 Interim Action was performed in accordance with Section 300.415 (b) of the National Contingency Plan (NCP) and CERCLA. The purpose of this non-time critical removal action was to clean-up and remove contaminated media in order to prevent damage to the public health or welfare of the environment. Prior to the removal action, an engineering evaluation/cost analysis (EE/CA) was completed which identified the objectives of the removal action and ananlyzed the various alternatives that were available to satisfy the objectives for cost, effectiveness, and implemtability. From this evaluation, the recommended alternative is implemented.

In 1994, IT Corporation conducted an investigation of affected soils/sediment at OU-5/Site WP-1. This work was performed under contract to USACE - Mobile District. The investigation consisted of completing 28 soil borings at the locations identified on Figure 2-5. Two borings, EWA-1 and EWA-2, were completed in the paved parking area west of Building 164. These borings were extended to a depth of 2 ft below the asphalt. The remaining 26 borings were located at points within the drainage swales and advanced to a depth of six inches. The draining swales have been divided in to a North and a South Area. The North Area is located between Building 164 and 153. The South Area is located between Building 179 and 185. Soil/sediment samples were collected from each boring to obtain a sample for laboratory analysis. The samples collected from the two borings drilled west of Building 164 and the five northern most borings (EWA-3 through EWA-7) are considered soil/ weathered rock samples. These five northern most points were considered to be representative of soil/weathered rock due to the absence of a well defined drainage system

SUMMARY OF CONSTITUENTS DETECTED IN SOILS SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA 1993, 1994, AND 1995 INVESTIGATIONS Homestead Air Reserve Base, Florida

	1995	Removal	Sample ID.	MW 1993 SPI-SL-0005	MW 1993 SP1-SL-0006	MW 1993 SPI-SL-0007	MW 1993 SP1-SL-9007
Analyte	FDEP Health Based	Action Level	Sample Interval (ft)	(0-2)	(0-2)	(0-1)	(0-1)
	Soil Target Levels		Date Collected	3/11/93	3/11/93	3/11/93	DUPLICATE
VOA TCL Compounds (ug/kg)	1,800,000	ND (1)		NA	NA	21000	27000
Aceione	1,800,000						
Pesticides/PCB TCL Compounds (ug/kg)						1	
Alpha-Chlordane	3.000	3.210		<2.0	1.9 J	<1.9	<1.9
Beta-Chiordane	3,000	3,210		<2.0	2.3 P	<1.9	<1.9
Chlordane (technical)	3,000	3,210		NA	NA	NA	NA
Endosulfan Sulfate	5.900,000	ND (1)		<3.8	0.77 J	6.4	8.8 P
p.pDDD	17.000	17,500		<38	28	0.75 J	141
p.p DDE	11,000	12,400		3.9	130	3.8	6.2
p.p [·] DDT	12.000	11,300		0. 83 J	34	6.3	20
BNA TCL Compounds (ug/kg)	30 000 000	1 000 (3)		NA	NA	51	171
Acenaphthene	30,000,000	1.000 (2)		NA	NA	25.1	391
Anthracene	300,000,000	1,000 (2)		NA NA	NA	260 1	320 1
Benzo(a)Anthracene	4,900	5,040			NA	300 1	360 J
Benzo(a)Pyrene	500	540		NA	NA	380	540
Benzo(b)Fluoranthene	5.000	5,010		NA		190 J	-
Benzo(g.h.i)Perylene	50.000	ND (1)		NA	NA	1	230 J
Benzo(k)Fluoranthene	48,000	4,970		NA	NA	390	420
Benzylbutyl Phthalate	310.000,000	ND (1)		NA	NA	24 J	143
Bis(2-Ethylexyl) Phthalate	110,000	ND (1)		NA	NA	93 BJ	79 BJ
Carbazole	120.000	224,000		NA	NA	351	62)
Chrysene	500.000	50.300		NA	NA	410	54(
Di-n-Butyl Phthalate	140.000.000	ND (1)		NA	NA	17 BJ	14 B)
Di-n-Octyl Phthalate	32,000.00	ND (1)		NA	NA	7 J	<360
Dibenz(a,h)Anthracene	500	505		NA	NA	87 J	68 J
Dibenzofuran	3,500,000	1,000 (2)		NA	NA	4.1	11.1
Fluoranthene	48,000,000	1,000 (2)		NA	NA	650	880
'10 16116	30,000,000	1,000 (2)		NA	NA	18	20 J
eno(1,2,3-C,D)Pyrene	5,000	5,040		NA	NA	200 J	230 J
, enanthrene	21,000.000	1,000 (2)		NA	NA	190 J	340 J
Pyrene	41,000,000	1,000 (2)		NA NA	NA	510	810
				8	+		
Metais (mg/kg)		1		ă		5.440	6,430
Aluminum	>1,000,000,000	ND (1)			NA	5,460	9.7 N
Arsenic	3	10 (3)			NA	9.1 N 17.7 B*	21.2 B*
Banum	4,000	4,940			NA	1	0 42 B
Cadmium	600	1,070		NA S	NA	<0.40 285,000	296.000
Calcium	ND (1)	ND (1)			NA	1	18.9*
Chromium, Total	430	160		NA NA	NA	18.6*	
Cobalt	110,000	ND (1)		NA NA	NA	0.62 B	0.66 B
Copper	ND (1)	ND (1)		S NA	NA	13.8	15.5 3210 E
Iron	ND (1)	ND (1)			NA	3070 E	
Lead	1,000	108		NA NA	NA	38.3*	44.5*
Magnesium	ND (1)	ND (1)		NA	NA	1,160	1,230
Manganese	5,500	ND (1)		NA NA	NA	72.4	65.1
Mercury	480	17		NA	NA	<0.11	<0.14
Nickel	2.600	3.24		NA	NA	7.2 B	6.1 B
Potassium	ND (1)	ND (1)		NA NA	NA	680 B	680 B
Sodium	ND (1)	ND (1)		NA NA	NA	513 B	468 B
Vanadium	4,800	ND (I)		S NA	NA	8.4 B	9.4 B
Zinc	560,000	ND (I)		8 NA	NA	20.1 E	23. 3 E
1	1	1		8			

< not detected at specified detection limit

NE - not established

NA - not analyzed

ND (1) - no data

(2) - not listed on the Soil Target Level Table

but was listed in 62-775 of the FAC. Total VOC C - Confirmed on second column

listed in 62-775 as having a max conc. of

bug/kg and 1 mg/kg for Total PAHs.

Removal Action Level as determined by BC T uld - exceeds Action Level

Bold & Shaded - exceeds Action Level and

FDEP Soil Target Level

** - 2 samples used to calculate mean

Data Qualifiers for Organic Compounds

J - Estimated value, <CRQL

P - >25 % difference in detected value between two columns

B - Compound detected in an associated blank

Data Qualifiers for Inorganic Compounds

B - Reading is less than CRQL but greater than IDL

E - reported value is estimated due to interference

N - spiked sample recovery not within control limits

* - duplicate analysis not within control limits

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SUMMARY OF CONSTITUENTS DETECTED IN SOILS SITE WP-IAOU-5, ELECTROPLATING WASTE DISPOSAL AREA 1993, 1994, AND 1995 INVESTIGATIONS Homestend Air Reserve Base, Florida

ſ	T	l		IT 1994	IT 1994	1T 1994	IT 1994	IT 1994	IT 1995
1	1995	Removal	Sample ID.	EWAI	EWA2	EWAJ	EWA4	EWAS	CS-01
Analyte	FDEP Health Based	Action Level	Sample Interval (ft)	(0-2)	(0-2)	(0-0.5)	(0-0_5)	(0-0.5)	(0-1)
	Soil Target Levels		Date Collected						
VOA TCL Compounds (ug/kg)									
Acetone	1.800.000	ND (1)		NA	NA	NA	NA	NA	NA
Pesticides/PCB TCL Compounds (ug/kg)]				
Alpha-Chlordane	3,000	3,210		NA	NA .	NA	NA	NA	NA
Beta-Chlordane	3,000	3,210		NA	NA	NA	NA	NA	NA
Chlordane (technical)	3,000	3,210		NA	NA	<50	42	200	1400
Endosulfan Sulfate	5,900,000	ND (I)		NA	NA	<200	<20	<40	<3.3
p,p'DDD	17,000	17,500		NA	NA	9.1 J.C	<4	5.5 C	1400
p.pDDE	11,000	12,400		NA	NA	240 C	7.8 C	25 C	89
p.pDDT	12,000	11,300		NA	NA	22 J.C	<4	13 C	<3.3
BNA TCL Compounds (ug/kg)									
Acenaphthene	30,000,000	1,000 (2)		<7500	<3700	<110000	<7200	<3700	<452
Anihracene	300.000.000	1,000 (2)		<7500	<3700	<110000	<7200	<3700	<452
Benzo(a)Anthracene	4,900	5,040		<7500	<3700	<110000	<7200	<3700	691
Benzo(a)Pyrene	500	540		<7500	3700	<110000	<7200	<3700	460
Benzo(b)Fluoranthene	5,000	5,010		<7500	<3700	<110000	<7200	<3700	530
Benzo(c.h.i)Perviene	50,000	ND (1)		<7500	<3700	<110000	<7200	<3700	420
Benzo(k)Fluoranthene	48,000	4.970		<7500	<3700	<110000	<7200	<3700	460
Benzylbutyl Phthalate	310,000,000	ND (1)		<7500	<3700	<110000	<7200	<3700	<452
Benzylouryl Phinalate Bis(2-Ethylexyl) Phihalate	110,000	ND(1)		<7500	<3700	<110000	<7200	<3700	<452
								t	
Carbazole	120,000	224,000		NA	NA	NA	NA <7200	NA <3700	NA 370
Chrysene	500,000	50,300		<7500	<3700	<110000 <110000		<3700	
Di-n-Butyl Phthalate	140,000,000	ND (1)		<7500	<3700		<7200	<3700	460
Di-n-Octyl Phthalate	32,000.00	ND (1)		<7500	<3700	<110000	<7200	<3700	<452 <452
Dibenz(a,h)Anthracene	500	505		<7500	<3700	<110000	<7200	<3700	<452
Dibenzofuran	3,500,000	1,000 (2)		<7500	<3700	<110000	<7200	1	
Fluoranthene	48.000,000	1,000 (2)		<7500	<3700	000011>	<7200	<3700	1200
1	30,000,000	1,000 (2)		<7500	<3700	<110000	<7200	<3700	<452
,2.3-C.D)Pyrene	5.000	5,040		<7500	<3700	<110000	<7200	<3700	<452
drene	21,000,000	1.000 (2)		<7500	<3700	<110000	<7200	<3700	490
Pyrene	41.000,000	1,000 (2)		<7500	<3700	<110000	<7200	<3700	710
Metals (mg/kg)									
Alumnum	>1,000,000,000	ND (1)		1.670	3010	6640	1250	1710	1600
Arsenic	3	10 (3)		0.43 B	<0.45	<1.5	0.56 B	<0.41	6.9
Banum	4,000	4,940		5.0 B	9.0 B	20.4 B	4.6 B	10.6 B	17
Cadmium	600	1,070		<11	<10.8	<t6.5< td=""><td><10.6</td><td><ા.ા</td><td>1.4</td></t6.5<>	<10.6	<ા.ા	1.4
Calcium	ND (1)	ND (1)		341,000	372000	30600	288000	190000	290000
Chromium, Total	430	160		<22.0	<21.6	<43.0	<21.2	<22.1	19
Cobalt	110,000	ND (1)		<22.0	⊲1.6	<33.0	<21.2	<22.1	1.4
Copper	ND (1)	ND (1)		<22.0	<21.6	109	<21.2	<22.1	160
Iron	ND (1)	ND (1)		1,050	1670	3980	1400	1220	5800
Lead	1,000	108		4.4	4.5	83.2	18.9	22.7	120
Magnesium	ND (I)	ND (1)		1,010 B	1080 B	1240	1110 B	970 B	1100
Manganese	5,500	ND (1)		45.6	24.1 B	63. I	21.7 B	31.9 B	230
Mercury	480	17		<0.11	<0.11	<0.15	<0.09	<0.10	0.4
Nickel	2,600	3.24		<44.0	<43.1	<66.0	<42.4	<44.2	300
Potassium	ND (1)	ND (1)		<2200	2160	<3300	<2120	<2220	110
Sodium	ND (1)	ND (1)		746 B	711 18	688 B	<212	<221	230
Vanadium	4,800	ND (1)		<22.0	<21.6	<33	<21.2	<22.1	9.9
Zinc	560,000	ND (1)*		<11.0	<10.8	154	26.0 B	42.7 B	300

< not detected at specified detection limit

NE - not established

NA - not analyzed

ND (1) - no data

- (2) not listed on the Soil Target Level Table
- but was listed in 62-775 of the FAC. Total VOC C Confirmed on second column

hurd in 62-775 as having a max conc. of • and 1 mg/kg for Total PAHs.

val Action Level as determined by BC T

. Jaceeds Action Level

Bold & Shaded - exceeds Action Level and

FDEP Soil Target Level

** - 2 samples used to calculate mean.

Data Qualifiers for Organic Compounds

J - Estimated value, «CRQL

P - >25 % difference in detected value between two columns

B - Compound detected in an associated blank

- Data Qualifiers for Inorganic Compounds
 - B Reading is less than CRQL but greater than IDL
 - E reported value is estimated due to interference
 - N spiked sample recovery not within control limits
 - · duplicate analysis not within control limits

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SUMMARY OF CONSTITUENTS DETECTED IN GROUNDWATER SITE WP 1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA 1993 AND 1996 INVESTIGATIONS

Homestead Air Reserve Base, Florida

Analyte	Florida Drinking Water Standard	EPA Drinking Water Standard	EPA Maximum Contaminant Level Goal	Sample ID Data Collected	MW 1993 SP1-I-902 3/1/93	MW 1993 SP1-I-902 3/1/93 DUPLICATE	MW 1993 SP1-EB-0001 3/1/93	OHM 1996 SP1-MW-001 1/25/96
VOA TCL Compounds (ug/l) Chloroform Bromodichloromethane 1,2-Dichloropropane Methyl Chloride Methyl Ethyl Ketone (2-Butanone	100 (a) (i) 100 (a) (i) 5 (i) 5 NS	100(a) 100(a) 5 NS NS	0 0 0 NS NS		2J 1J <10 <10 <10	2J 2J <10 <10 <10	<10 <10 2J <10 6J	<10 <10 <10 2J <10
BNA TCL Compounds (ug/l) Bis(2-Ethylhexyl)Phthalate Di-n-Butyl Phthalate 2-Methylnaphthalene Naphthalene	6 NS 100(b) 100(b)	6 NS NS NS	0 NS NS NS		0.3 J 0.5 J <11 <11	<11 0.5 J 2 J 1 J	<11 <11 <11 <11	<10 <10 <10 <10
Metals (ug/l) (c) Aluminum Arsenic Barium Calcium Chromium Cobalt Copper Iron Lead Magnesium Manganese Nickel Potassium Sodium Vanadium Zinc	200 (j) 50 (i) 2000 (i) NS 1000(j) 100 (i) NS 300 (j) 15 (i) NS 50 (j) 100 (i) NS 160000 (i) NS 5000 (j)	50-200 (h) 50 (g) 2000 (g) NS 1000 (h)/1300 (d) 100 (g) NS 300 (h) 15 (d) NS 50 (h) 100 NS NS S5000 (h)	NS NS 2000 NS 1000 100 NS NS 0 NS 100 NS NS NS NS NS		1850 18.4 36.9 B 209000 3.1 B <3.0 <2 1560 9.3 2780 B 26.7 <6 4410 B 14300 5.5 B 33.5	2610 18.1 37.5 B 220000 3.7 B <3.0 <2 1950 9.1 2980 B 28.5 <6 4950 B 14500 6.5 B 18.2 B	41.8 B <5.0 <1.0 95.3 B <2.0 <3.0 <2 <7.0 <3.0 <30.0 <1.0 <6 <325 <30.0 <3.0 27.1	<43.4 7.8 B 4.5 B 83900 <3.0 4.1 B <2.4 <3.8 <0.90 1490B <0.40 <4.4 642B 3870B <2.6 <3.9

Data Qualifiers for Organic Compounds J - estimated quantity, quality control criteria were not met

Data Qualifiers for Inorganic Compounds B - Reading is less than CRQL but greater than IDL

< not detected at specified detection limit Bold - equal to or greater than BG

NS - No Standard

Notes;

- MCL of 100 ug/L is for total THM's (a)
- total napthalenes must be <100 to meet FAC 62-770 guidlines (b)
- Metals listed are total metals (c)
- Federal Action Level- the action level is exceeded if the copper/lead in more than (d) 10% or targeted tap samples is greater than the action level.
- Numbers represent EPA's Proposed Primary MCL, Federal Register, Vol. 55. No. 143, July 1990. (f)
- Numbers represent EPA's Primary MCL for Inorganics. (g)
- (h) Numbers represent EPA's Secondary MCL for Inrganics which are non-efforocable, taste, odor or appearance guidlines
- (i) Florida Primary Drinking Water Standard.
 (ii) Florida Secondary Drinking Water Standard.

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SUMMARY OF CONSTITUENTS DETECTED IN GROUNDWATER SITE WP 1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA 1993 AND 1996 INVESTIGATIONS Homestead Air Reserve Base, Florida

Analyte	Florida Drinking Water Standard	EPA Drinking Water Standard	EPA Maximum Contaminant Level Goal	Sample ID Data Collected	OHM 1996 SP1-I-01 1/25/96	OHM 1996 SP1-I-02 1/25/96	OHM 1996 SP1-I-02 1/25/96 DUPLICATE	OHM 1996 SP1-I-03 1/24/96
VOA TCL Compounds (ug/l) Chloroform Bromodichloromethane 1,2-Dichloropropane Methyl Chloride Methyl Ethyl Ketone (2-Butanone	100 (a) (i) 100 (a) (i) 5 (i) 5 NS	100 (a) 100 (a) 5 NS NS	0 0 0 NS NS		<10 <10 <10 7BJ <10	<10 <10 <10 <10 <10	<10 <10 <10 1J <10	<10 <10 <10 <10 <10
BNA TCL Compounds (ug/l) Bis(2-Ethylhexyl)Phthalate Di-n-Butyl Phthalate 2-Methylnaphthalene Naphthalene	6 NS 100(b) 100(b)	6 NS NS NS	0 NS NS NS		<10 <10 <10 <10	<10 <10 <10 <10	<10 <10 <10 <10	<10 <10 <10 <10
Metals (ug/l) (c) Aluminum Arsenic Barium Calcium Chromium Cobalt Copper Iron Lead Magnesium Manganese Nickel Potassium Sodium Vanadium Zinc	200 (j) 50 (i) 2000 (i) NS 1000 (j) 100 (i) NS 300 (j) 15 (i) NS 50 (j) 100 (i) NS 160000 (i) NS 5000 (j)	50-200 (h) 50 (g) 2000 (g) NS 1000 (h)/1300 (d) 100 (g) NS 300 (h) 15 (d) NS 50 (h) 100 NS NS S50 (h)	NS NS 2000 NS 1000 100 NS NS 0 NS NS 100 NS NS NS NS NS		< 43.4 9.6B 7.6B 1140000 11.9 B <2.8 <2.4 11.4B <0.90 2740B 5.1B <4.4 3130B 10600 <2.6 6.7B	< 43.4 14.8 6.3B 99700 <3.0 4.4B <2.4 <3.8 <0.90 1970B <0.40 <4.4 <401 4880B 2.9B <3.9	<43.4 15.3 6.2 98900 <3.0 <2.8 <2.4 <3.8 <0.90 1950B <0.40 <4.4 838 4860 <2.6 <3.9	<43.4 3.9B 8.2B 10100 3.8B 6.2B 3.5B 4.9B <0.90 2210B 0.81B 8.0B 4030B 9310 4.5B 15.2

Data Qualifiers for Organic Compounds J - estimated quantity, quality control criteria were not met

Data Qualifiers for Inorganic Compounds B - Reading is less than CRQL but greater than IDL

< not detected at specified detection limit Bold - equal to or greater than BG NS - No Standard

Notes;

- (a) MCL of 100 ug/L is for total THM's
- total napthalenes must be <100 to meet FAC 62-770 guidlines (b)
- Metals listed are total metals (c)
- Federal Action Level- the action level is exceeded if the copper/lead in more than (c) 10% or targeted tap samples is greater than the action level. Numbers represent EPA's Proposed Primary MCL, Federal Register, Vol. 55. No. 143, July 1990.
- (f)
- Numbers represent EPA's Primary MCL for Inorganics. (g)
- Numbers represent EPA's Secondary MCL for Inrganics which are non-efforocable, taste, odor (h) or appearance guidelines.
- Florida Primary Drinking Water Standard. (i)
- Florida Secondary Drinking Water Standard. (j)

SUMMARY OF CONSTITUENTS DETECTED IN SEDIMENTS SITE WP-HOUG, ELECTROPLATING WASTE DISPOSAL AREA 1993, 1994, AND 1995 INVESTIGATIONS Homesicad Air Reserve Base, Florida

.

Analyte	Hackground	Sumple ID. Sumple Interval	MW 1993 SPL-SD-0007	MW 1993 SPI-SD-4HOR	IT 1994 EWA9 8-6"	17 1994 EWAL4 8-6"	IT 1994 EWA16 8-6"	1T 1994 EWA17 8-6"	IT 1994 EWAIE 84"	1T 1994 EWA19 8-4"
	}	Date Collected	3/12/93	3/12/93		1	-	-	-	-
VOA TCL Compounds (verka)	1						1	1		
Actions	NA		<1600 D	<12	NA	NA	NA	NA	NA	NA
Methylene Chionde	NA		<65	<12	NA	NA	NA	NA	NA	NA
	1					<u> </u>	t	t	<u>}</u>	
Particides/PCB TCL Compounds (ug/ug) Alpha-Chiordane	NE		847	120	NA	NA	NA	NA	NA	NA
Beta-Chirrinate	NE		92	120	NA	NA	NA	NA	NA	NA
Chierdane (sechnical)	NE		NA	NA	680 **	#2 **	210 **	180 **	330**	370 **
Endosalfan Selfanc	NE		ব্যয	<20	<11XD	<100	<100	<100	<200	<400
p.p DDD	NE		3477	<20	E307	6.6 3	4,4.1	4.21	210 C	19 J
p.#`DD€ p.#`DDT	NE NE		31.1P 59.1P	<20 <20	25 J 12 J	44C 46C	8.7 C 11 J	8.9 C 12 J	59 C 490 C	173
PCB-1260 (Aroutor 1260)	NE		<850	870	NA	NA	NA	NA	NA	NA
BNA TCL Compounds (ug/kg)									1	
Alexaphthese Alexaphthylese	<1400 NE	FXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXXX	73]	93 J <3900	<1 (600) <1 (600)	<11000 <11000	12003	<11000	<11000	<9800
Anthrauene	<140			2201	2006 J	<11000	4000 3	1700 J	12001	1500 J
Benzo(a)Anthracene	<14(X)		140.1	13(x) /	12000	1 0000	8500 3	5600 3	9788 3	13000.3
BenzikalPyrene	<14(1)		2003	1300 J	12000	2400 J	6800 3	4700 3	12000	1,2000
BenzuityFluuranchene	<14(3)		380.5	2300 5	13000	3000 3	6690 3	5004.3	14000	13000
Benzeigh.()Perviene	<1400 <1400		3103	3480.3	13000	<11000	<11880	<f1000< td=""><td>12000</td><td>1 90EF</td></f1000<>	12000	1 90EF
Bonzo(k)Flueranchene	<14(X) <14(X)		29013 17613	1300 J 71 J	17000	3000 J	\$100 J	4000 J	17000	16808
Benzyl Butyl Phthalaic Bis(2-Eihylhezyl) Phthalaic	<14(X)		1201	2301	<11080 1180 J	<11000 <11000	<11000	<11000 <11000	<11080	<9800
Carbezole	NE		<450	330 J	NA	NA	NA	NA	NA	NA
4-Chleronniline	NE		<450	<3900	<11000	<11000	<11000	<11000	<11000	<9800
Chrysenc	<1400		330.2	2200 J	13000	2706.2	Laure	6400 J	25006	17808
Di-n-Buiyl Philulate	NE		72 BJ	<3900	<11000	<11000	<11000	<11000	<11000	<9800
Dure-Octyl Phihalme	NE		<450	<3900	<11000	<11000	<11000	<11000	<11000	<9800
Dubenz(s.h)Amhraonne Dubenzofuran	<14(2) <14(3)		≪450 41	<3900 543	<11000	<11000	<11000	<11000 <11000	<11000	<9800
Diethylabhalate	NE		<450	<3900	<11000	<11000		<11000	<11000	<9400
Dunethyl Philalate	NE		<450	<3900	<11000	<11000	<11000	<11000	<11000	<9800
2.6-Destroiotuene	NE		<450	<39(1)	<11000	<11000	<11000	<11000	<11000	<9800
Flucturenthous	< (4(K)		430 J	3600 J	3,5000	5(00.3	1,9000	13000	21600	25966
Fluorene	<1400 <1400		81 230 J	110.1	<11000 13000	<11000	1600 1	<11000	<11000	<9300
Indeno(1.2.3-odiPyrene 2-Methylrophihalene	NE		31	<3900	<11000	<11000	<11000	<11000	<11000	0000
N-Nitrosodiphenytamine	NE		<450	<3900	<110XX	<11088)	<11000	<11000	<1100	<9800
Naphshalene	NE		<450	<3908	<11000	<11000	<11000	<11000	<11000	<9600
Phenanderone	<1400		1203	1500.3	20068	2409 J	17098	3540 J	6008 J	19009
Phenol	NE		<450	<3900	<11000	< 1(KX)	<11000	<11000	<11000	<9800
Pysche	<1408		550	3306 J	35668	4646]	57000	9908.3	23066	29090
Netais (mg/kg)									j i	1
Aluminum	2700		2770	350	2140	28,90	2070	4060	2380	2100
Anumosy Arsenic	NE 1		-3.0 N 7.1 N	3.0 BN 44 N	<129 16	<116 3.9	<120 \$51	<116 13.2	<133 16	<116 2.8
Arsenic Banvm	14		536+	52.9 -	2458	3.3 26.5	38.4	29.0	5290	3970
Beryllium	NE		NA	NA	<2.6	23	4	43	<2.7	<23
Cadmium	<2.1		4	1.7	<12.9	<11.6	<12.0	<11.6	<13.3	<11.6
Calcium	310,000		212000	337000	281000	336808	333000	329000	285000	225000
Chromium, Total Cobalt	(11		85.8 ° 3.3 B	15.8 * <0.40	47.8 <25.6	49.7	973 - 240	44.7	91.5 (36.6	76.1 <21.1
Copper	<6.2 16 J		3.3 B 52	-(1).40	<23.8 141	<23.2 <23.2	<24.0 <24.0	23.1	<36.6 60.3	-21.1 - 29.5
Logger Irres	1700		4210 2	352 E	9630	23/0	1830	2710	4468	6218
Lead	11		216 *	189 *	406	195	370	397	353	\$29
Magnesium	1000		1470	510 8	2548	1964	1296	1340	1696	1550
Manganese	<29		52.5	6.6	49.5	66.1	***	61.1	31.	55.2
Mercury , Nickel	0.043 NE		6.36 10.3	<n.10< td=""><td>82 <315</td><td><0.13</td><td>8.4 <48.0</td><td>0.30 <46.3</td><td>11 <313</td><td>8.34 <46.2</td></n.10<>	82 <315	<0.13	8.4 <48.0	0.30 <46.3	11 <313	8.34 <46.2
Printing	NE		560 8	<1.2 757 B	<2580	<46.4 <2320	<2400	<2310	<2660	<2310
Silver	NE		NA	NA	<12.9	13.8	<12.0	411.4	<13.3	<11.6
Sodium		000000000000000000000000000000000000000	269 8	341 B	346	638	407	507	361	24E
	2901	000000000000000000000000000000000000000								
Vanadsum Ziac	2901 5.71 271		19.R 612.E	2.4 B 35.9 E	-25.3	<23.2 162	<24.0 152	43.1 196	<26.6 787	<23.1 722

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 Data Qualifiers for Organic Compounds
 Decodecodecode

 J - estimated quantity, less than CRQL
 - compound descende is an associated blank.

 B - compound descende is an associated blank.
 D. Result is reported from secondary dilution

 P - >25% difference is descend value between two entimest
 E. Exceeded calibration range of instrument

 C - Confirmed on second origination
 * Elevated detection limit due to either pesticides present or matrix interferences

 ** - Altered patters
 **

Data Qualifiers Re laorganic Compounds B - Reading is less than CRQL but greater than IDL E - reported value is estimated due to inserference N - spited sample recovery not within control limits W - post digestion spike for furnace AA out of control limits * - duplicate analysis not within control limits

< not detected at specified detection limit NE - not established

NA - not analyzed

Bold > coust or greater than *BG Bold & Shaded > coust or greater than 2*BG

Page 1 of 1

SUMMARY OF CONSTITUENTS DETECTED IN SEDIMENTS SITE WP-IADU-5, ELECTROPLATING WASTE DISPOSAL AREA 1993, 1994, AND 1995 INVESTIGATIONS Homestend Air Reserve Hase, Florida

Hameslend Air Reserve Hase, Florida										
			17 1994	IT 1994	IT 1994	IT 1994	AT 1794	IT INH	11 1994	8T 1994
	Ruckground	Sample ID.	EWA20	EWA2I	EWA22	EWA23	EWA24	EWA25	EWA26	EWA27
Aastyle	•	Sample Interval	0-6"	4.6"	8-4"	8-6"	0-6"	0.6"		
		Date Collected				l				
VOA TCL Comprueds (surfig)						1	[
Actions	NA		NA	NA	NA	NA	NA	NA	NA	NA
Methylene Chiwade	NA		NA	NA	NA	NA	NA	NA	NA	NA
······										
Particides/PCB TCL Compounds (ug/kg)				ĺ			1	1	ļ	
Alphe-Chiordene	NE		NA	NA	NA	NA	NA	NA	NA	NA
Beta-Chioniane	NE		NA	NA	NA	NA	NA	NA	NA	NA
Chlordane (lechnical) Endosulfan Selfme	NE		<1000	270 ** <1000	270**	1400** <1000	340** <1000	620 ** < 1000	360++ <1000	\$4(I)** <1000
DDD DDD	NE		<200	<200	1.31	251	22.1	32.5	181	213
PD DDE	NE		INC	30	8,23	31.1	8.9J	16.1	9.4.1	152
TOOT	NE		<200	14.1	163	60.1	1101	791	41.5	291
PCB-1260 (Amulor 1260)	NE		NA	NA	NA	NA	NA	NA	NA	NA
SNA TCL Compounds (ug/lig)							1	1		1
Acesaphihene	<}4(X)		<9703	440 1	630 J	<46(X)	<4310	<2700)	<5800	<3400
Accessphilityleate	NE		<97(X)	6601	1500 J	6500	821) J	7900	11003	920)
Anthescene Benzola)Anthescene	<1400 <1400		<97(ii) 1 9000	1909 J 7200	3400 J 12800	12011	1709 J 5780	1400 J 6900	1600 J	1 100 J 3000 J
BCTZ(KB)ARITERCERE BCTZ(KB)ARITERCERE	<1400) <14003		19000	7,00	1,5004	4394 1	1340	9406	6968 11008	
Benzo(b)Flucenthese	<1400		23000	13000	21005	5400	12000	22000 E	18000	Mai
Benzoig h.i/Perylene	<1400		12000	VION	17066	3700 3	8300	9200	18086	6086
Benzo(k)Fluoramhene	<1400		23000	9900	19000	7008	9199	<1700	11000	7766
Benzyl Butyl Phihalane	<140)		<9700	<4200	<4900	<4600	<4300	<2700	<3100	<3400
Bis(2-Eihylhesyl) Phihalaic	<14(1)		3500 J	1100 J	2000 J	14004	<4300	1700 3	1700 3	-3400
Carbazoie	NE		NA	NA	NA	NA	NA	NA	NA	NA
4-Chierosailine	NE		<97031	<4200	<4910	<4600	<4310	<2700	<5800	<\$400
Chayache	<1400		29000	12060	21,008	7600	7600	14008	14000	3464
Den-Buiyi Phthalme	NE		<9700	<4200	<4900	<4600	<4300	<2700	<\$100	<5400
Di-n-Octyl Ptabalaie	NE		<97(1)	<4200	<4900	<4600	<4300	<2700	-3800	<5400
Dihena(&h)Ambracane	<1400	************	<9700	3300 J	6598	1500 J	<4300	<2710	් 300	<\$400
Dibestofuran	<1400		<9700	<4210	<490	<4600	<43(10	<27(1)	-5100	<\$400
Dicthyiphthalain	NE NE		<97(K)	<42(X) <42(X)	<4900 <4900	<4600 <4600	<4300 <4310	<2700 <2700	<5800 <5800	<\$400 <\$400
Dimethyl Phihalais 2.6-Diaistosolucas	NE		<9700	<4200	<49(1)	<4600	<4300	<2700	<5400	-5400
Paranticae	<1400		55000	20000	33000	13004	15000	1,9000	20000	12000
Flucence	<1400		<9700	5901	10007	<4600	490.2	3703	<5800	<3400
Indene(1.2.3-cd)Pyrene	<1403		14000	<4310	<4900	<4610	<43(X)	\$300	9196	\$306.3
2-Methylnaphthalene	NE		<9710	<421K)	<49(3)	<4600	<4300	<2700	<3800	<3400
N-Nitroendiphenylamuse	NE		<9700	<4200	<49(1)	<46(1)	<43(J)	<2700	<5800	<3400
Naphthaliene	NE		<9700	<42(R)	<4900	<4600	<4300	<2700	<5800	<\$400
Phonantheone	<1400	************	12000	7508	13000	4400 3	4900	5790	5900.3	3200 J
Pheaol	NE		<9700	<4210	<4900	<4600	<4 ¥Ю	<2700	<5800	<\$400
Рутелс	<1400		52090	15088	25908	10000	12000	17996	17000	11000
Meinis (main)										
Aluminum	2700	***************************************	1670	2050	2330	3480	2340	3548	3568	3130
Antimony	NE		<109	<128	<135	<137	<129	<149	<167	<157
Arsenic	2		2.8	51	41	65	53	9.7	12.0	9.4
Barum	14		3000	202	79.5	61.8	55.6	103	113	77.9
Beryllium	NE		41	6.0	7.0	<2.7	<2.6	<1.0	<33	<3.1
Cadmaum	বা		<10.9	<12.8	<13.5	<13.7	<12.9	<14.9	<16.7	<15.7
Calcium	310,000		242000	343080	303000	286000	344008	278000	290000	323608
Chronisten, Total	111		6 .4	54.5	72.8	169	52.0	116	116	96.9
Cobait	<4.2		<21.9	<25.6	<27.1	<27.4	<25.8	<29.1	-03.3	013
Copper	16 J 1700	***************************************	27.8 3150	27.6 2610	43.7 3070	43.4 3050	31.2 2240	51,1 4310	68.4 4440	41.2 3430
iron Lead	11	***************************************	3130	2640 341	_3076 296	.3050	354	4310	633	643
Magnesium	ana	***********	1334	1500	1340	1470	1350	icii	1639	1520
Manganese	<29		37.8	34.8	45.6	33.9	50.7	44	76.6	46.8
Mercury	0.043		4.21	0.16	634	111	4.34	8.41	8.48	8.27
Nickel	NE	***********	<43.7	<51.2	-54.1	-54.9	41.6	<59.5	<66.7	<62.7
Pritassium	NE	******************	<2190	<2560)	<2710	<2740	<2510	<2990	<3330	<1130
Silver	NE		<10.9	<12.8	<13.5	<13.7	<12.9	<14.9	<16.7	<15.7
Sodiest	290 J		444	841	599	614	707	<i>446</i>	683	578
		****************					~ ~ ~	<29.1		<31.3
Vanadium Zinc	5.7 x 27 x		<21.9 385	<25.6 331	<27.1 421	<27.4 329	-25.L 313	507. 500	<33.3 \$51	<31.5 666

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Data Qualifiers for Organic Compounds
 J - estimated quantity, fees than CRQL
 J - compound detected in an annotated Mank
 D- Result is reported from secondary dilution
 P - >21% difference is detected value between two columns
 E - Exceeded calibration maps of instrument
 C - Confirmed on second column
 · Elev Sted detection limit due to either perticides present or matrix interferences
 ** - Altered pattern

Data Qualifiers for languate Components B - Reading is less than CRQL but greater than IDL E - reported value is estimated due to insoference V - spiked sample recovery not within control limits W - poor digestion apile for furmers AA out of control limits * - deplicing analysis not within control limits

< was detected at specified detection limit

WR OCICIO H BOCHEL WORLAAM HANN
 NE - NOL BARDYCAI
 NA - act manipred
 Bold > expusion or greater that *BG
 Bold & Shaded > expusion greater that 1*BG

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SUMMARY OF CONSTITUENTS DETECTED IN SEDIMENTS SITE WP-1/OU-S. ELECTROPLATING WASTE DISPOSAL AREA 1993, 1994, AND 1995 INVESTIGATIONS Homestead Air Reserve State, Parida

	Homestend Ai	Reserve Huse, Florid	*		
	Buckground	Sample ID.	IT 1994 EWA28	17 1995 CS-02	1T 1995 CS-03
Analyte		Sample Interval Date Collected	ð-6"	(0-1)	(0-1)
	f	0000000000000		[[
VOA TCL Compounds (ug/kg) Autome	NA		NA	NA	NA
Methylene Chlonde	NA	K88888888888888	NA	NA	NA
Pesticides/PCB TCL Compounds (up/kg)		Recessessesses			
Alpha-Chiordane	NE		NA	NA	NA
Beta-Chlordane Chlordane (technical)	NE		NA 4(K) ==	NA 1400	NA <17
Endewalian Sullate	NE	888888888888888	<400	<3.3	<3.3
p.p.OOD	NE		7.7 1	33	<3.3
p.p/DDE	NE	6333333338888888	6.03	150	16
P.P.DDT	NE		13.1	24()	<3.3
PCB-1260 (Ameior 1260)	NE	8888888888888	NA	<33	<33
BNA TCL Compounds (ug/kg)					
Accessible ne	<14(8)	888888888888888	<24(3)	<330	<493
Aucnaphthylene	NE		<24(X)	<330	<493
Anthracear	<1400	888888888888888888888888888888888888888	190.1	<330	<493
Senzo(#)Anthracene	<1400	8888888888888888	2200 J	3190	480
Benzo(2)Pyrene	<14(3)		2780	2400	370
Benzo(b)Fivoranthene	<14(K)	838888888888888	4100 2200 J	3600	690 <493
Benzo(g.h.i)Peryiene Benzo(k)Fluccianthene	<14(8)	833333333333333	2200 1	2100	<493
Benzyl Butyl Phthalaic	<1400		<24(X)	<330	<443
Bis(2-Ethylhexyl) Phthalate	<1400		<24(1)	<330	<493
Carbazole	NE	8888888888888888888888	NA	NA	NA
4-Chlomeslline	NE		<24(3)	<330	<493
Chrysene	<1400	832222222222222222222222222222222222222	3609	<330	<493
Di-s-Butyl Phthalme	NE	8388888888888888888	<2400	<330	410
Di-s-Octyl Phthalate	NE		<2400	<330 <330	<493 <493
Dibenz(a.h)Anthracene Dibenzofursa	<1400 <1400	8888888888888888888888	<2400	<330	<493
Dicthyighthalac	NE	188888888888888	<24(1)	<330	<493
Dimothyl Phthalate	NE		<2400	<330	<493
2.6-Dirutrolucne	NE	888888888888888	<24(K)	<330	<493
Fluoranthese	<1400	833333333333333	3780	5694	830
Fluorene	<14(X)		<2400	<330	<493
Indeno(1.2.3-cd)Pyrene	<1400	18888888888888	2000 J	1700	<493
2-Methyinaphthalene	NE		<2400 <2400	<330 <330	<493 <493
N-Nitronodiphenytamine Nachthalens	ME	K888888888888888	<24(8)	<330	<493
Phenanthrene	<14(X)		2100 J	<330	530
Phenol	NE	188888888888888	<2400	<330	<493
Pyrene	<1490		2000	3708	530
Metais (mg/kg) Aluminum	27(X)	888888888888888888888888888888888888888	3550	4040	3708
Antimony	NE		<131	- 3	-3
Arsenic	2	888888888888888888888888888888888888888	3.6	25	22
Barium	14		42.5	25	36
Beryllium	NE		<2.6	<0.5	-60.5
Cadmium	<2.1 310,000	83333333333333333	<13.1 .318000	2.7 290000	3.5
Chicium Chromium, Totai	310,000		318000 71.9	290000 13	310000
Cristing a dias	<4.2	888888888888888	<26.1		<1
Copper	16.5		43.8	2#	14
irca	1700	888888888888888888888888888888888888888	3836	3100	1308
Lesd	I II I	88888888888888	157	230	100
Magnessur	HXK		1130	1100	1100
Manganese	<29	888888888888888	32.1	67	44 91.0
Morcary Nickel	0.043 . NE		<0.13	1.9 6.2	 _<[
Potassium	NE	8888888888888888888	<2610	110	160
Silver	ME		<13.1	L>	<1
Sociam	2903		750	560	500
Vannium	5.73	88888888888888888888888888888888888888	<26.t	10	7.8
Zinc	27 J		239	338	718

Data Qualifiers for Organic Compounds J - exumered quantity, less than CRQL B - compound desorted in an associated blank

a composing detected is an associated blank.
 D-Result is reported from accordancy dilution.
 > >25% difference in detected value betware two columns.
 E - Exceeded calibration range of instrument.
 C - Confirmed on second column.
 * - Elevated detection limit due to either pesticides present or matrix interferences.
 * - Altered pattern.

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Data Qualifiers for Incrganic Compounds B - Reading is less than CRQL but greater than IDL E - reported value is estimated due to interference N - spitcel sample recovery nor within control limits W - past digestion spike for furnace AA out of control limits * - duplicate analysis not within control limits

< not detected at specified detection limit

NE - ext established NA - not enalyzed Bold > equal or greater than *BG Bold & Shaded > equal or greater than 2*BG

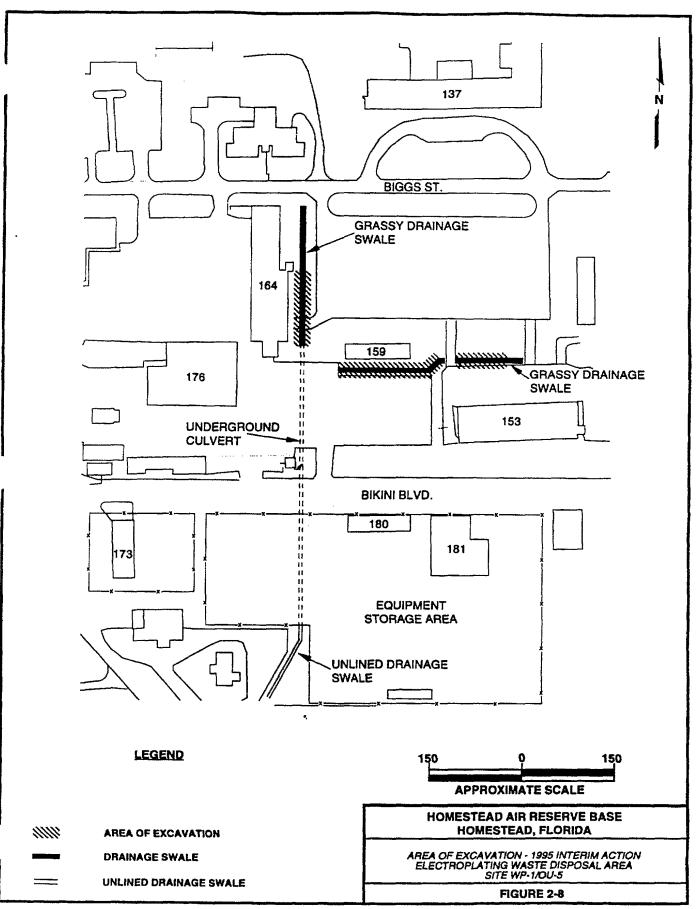
in this area and lack of sediment observed at these locations. The remaining samples were considered sediments based on their location ina well defined drainage ditch or swale and the presence of sediment. The 28 soil/sediment samples were analyzed for base neutral and acid extractable compounds (BNA's) and TAL metals. Additionally, the 21 soil/sediment samples from the drainage swales were further analyzed for pesticides. All analyses were performed in accordance with USEPA SW-846 protocols. Analytical results of the 1994 sampling are discussed in Sections 2.6.2 and 2.6.4. Tables 2-9 and 2-11 provide a summary of analytical results for the 1994 investigation.

In 1995, IT Corporation completed an Interim Action which included the excavation of soil/sediment from selected areas within the northern swale. This work was performed under contract to USACE - Mobile District. The excavations were completed to a depth of 1 ft bls and extended three ft to either side of the centerline of the swale (Figure 2-8). The excavations were completed using a front loader/backhoe combination unit. Excavated soils/sediments were stockpiled on visqueen, bermed with hay bails and covered at the end of each day's work shift. For the purposes of evaluation, the confirmation sample CS-01 was evaluated as representative of soil/weathered rock while, CS-02 and CS-03 were evaluated as sediments.

Soil/sediment excavation activities were not completed in the south area. Upon completion of the excavation, confirmation samples were collected from each of the thre**e** orth area excavations (Figure 2-6). The confirmation samples were analyzed for BNAs, TAL metals, and pesticides. Analytical results of the 1995 sampling are discussed in Sections 2.6.2 and 2.6.4. Tables 2-9 and 2-11 provide a summary of analytical results for the 1995 investigation.

2.6.1.4 1996 Confirmation Groundwater Sampling. On January 24 and 25, 1996, OHM Corporation conducted acomplete round of groundwater sampling from the four groundwater wells located on OU-5/Site WP-1. This work was performed under contract to AFCEE. This groundwater sampling event was recommended by the USEPA to provide current groundwater quality information for site characterization purposes. OHM Corporation completed the groundwater sampling program in accordance with USACE internal statement of work dated August 22, 1995.

In accordance with the scope of work, OHM Corporation collected groundwater samples from monitoring wells MW-0001, I-01, I-02, and I-03. Groundwater sampling locations are shown on Figure 2-7. Two equipment blanks, one blind duplicate, and two trip blanks were



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also collected and analyzed as part of the sampling event. Groundwatesamples were analyzed for TCL VOCs, TCL BNAs, TCL organcohlorine pesticides/PCBs, TAL metals, and cyanide. Laboratory analyses were performed by Analytical Technologies, Inc., located in Pensacola, Florida. Analytical results of the 1996 sampling are discussed in Section 2.6.3. Table 2-10 provides a summary of analytical results for the 1996 investigation.

A summary of the detected parameters from the 1991, 1993, 1994, 1995, and 1996 investigations are provided in Sections 2.6.2, 2.6.3, 2.6.4, and 2.6.5.

2.6.2 Soil Investigations

This section presents the results from the soil samples that were collected by G&M during the 1991 Remedial Investigation, results from the 1993 Montgomery Watson Remedial Investigation Addendum, and the 1994 and 1995 IT Corporation Interim Action Investigations. Results of soil analyses for the previous and the current investigations are presented by analytical group (i.e., VOCs, metals, etc.). Figures 2-3 and 2-4 provides an illustration of the locations of the soil sampling points for 1991 and1993, Figures 2-5 and 26 present the soil sampling points for the 1994 and 1995 investigations, respectively.

2.6.2.1 Volatile Organic Compounds. 1991 Investigation. A summary of laboratory results of constituents detected in soil during G&M's 1991 investigation are presented in Table 2-5. Acetone was the sole volatile organic compound (VOC) detected in 1991 and was seen in two of four samples, including the background sample (SP1-SL-0002). Concentrations ranged from 1,268 μ g/kg to 7,301 μ g/kg. G&M calculated the verage background to be 1,029 μ g/kg for Homestead ARB at the 2-4 ft bls level. Based on current sampling information, and the backgroundsoil data, the acetone appears to be related to the degradation of isopropanol used in the decontamination of field sampling equipment.

1993 Investigation. In 1993, only soil sample SP7-SL-0007 and its duplicate, SP7-SL-9007, were analyzed for VOCs as required by the Work Plan. Table 2-9 presents concentrations of compounds detected in soils during the, 1993 investigation. Acetone was detected at 25,000 and 27,000 μ g/kg, respectively. These concentrations are well below the State of Florida Health-Based Soil Target Levels. The acetone detected in these samples is believed to be attributable to the degradation of the isopropanol utilized for field decontamination of sampling equipment. Isopropanolsamples were analyzed and found to contain acetone at concentrations up to 120,000 μ g/L. Acetone, therefore, is most likely a field contaminant introduced into the samples during the decontamination process. A discussion of the

isopropanol analysis is included in the Quality Control Summary Report (QCSR) submitted to the USACE-Omaha District under separate cover.

1994 and 1995 Interim Action. Confirmation samples collected during the 1994 and 1995 Interim Action investigations were not analyzed for VOCs.

2.6.2.2 Base Neutral/Acid Extractable Compounds. 1991 Investigation. During the 1991 investigation, five soil samples were collected and analyzed for Base Neutral/Acid Extractable Compounds (BNAs). These samples were identified as SPI-SL-0001, SPI-SL0002-split, SP1-SL-9002 (the duplicate of SP1-SL-0002), SP1-SL-0003, and SP1-MW-0001S. Soil boring SP1-SL-0002 was identified prior to sampling as a background sampling location. All soil samples for the event were collected from the 2-4 ft bls interval. Detected BNAs from the 1991 sampling event are listed in Table 2-5.

Several BNA compounds (mainly polynuclear aromatic hydrocarbons [PAHs]) were detected in background sample SP1-SL-0002-split, although, at very low levels. Concentrations of PAHs detected in the background sample, including benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(g,h,i)perylene, benzo(k)fluoranthene, chrysene, fluoranthene, phenanthrene, and pyrene, are estimated values because they are between the method detection limit and the practical quantitation limit (PQL). Additionally, two non-PAH BNAs, bis(2-ethylhexyl)phthalate and 2-chlorophenol, were detected in sample SP1-SL0002-split at estimated concentrations of 110 and 10 μ g/kg, respectively. PAHs were not detected in sample SP1-SL-9002, or the three other OU-5/Site WP-1 soil samples.

The concentrations of PAHs, bis(2-ethylhexyl)phthalate and 2-chlorophenol detected in OU5/Site WP-1 background samples were equal to or less than the average Homestead ARB background concentrations for the 2-4 ft depth interval. The PAH, bis(2-ethylhexyl)phthalate, and 2-chlorophenol concentrations are similar because the background data collected at OU-5/Site WP-1 was included in the average Homestead ARB calculations and some PAH compounds and 2-chlorophenol were only detected in background data collected at OU-5/Site WP-1. The PAH, and bis(2-ethylhexyl)phthalate concentrations detected at OU-5/Site WP-1 were less than the average Homestead ARB background concentrations for the 0-2 ft depth interval. PAHs and 2-chlorophenol were not detected in the Homestead ARB background samples included in the 4-6 ft depth interval. Additionally, the concentrations of bis(2-ethylhexyl)phthalate detected at OU-5/Site WP-1 were below the average Homestead ARB concentration of this compound calculated for the 4-6 ft interval.

1993 Investigation. Of the four samples collected in 1993, only one sample (SP1-SL-0007) and its duplicate (SP1-SL-9007) were analyzed for BNAs (Table 2-9). This sample was collected from the 0-1 ft bls interval. Fifteen PAH compounds were detected at comparable levels in both samples. Additionally, dibenzofuran, carbazole and four phthalates were detected. Carbazole concentrations were 35 and 62 μ g/kg and are estimated below the contract required quantitation limit (CRQL). Di-n-butyl phthalate and bis-2-ethylhexyl phthalate were detected at less than the CRQL and were also detected in the laboratory blank. Benzylbutyl phthalate and di-n-octyl phthalate at 24 µg/kg and 7 µg/kg, respectively, were detected in SP1-SL-0007. SP1-SL-9007 contained 14 µg/kg of benzylbutyl phthalate. All phthalate concentrations were below the CRQL. The total phthalate concentration of 141 μ g/kg is slightly greater than the average Homestead background of 126 µg/kg. The total PAH concentrations for SP1-SL-0007 and SP1-SL-9007 were 3,609 µg/kg and 4,825 µg/kg, respectively. However, only five of the PAH compounds were above the CRQL: fluoranthene (650 and 880 μ g/kg), pyrene (510 and 810 μ g/kg), benzo(b) fluoranthene (380 and 540 µg/kg), benzo(k)fluoranthene (390 and 420 µg/kg), and chrysene (410 and 540 µg/kg). All other individual PAH compounds were detected at less than the CRQL. None of the detected BNA compounds exceed the State of Florida Health-Based Soil Target Levels.

The background soil samples for OU-5/Site WP-1 are samples SP1-SL-0002 and its duplicate, SP 1-SL-9002, collected by Geraghty & Miller in 1991. Soil sample SP1-SL-0007 (0-1 ft bls), collected during the 1993 investigation, indicates an order of magnitude greater concentration of BNAs when compared to the 1991 background sample SP1-SL-0002 (2-4 ft bls). Background total PAHs at Homestead ARB as presented by G&M for the 0-2 ft bls interval were 739 μ glkg. In 1993, Total PAHs were 3,609 and 4,825 μ g/kg detected in samples SP1-SL-0007, and SP1-SL-9007, respectively.

A comparison of the 1991 sample, SP1-MW-0001-S, with the 1993 sample SP1-SL-0007 indicate an absence of detected PAHs above the method detection limit in 1991, while 17 PAH compounds were detected in 1993. However, with the exception of pyrene ($<361 \mu g/kg$ in 1991 and 510 $\mu g/kg$ in 1993), the reported quantitation limits for the PAHs analyzed in 1991 were above the quantities reported in 1993. PAH compounds are not generally associated with electroplating waste. Sample SP1-SL-0007 is located in close proximity to the asphalt parking lot. These PAH results may be indicative of run-off from the parking area. The PAH concentrations reported for OU-5/Site WP-1 were near the values reported for urban areas and are within the range of values reported for road dust (Menzie, et al., 1992).

The background soil samples for OU-5/Site WP-1 are samples SP1-SL-0002 and its duplicate, SP1-SL-9002, collected by Geraghty & Miller in 1991. Soil sample SP1-SL-0007 (0-1 ft bls), collected during the 1993 investigation, indicates an order of magnitude greater concentration of BNAs when compared to the 1991 background sample SP1-SL-0002 (2-4 ft bls). Background total PAHs at Homestead ARB as presented by G&M for the 0-2 ft bls interval were 739 μ g/kg. In 1993, Total PAHs were 3,609 and 4,825 μ g/kg detected in samples SP1-SL-0007, and SP1-SL-9007, respectively.

A comparison of the 1991 sample, SP1-MW-0001-S, with the 1993 sample SP1-SL-0007 indicate an absence of detected PAHs above the method detection limit in 1991, while 17 PAH compounds were detected in 1993. However, with the exception of pyrene (<361 μ g/kg in 1991 and 510 μ g/kg in 1993), the reported quantitation limits for the PAHs analyzed in 1991 were above the quantities reported in 1993. PAH compounds are not generally associated with electroplating waste. Sample SP1-SL-0007 is located in close proximity to the asphalt parking lot. These PAH results may be indicative of run-off from the parking area. The PAH concentrations reported for OU-5/Site WP-1 were near the values reported for urban areas and are within the range of values reported for road dust (Menzie, et al., 1992).

Dibenzofuran was detected in SP1-SL-0007 and the duplicate at 4 and 11 μ g/kg, respectively. Carbazole was also detected at 35 and 62 μ g/kg. All values reported for dibenzofuran and carbazole are less than the CRQL. Carbazole was not analyzed in the 0-2 ft Homestead ARB background samples and was not detected at a quantitative limit of 1,250 μ g/kg in the 2-4 ft background sample. Dibenzofuran was not detected in the background samples for the Base or OU-5/Site WP-1.

1994 and 1995 Interim Action. During the 1994 confirmation sampling program, 7 of the 28 samples collected were considered soil/weathered rock due to their position either underlying the asphalt parking area (EWA-1 and EWA-2) or within the grassy swale east of Building 164 (EWA-3 through EWA-7). Two of the soil samples EWA-6 and EWA-7, are no longer considered representative of site conditions, given that the area from which they were sampled, was excavated during the 1995 excavation and removal activity. Analytical results of the 5 remaining samples did not indicate the presence of BNA compounds above the specified detection limit. The 1994 BNA soil analytical results for these 5 samples have been summarized and are presented in Table 2-9. However, elevated detection limits were reported in each of the samples. Ten BNAs were reported in the two soil samples (EWA-6 and EWA-7) collected from areas that were subsequently excavated. Five of the BNAs

detected exceeded the State of Florida Health-Based Soil Target Levels in sample EWA-7. The BNAs exceeded include; benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, fluoranthene, and indeno (1, 2, 3-C, D)pyrene. Concentrations of these compounds ranged from $11,000 \mu g/kg$ to $32,000 \mu g/kg$.

Soil analytical results from the 1995 confirmation sample (CS-01) collected from the base of the excavation associated with the 1994 soil samples EWA-6 and EWA-7, indicate detectable concentrations of 10 BNA compounds, primarily PAHs. The compounds detected include 9 of the compounds found in the pre-excavation sample, di-n-butyl phthalate was the only compound not previously reported. The concentration of detected BNA compounds ranged from $370 \,\mu g/kg$ to 1260 $\mu g/kg$. None of the compounds detected were reported above the State of Florida Health-Based Soil Target Levels. However, the concentration of PAHs in this sample exceed the FDEP 62-775 maximum level of 1000 $\mu g/kg$ Total PAHs.

A summary of the 1995 soil BNA analytical results is provided in Table 2-9.

2.6.2.3 Organochlorine Pesticides/PCBs. 1991 Investigation. Soil samples collected by G&M during the 1991 investigation were not analyzed for Pesticides and PCBs.

1993 Investigation. Four soil samples plus one duplicate were collected and analyzed for organochlorine pesticides and PCBs. A summary of the pesticide/PCB constituents detected is provided in Table 2-9. No PCBs were detected in any of the soils collected. Concentrations of pesticides detected in 1993 soil samples are well below the CALs as well as the State of Florida Health-Based Soil Target Levels. Sample SP1-SL-0004 at the 0-2 ft bls interval showed no detectable concentrations of pesticides. DDT and its metabolites were detected in the three soil samples SP1-SL-0006, SP1-SL-0007, and the background sample, SP1-SL-0005. DDT ranged in concentration from 0.83 (background SP1-SL-0005) to 34 μ g/kg. The DDE metabolite was seen at 3.8 to 130 μ g/kg (SP1-SL-0006). The background sample was reported at 3.9 μ g/kg DDE. The DDD metabolite was not observed in SP1-SL-0005, but was detected in samples SP1-SL-0007 and the SP1-SL-0007 duplicate at concentrations ranging from 0.75 to 28 μ g/kg. Endosulfan sulfate was detected in two locations: SP1-SL-0006 (0.77 μ g/kg) and SP1-SL-0007 (6.4 and 8.8 μ g/kg). Alpha and beta chlordane were detected in SP1-SL-0006 at 1.9 and 2.3 μ g/kg, respectively.

soil samples were collected adjacent to the grassy swale located east of Building 164 (Figure 2-2) where runoff from the surrounding area collects. Analytical data for the 1984 sampling is presented in Table 2-2. Low concentrations of cyanide and all metals analyzed, except cadmium, were detected in the four samples as follows: cyanide concentrations ranged from 0.60-1.3 mg/kg, total chromium concentrations ranged from 0.02-0.07 mg/kg; copper concentrations ranged from 0.08-0.11 mg/kg; lead concentrations ranged from 0.09-0.21 mg/kg; nickel concentrations ranged from 0.01-0.05 mg/kg; and zinc concentrations ranged from 0. 12-0.72 mg/kg. Concentrations of chromium, copper, lead, nickel, and zinc detected in these surficial samples are below concentrations detected in background soil sample SP1-SL-0002, collected northwest of Building 164 during the 1991 investigation. In addition, concentrations of chromium, copper, lead, nickel, and zinc were well below the average Homestead ARB background concentrations (7.6, 1.8, 3.3, 1.1, and 3.3 mg/kg, respectively) for the 2-4 ft bls soil interval and the average carbonate composition concentrations reported by Hem (1989).

1991 Investigation. TAL metals detected in the four soil samples collected from 2-4 feet bls in 1991 included aluminum, arsenic, barium, beryllium, calcium, chromium, cobalt, copper, iron, lead, magnesium, manganese, mercury, nickel, potassium, sodium, vanadium zinc, and lead (Table 2-5). These constituents are typically present in carbonate rocks and soils. According to the average carbonate-composition data presented by Hem (1989), calcium, magnesium, aluminum, iron, potassium, manganese, and sodium are the most common constituents of carbonates. Arsenic, barium, beryllium, chromium, cobalt, copper, lead, nickel, vanadium, and zinc occur at trace levels. Concentrations of most of the metals detected in the background samples for OU-5/Site WP-1, SP1-SL-9002 and SP1-SL-0002-split, were below the average carbonate composition concentrations with the exception of calcium, chromium, and sodium. Similarly, most metals detected in the three soil samples collected at OU-5/Site WP-1, SP1-SL-0001, SP1-SL-0003, and SP1-MW-0001-S, were below the average limestone composition concentrations except for concentrations of calcium and chromium. Additionally, concentrations of aluminum, beryllium, and vanadium in sample SP1-SL-0003 were above the average carbonate concentrations. Cyanide, which is not a common constituent of limestone, was not detected in the soil samples collected in 1991.

Concentrations of chromium (7.2 to 23.4 mg/kg), copper (1.7 to 2.6 mg/kg), lead (0.64 to 5.5 mg/kg), nickel (0.87 to 5.2 mg/kg), and zinc (2.2 to 2.9 mg/kg) detected in the soil samples collected in 1991 were higher than concentrations detected in 1984 except for cyanide which was not detected in the 1991 samples. The different sampling intervals used during these

investigations may explain the difference in detected concentrations of these metals: the samples collected in 1984 were surficial samples and the samples collected in 1991 were bedrock samples of Miami Oö1ite collected from 2 to 4 ft b1s.

Concentrations of TAL metals detected at OU-5/Site WP-1 were compared to background concentrations, which were determined by combining results from background soil samples SP1-SL-9002-split and SP1-SL-0002. Concentrations of calcium, magnesium, and vanadium detected in samples SP1-SL-0001, SP1-SL-0003, and SP1-MW-0001S were above the background concentrations of these metals. In addition, concentrations of aluminum, copper, iron, manganese, nickel, potassium, and arsenic detected in samples SP1-SL-0003 and SP1-MW-0001-S were greater than background concentrations. Sample SP1-SL-0003 also contained concentrations of barium, beryllium, chromium, cobalt, and lead that were greater than background concentrations.

Concentrations of TAL metals detected at OU-5/Site WP- 1 were also compared to average Homestead ARB background concentrations for the 2-4 ft bls depth interval. Concentrations of calcium, magnesium, and vanadium detected in samples SP1-SL-0001, SP1-SL-0003, and SP1-MW-0001S were above the average Homestead ARB background concentrations of these metals in the 2-4 ft bls depth interval. Concentrations of aluminum, arsenic, chromium, copper, iron, nickel, and potassium detected in samples SP1-SL-0003 and SP1-MW-0001-S were greater than the average Homestead ARB background concentrations for the 2-4 ft bls depth interval. Additionally, concentrations of barium, lead, and manganese detected in sample SP1-SL-0003 and sodium detected in sample SP1-SL-0001 were also above the average Homestead ARB background concentrations for the 2-4 ft bls depth interval.

1993 Investigation. Four soil samples were collected in 1993 and analyzed for cyanide: SP1-SL-0004 (0-2 ft b1s), SP1-SL-0005 (0-2 ft b1s), SP1-SL-0006 (0-2 ft b1s), and SP1-SL0007 (0-1 ft b1s), and a duplicate, SP1-SL-9007 (0-1 ft b1s). Cyanide was not detected above the CRQLs which ranged from 0.28 to 0.29 mg/kg dry weight.

Only sample SP1-SL-0007 and its duplicate, SP1-SL-9007, were analyzed for TAL metals. Aluminum, arsenic, barium, cadmium (duplicate only), calcium, chromium, cobalt, copper, iron, lead, magnesium, manganese, nickel, potassium, sodium, vanadium, and zinc were detected at this location (Table 2-9). With the exception of cadmium in the duplicate, and cobalt, these same metals were detected during the 1991 sampling. Calcium (285,000 mg/kg), magnesium, (1,160 mg/kg), and sodium, (513 mg/kg) were below the OU-5/Site WP-1 background sample collected by G&M in 1991 (SP1-SL-0002). Barium, (17.7 mg/kg)

and cobalt (0.62) were below the Homestead ARB average of 1.2 mg/kg. The remaining metals were detected at concentrations slightly above site background and also above Homestead ARB background. Aluminum (5,460 mg/kg), iron (3,070 mg/kg), manganese (72.4 mg/kg), nickel (13 mg/kg), potassium (630 mg/kg), and vanadium (8.4 mg/kg) were below average carbonate composite values (HEM, 1989).

Arsenic is the only metal detected which exceeds the State of Florida Health-Based Soil Target Level of 3 and was reported at 9.1 mg/kg and 9.7 mg/kg in the duplicate. However, arsenic is a common constituent in the environment and is present throughout Homestead ARB. A CAL of 10 mg/kg was established for the base by USEPA, FDEP, and DERM. While the arsenic levels approached the CAL of 10 mg/kg, it was not exceeded. Nickel was also detected in soil sample SP1-SL-0007 and its duplicate SP1-SL-9007 at concentrations of 7.2 mg/kg and 6.1 mg/kg, respectively. These nickel concentrations exceed the CAL of 3.24 mg/kg but are well below the State of Florida Health-Based Soil Target Level of 2,600 mg/kg. Furthermore, the nickel concentrations are below the average carbonate concentration and the common concentration of nickel found in the Eastern United States.

Chromium was detected at 18.6 and 18.9 mg/kg. Copper (13.8 and 15.5 mg/kg), lead (38.3 and 44.5 mg/kg), and zinc (20.1 and 23.3 mg/kg estimated due to interference) were also detected. Arsenic, copper, lead, and zfnc levels were slightly higher in the 1993 soil samples than those reported by G&M in 1992.

The furnace metals, arsenic, lead, selenium, and thallium, are qualified as estimated due to inherent interference from the limestone nature of the soil. In general, the metals concentration values for aluminum, beryllium, cadmium, calcium, cobalt, copper, magnesium, manganese, mercury, nickel, potassium, sodium, and vanadium are unqualified data. Data for antimony, barium, chromium, iron, lead, silver, and zinc are qualified as estimated due to interference or difficulty with reproducibility, again caused by the nature of the samples. Further discussion of these interferences is addressed in the associated QCSR.

As reported by G&M in 1991, the metals constituents observed are typically present in carbonate soils and rocks. With the exception of arsenic (9.1 mg/kg), calcium (285,000 mg/kg), chromium (18.6 mg/kg), cobalt (0.62 mg/kg), copper (13.8 mg/kg), lead (38.3 mg/kg), sodium (513 mg/kg), and zinc (20.1 mg/kg), the metals were below the carbonate soils averages.

1994 and 1995 Interim Action. Soil samples collected during the 1994 and 1995 Interim Action were each analyzed for TAL metals. Cyanide was not analyzed for during these investigations. Results from the 1994 and 1995 investigations have been summarized and are presented in Table 2-9. With the exception of cobalt, copper, lead, mercury, and nickel, metal results are comparable to previous sampling results. The concentrations of these metals were typically higher than 1993 and 1991 results. Cobalt was not detected in any of the 1994 samples, but was detected in the 1995 soil samples, CS-01 at a concentration of 1.4 mg/kg. The maximum concentrations of copper were detected in the 1994 soil samples EWA-3 at a concentration of 109 mg/kg and the 1995 soil sample CS-01 at a concentration of 160 mg/kg. Elevated detection limits were reported for cobalt and copper in the 1994 soil samples. Lead was detected in each of the 1994 and 1995 samples ranging in concentration from 4.4 mg/kg to 799 mg/kg. The maximum lead concentration was reported in sample EWA-7. This sampling point has been subsequently excavated. Mercury and nickel were only detected in the 1995 soil samples at concentrations of 0.40 mg/kg and 300 mg/kg. The levels of cobalt, copper, lead, mercury, nickel and silver appear to have higher concentrations in the excavated portions of the North Area swale, i.e., at sampling points EWA-6 and EWA-7.

The concentrations of metals in the 1994 and 1995 soil samples are below the State of Florida Health-Based Soil Target Levels with the exception of arsenic. Arsenic, lead, and nickel concentrations exceed the CALs of 10 mg/kg, 108 mg/kg and 3.24 mg/kg in the 1994 soil sample EWA-7 (subsequently excavated) and the 1995 confirmation sample CS-01.

2.6.2.5 Summary Section for Soils. In general, analytical results do not indicate significant impact in the areas of soil sample collection. Concentrations of compounds detected in soils at OU-5/Site WP-1 include BNAs, pesticides, and metals. The VOC detected (acetone) has been traced to field decontamination of sampling equipment. Acetone has been widely detected in soil samples throughout Homestead ARB and has been identified as an artifact of the degradation of isopropyl alcohol used during field decontamination procedures. Field samples of the isopropanol alcohol were obtained for analysis which indicated up to 120,000 μ g/L acetone content. Results of the acetone sampling are documented in the quality control summary report (QCSR) provided to the USACE, Omaha District.

BNAs, primarily PAHs, were detected in the one sample collected in 1993, the background sample collected in 1991, and the confirmation samples collected in 1995. BNA compounds were not detected in the 1994 samples due to elevated detection limits. The proximity of

sample SP1-SL-0007 to the asphalt parking lot and the surficial location of this sample indicate the potential source of the PAH compounds may be run-off from the parking lot. The PAH compounds seen in the 1991 background samples were reported at higher concentrations than those detected in the 1993 samples. Only the background sample collected in 1991 had detectable quantities of PAHs with all results estimated at values less than the PQLs. The PAH concentrations were near the range of background concentrations reported for urban areas and those affected by anthropogenic influences.

Cyanide has not been detected in any of the soil samples collected at OU-5/Site WP-1. A comparison of the 1993, 1994, and 1995 metals analytical results indicate elevated concentrations above State of Florida Health-Based Soil Target Levels or CALs for arsenic, lead, and nickel in the 1995 excavation sample. Again, this may be indicative of their location within the swales which receive runoff from the surrounding area. Nickel was above the CALs in both 1993 and 1995 soil samples, while lead was above CAL in only the 1995 soil sample collected from the swale excavation. Arsenic exceeded the State of Florida Health-Based Soil Target Levels in the 1993 and 1995 soil samples.

2.6.3 Groundwater Investigations

The Electroplating Waste Disposal Area was identified during the initial IRP Phase I investigations. Groundwater samples have been collected at OU-5/Site WP-1 in every phase of field investigations conducted at Homestead ARB since 1984 with the exception of the 1994 and 1995 Interim Action.

During the 1984 IRP Phase II investigation, three monitoring wells (I-01 to I-03) were installed (Figure 2-3) and groundwater samples were collected and analyzed for cyanide and total metals including cadmium, total chromium, hexavalent chromium, copper, lead, nickel, and zinc. A summary of the analytical data generated in 1984 is presented in Table 2-1. Low concentrations of cadmium (<0.2 to $0.4 \mu g/L$), total chromium (<0.5 to $19.7 \mu g/L$), hexavalent chromium (<0.1 to $1.7 \mu g/L$), copper (4.2 to $7.0 \mu g/L$), lead (5.7 to $9.0 \mu g/L$), nickel (9.2 to $16.9 \mu g/L$), and zinc (15.1 to $16.3 \mu g/L$) were detected in the three groundwater samples. Cyanide was not detected in the groundwater samples. The highest concentration of total chromium was detected in sample I-01; the highest concentration of lead was detected in the duplicate of sample I-02; the highest concentrations of zinc was detected in samples I-02 and I-03. The concentrations of these constituents detected at OU-5/Site WP-1 were well below their applicable Florida Primary and Secondary

Drinking Water Standards and the Federal Maximum Contaminant Levels (MCLs) or Action Levels.

During the 1988 IRP Phase IV-A investigation, groundwater samples were collected from the three existing wells installed during the Phase II investigation (Figure 2-3). The groundwater samples were analyzed for total cyanide and total metals including arsenic, barium, cadmium, chromium, copper, lead, mercury, nickel, selenium, sodium (Table 2-3). Low concentrations of total metals included arsenic (8.1 to $24 \mu g/L$), barium (5.7 to 7.6 $\mu g/L$), cadmium (<0.12 to 0.48 μ g/L), chromium (8.9 to 9.4 μ g/L), copper (<7.8 to 8.3 μ g/L), lead (1.4 to 2.5 μ g/L), mercury $(<0.13 \text{ to } 0.16 \mu g/L)$, nickel $(<11 \text{ to } 14 \mu g/L)$, selenium $(0.54 \text{ to } 1.0 \mu g/L)$, and sodium (7,710)to 33,500 μ g/L). All metals detected except for sodium and arsenic exhibited concentrations between the practical quantitation limit and the method detection limit. The highest concentrations of total arsenic, barium, nickel, selenium, and sodium were detected in sample I-01; the highest concentrations of cadmium, total chromium, and total lead were detected in sample I-02; and the highest concentrations of total copper and total mercury were detected in sample I-03. Concentrations of metals detected in groundwater were below applicable Florida Primary and Secondary Drinking Water Standards and Federal MCLs or Action Levels. Total cyanide, which was not detected in 1984, was detected in samples I-01 and I-03 at concentrations of 7.4 and 24 μ g/L, respectively; however, these concentrations, are well below the Federal MCL of 200 μ g/L.

Based on results from the previous investigations, G&M conducted a Remedial Investigation in 1991, during which additional groundwater samples were collected. Three existing wells I-01, I-02, and I-03 were sampled. One new monitoring well, SP1-MW-0001, was installed east of Building 164 in 1991 (Figure 2-3). This new well was also sampled. VOCs, BNAs, and TAL metals were included in the analyses.

The remedial investigation continued in 1993, with Montgomery Watson collecting samples from all four monitoring wells (SP1-MW-0001, I-01, I-02, and I-03) associated with OU5/Site WP-1 (Figure 2-4). All groundwater samples were analyzed for TCL pesticides/PCBs and cyanide. Additionally, groundwater sample SP1-I-02 was analyzed for TCL VOCs, TCL BNAs, and TAL metals (total and dissolved).

In 1996, OHM Corporation completed an additional round of groundwater sampling from the four OU-5/Site WP-1 monitoring wells SPI-MW-0001, I-01, I-02, and I-03. This groundwater sampling event was completed to supplement previous groundwater sampling events and provide a current characterization of site conditions. Groundwater samples

collected by OHM Corporation were analyzed for TCL VOCs, TCL BNAs, TCL pesticides/PCBs, total metals and cyanide.

Groundwater results are compared to Florida Primary and Secondary Drinking Water Standards, Florida 62-770 Target Cleanup Levels, and Federal Primary and Secondary Drinking Water Standards (MCLs), presented in Table 2-12.

2.6.3.1 Volatile Organic Compounds. 1991 Investigation. In 1991, groundwater samples were collected from three existing monitoring wells (I-01, I-02, and I-03) and one newly installed monitoring well (SP1-MW-0001) (Figure 2-3). One VOC, methylene chloride, was detected in monitoring wells I-01 and I-02 at concentrations of 1.4 and 3.1 μ g/L, respectively (Table 2-6). These reported concentrations are below the PQL for methylene chloride (5 μ g/L) and also below the Florida Primary Drinking Water Standard, 5 μ g/L. Methylene chloride is also a common laboratory contaminant and concentrations at these low levels are not necessarily indicative of site contamination.

1993 Investigation. Groundwater samples were collected for VOC analysis at only one monitoring well at OU-5/Site WP-1 during the 1993 field investigation. Monitoring well SP1-I-02 was sampled and a duplicate was collected. Chloroform (2.52 μ g/L) and bromodichloromethane (1 and 2 μ g/L) were detected at less than the CRQL of 10 μ g/L. These values are also significantly below the Federal MCL of 100 μ g/L for each compound.

Field QA/QC samples indicated the presence of 1,2-dichloropropane in the equipment blank associated with the collection of SP1-I-02 and SP1-I-902. The concentration $(2 \mu g/L)$ is estimated below the CRQL. This compound has been detected in other equipment blanks and analyte-free water samples obtained during the 1993 investigation and is most likely associated with the use of the analyte-free water system. The QCSR discusses the full scope of quality assurance for the 1993 investigation and is submitted under separate cover. 1,2-dichloropropane was not detected in any of the groundwater samples collected. A summary of the 1993 groundwater analytical results for VOCs are presented in Table 2- 10.

1996 Investigation. Groundwater samples were analyzed for VOCs in the 1996 investigation from each of the four OU-5/Site WP-1 monitoring wells. Methylene chloride was the only compound detected and was reported in two samples and one duplicate, ranging in concentration from 1 μ g/L to 7 μ g/L. Methylene chloride was also present in an equipment blank at a concentration of 3 μ g/L. Methylene chloride is a common laboratory contaminant and was also present in the equipment blank. Methylene chloride was detected

GROUNDWATER QUALITY CRITERIA

Analyte	Florida Drinking Water Standards	Florida 62-770	EPA Drinking Water Standards	EPA Maximum Contaminant Level Goal
VOLATILE ORGANIC COMPOUNDS (ug/L): Chloroform Bromodichloromethane	NS NS	NS NS	100 100	0 0
BASE/NEUTRAL AND ACID EXTRACTABLE ORGANIC COMPOUNDS (ug/L) bis(2-Ethylhexyl) phthalate Di-n-octylphthalate 2-Methylnaphthalene Naphthalene	6 NS NS NS	NS NS d d	6 NS NS NS	0 NS NS NS
METALS (ug/L): Aluminum Arsenic Barium Beryllium Calcium Chromium Cobalt Copper Iron Lead Magnesium Manganese Mercury Nickel Potassium Selenium Sodium Thallium Vanadium Zinc	200 1 50 k 2000 k 4 NS 100 k NS 1000 1 300 1 300 1 15 k NS 50 1 2 k 100 k NS 50 k 160000 k 2 NS 5000 1	NS NS NS NS NS NS NS NS NS NS NS NS NS N		NS NS 2000 i 4 NS 100 i NS 1000 NS 0 NS 2 i 100 NS 2 i 100 NS 50 i NS 50 i NS 50 i NS
TOTAL RECOVERABLE PETROLEUM HYDROCARBONS (mg/L)	NS	5	NS	NS
TOTAL DISSOLVED SOLIDS (mg/L)	500 1	NS	500 h	NS
BIOCHEMICAL OXYGEN DEMAND (mg/L)	NS	NS	NS	NS
TOTAL SUSPENDED SOLIDS (mg/L)	NS	NS	NS	NS
ALKALINITY (mg/L)	NS	NS	NS	NS
TOTAL ORGANIC CARBON (mg/L)	NS	NS	NS	NS
SULFATE (mg/L)	250	NS	250	400/500 g
SULFIDE (mg/L)	NS	NS	NS	NS
HARDNESS as CaCO3 (mg/L)	NS	NS	NS	NS

ug/L - micrograms per liter

mg/L - milligrams per liter

NS - No Standard

b - The total of volatile organic aromatics (benzene, toluene ethylbenzene and xylenes) must be <50 ug/L to meet FAC 62-770 guidelines.

c - The total of polynuclear aromatic hydrocarbons excluding naphthalenes must be <10 ug/L to meet FAC 62-770 guidelines.

d - The total of naphthalenes and methyl naphthalenes must be <100 ug/L to meet FAC 62-770 guidelines.

f - Numbers represent EPA's Proposed Primary MCL or Proposal MCLG, Federal Register, Vol. 55, No. 143, July 1990.

g - Numbers represent EPA's Primary MCL for Inorganics.

 Numbers represent EPA's Secondary MCL for Inorganics which are non-enforceable taste, odor or appearance guidelines. Numbers represent EPA's Final MCL effective July 1992, Federal Register, January 30, 1991 and July 1, 1991.

- Florida Primary Drinking Water Standard.

1 - Florida Secondary Drinking Water Standard.

s - Final Action Level -The final lead action level is exceeded if the level of lead/copper in more than 10 percent of the targeted tap samples is greater than the action level (90th percent). in groundwater samplescollected from the same two wells (I-01 and I-02) during the 1991 RI. However, methylene chloride was not reported above the method detection limit in the 1993 groundwater sample collected from monitoring well SP1-I-02. A summary of the 1996 VOC analytical results are provided in Table 2-10.

2.6.3.2 Base Neutral/Acid Extractable Compounds. 1991 Investigation. All four monitoring wells (I-01 through I-03, and SP1-MW-0001) were sampled and analyzed for BNAs during the 1991 investigation. Only one BNA, bis(2-ethylhexyl)phthalate, was detected in monitoring wells I-02, I-03, and SP1-MW-0001 at concentrations of 0.9,320, and 0.7 mg/L, respectively (Table 2-6). Concentrations of bis(2-ethylhexyl)phthalate were classified as qualitative because they were either detected below the PQL or were observed in an associated blank sample. Bis(2-ethylhexyl)phthalate is a common laboratory contaminant. This compound is a plasticizer and is also commonly encountered in samples which have been in contact with plastics (gloves, sample containers, etc.).

1993 Investigation. One well was sampled and analyzed for BNAs during the 1993 field investigation. The sample, SP1-I-02 and its duplicate, SP1-I-902, were collected from monitoring well I-02. Sample SP1-I-02 indicated 0.5 μ g/L of di-n-butyl-phthalate and 0.3 μ g/L of bis(2-ethylhexyl)phthalate (Table 2-10). Both concentrations are below the CRQL of 11 μ g/L. Duplicate sample SP1-I-902 had reported concentrations of naphthalene (1 μ g/L), 2-methylnaphthalene (2 μ g/L), and di-n-butyl phthalate (0.5 μ g/L). Concentrations reported for all BNA constituents are estimated since they are below the CRQLs (11 μ g/L). The Florida Primary Drinking Water Standard for bis(2-ethylhexyl)phthalate is 6 μ g/L and the Federal MCL is 6 μ g/L. Groundwater concentrations of these compounds detected at OU-5/Site WP-1 were below these values. There is no established groundwater criteria for di-n-butylphthalate. As stated previously, phthalates are commonly encountered in samples exposed to plastics. The concentrations of naphthalene and 2-methylnaphthalene detected in the duplicate sample are below the Florida 62-770 guidelines of 100 μ g/L with a total naphthalenes concentration of 3 μ g/L.

1996 Investigation. Groundwater samples were collected and analyzed for BNAs from each of the OU-5/Site WP-1 wells in 1996. BNAs were not reported above the method detection limit in any of the groundwater samples. The 1996 groundwater analytical results for BNAs are presented in Table 2-10.

2.6.3.3 Organochlorine Pesticides/PCBs. 1991 Investigation. Groundwater samples collected in 1991 were not analyzed for pesticides and PCBs.

1993 Investigation. Groundwater samples were collected from all four monitoring wells at OU-5/Site WP-1 in March of 1993 and analyzed for TCL organochlorine pesticides and PCBs. Wells SP1-MW-0001, SP1-I-01, SP1-I-02, SP1-I-03, were sampled and submitted for analysis. Pesticides and PCBs were not detected above CRQLs in any of the four groundwater samples, and one duplicate sample collected. A summary of the 1993 groundwater analytical results for pesticides/PCB s is provided in Table 2-10.

1996 Investigation. Groundwater samples were collected from the four monitoring wells at OU-5/Site WP-1 during the 1996 investigation. Pesticides and PCBs were not reported above the detection limit in any of the samples collected. A summary of the 1996 groundwater analytical results for pesticides/PCBs is provided in Table 2-10.

2.6.3.4 Metals and Cyanide. 1991 Investigation. Three existing monitoring wells (I-01 through I-03) and the one newly installed well (SP1-MW-0001) were sampled during the 1991 investigation. A duplicate of sample I-02 was also collected (I-9002). The following TAL metals were detected in groundwater samples collected at OU-5/Site WP-1: aluminum, barium, calcium, chromium, iron, magnesium, manganese, potassium, sodium, vanadium, zinc, arsenic, and lead (Table 2-6). Calcium, magnesium, and potassium were detected in the four wells sampled; however, no Florida Drinking Water Standards or Federal MCLs exist for these metals. Groundwatersamples I-01, I-02 and I-9002 contained very high concentrations of total calcium, 3,100,000, 5,400,000, and $2,600,000 \mu g/L$, respectively, in addition to significant concentrations of many other TAL metals. Review of the groundwater sampling logs for these samples indicated that all samples were turbid during sample collection. It is possible that the high TAL metal concentrations, particularly that of calcium, were a result of suspended sediments as artifacts of well construction and thereby overstated the actual concentrations of the analytes at the site. Calcium concentrations reported in groundwater samples I-02, I-9002, and I-03 were much higher than the range of dissolved calcium concentrations (55,000 to 140,000 µg/L) reported in the Biscayne Aquifer by Sonntag (1987).

Concentrations of barium detected in groundwater samples collected at OU-5/Site WP-1 were well below the Florida Primary Drinking Water Standard of 2,000 μ g/L and the Federal PMCL of 2,000 μ g/L. In addition, concentrations of sodium detected were well below the Florida Primary Drinking Water Standard of 160,000 μ g/L. Estimated concentrations of arsenic detected in monitoring wells I-01 (92 μ g/L) and I-02 (60 μ g/L) exceeded the Florida Primary Drinking Water Standard and Federal MCL for arsenic of 50 μ g/L. Concentrations of lead detected in samples I-02 and I-9002 (duplicate of I-02) exceeded both the Florida

Primary Drinking Water Standard and the Federal Action Level of 15 μ g/L. Concentrations of chromium detected in sample I-02 exceeded the Florida Primary Drinking Water Standard and the Federal MCL of 100 μ g/L. The high concentrations of calcium interfered with the ability to detect cadmium in samples I-01, I-02, and I-9002 resulting in detection limits for cadmium which exceed the Florida Primary Drinking Water Standard and the Federal MCL of 5 μ g/L. Additionally, the high concentrations of calcium interfered with the ability to detect selenium resulting in detection limits for all samples of 50 μ g/L.

Federal Secondary Drinking Water Standards establish recommended limits and deal with the aesthetic qualities of drinking water; however, the FDEP has adopted these standards as the Florida Secondary Drinking Water Standards and equires that potable groundwater shall meet these recommended limits. Concentrations of iron are naturally high and commonly exceed the Florida standard (Sonntag, 1987). Concentrations of iron detected in all groundwater samples collected in 1991 at OU-5/Site WP-1 exceeded the Florida Secondary Drinking Water Standard and Federal Secondary MCL of 300 μ g/L. Concentrations of manganese detected in groundwater samples SP1-MW-0001, I-01, I-02, and I-9002 exceeded the Florida Secondary Drinking Water Standard and Federal Secondary MCL of 50 μ g/L. Concentrations of aluminum detected in all groundwater, samples exceeded the Federal Secondary MCL for aluminum of 50-200 μ g/L. However, concentrations of total dissolved solids and zinc did not exceed the Florida Secondary Drinking Water Standards and Federal Secondary MCL sin samples analyzed for these constituents.

The metal concentrations detected in the groundwater samples collected during the 1991 CERCLA investigation had the highest concentrations of metals of all sampling events. As discussed above, it is likely that the increase in metal concentrations is the result of suspended sediments in the groundwater samples collected.

Cyanide was not detected in any of the groundwater samples collected in 1991.

1993 Investigation. All four existing monitoring wells (SP1-MW-0001, SP-I-01, SP1-I-02, and SP1-I-03) were sampled in 1993 and analyzed for cyanide. Cyanide was not detected above the CRQL in any of the groundwater samples.

Only well SP1-I-02 was sampled and analyzed or TAL metals. Monitoring well SP1-I-02 was sampled in duplicate (SP1-I-902) and an equipment blank was collected. Due to high turbidity encountered in groundwater samples during the 1991 sampling events, both total (unfiltered) and dissolved (filtered) metals were analyzed (Table 2-10). Dissolved metals

samples were filtered at the time of collection using a 0.45 micron disposable (single use) filter. The filtered sample was placed in the appropriate sample container and preserved with nitric acid to achieve a pH of <2. The pH was tested on all metals samples prior to shipment to the laboratoryto ensure that proper preservation (pH <2) had been achieved. Savannah Laboratories, Tallahassee, again checked the pH upon receipt at the laboratory. No report of inadequate preservation was noted in the analytical data.

Aluminum, arsenic, barium, calcium, copper, iron, lead, magnesium, manganese, potassium, sodium, vanadium, and zinc were detected in the unfiltered samples. Analysis of filtered samples showed lead, manganese, and vanadium had been removed to below detectable limits. Aluminum concentrations were reduced significantly (by as much as three orders of magnitude) in the filtered vs. unfiltered groundwater samples.

Evaluation of dissolved metals indicate that arsenic, copper, magnesium, potassium, sodium, and zinc are comparable to the total concentrations detected in groundwater. Total arsenic concentrations were 18.4 and 18.1 μ g/L in SP1-I-02 and the duplicate, while dissolved arsenic concentrationswere 16.1 and 16.7 μ g/L. These values are well below the 60 μ g/L reported during the 1991 sampling event. Total copper was reported at 3.1 and 3.7 μ g/L and dissolved copper at 3.5 and <3.0 μ g/L. Copper results were below the CRQL. Magnesium was detected in the unfiltered samples at 2,780 and 2,980 μ g/L and at 2,230 and 2,290 μ g/L in the filtered samples. Potassium was reported at 4,410 and 4,95 μ g/L (total) and 3,250 and 3,750 μ g/L (dissolved). Potassium results are below the CRQL. Sodium data indicated 14,300 and 14,500 μ g/L (total) and 13,700 and 13,800 μ g/L (dissolved).

Zinc was detected at 33.5 and 18.2 μ g/L in the unfiltered samples and at 44.1 and 15.3 μ g/L in the filtered samples. Zinc was also detected in the equipment blank at 36.1 μ g/L and is a known contaminant at low levels in acids used for preservation and digestion of samples.

Arsenic, copper, magnesium, potassium, sodium, and zinc concentrations were all below the established regulatory guidance levels or fell within the ange of concentrations reported in the Biscayne Aquifer, if no guideline for the metal in groundwater was available.

Aluminum, barium, calcium, iron, lead, manganese, and vanadium showed significant decreases in concentrations between total and dissolved metal results. Total aluminum was reported at 1,850 and 2,610 μ g/L in SP1-I-02 and the duplicate SP1-I-902. The Florida Secondary Drinking Water Standard is 200 μ g/L. The Federal Secondary MCL is 50-200 μ g/L. The filtered samples indicated a significant decrease in aluminum concentrations to <20 and 23.8 µg/L in SP1-I-02 and SP1-I-902, respectively. High concentrations of aluminum have been observed in background soil samples at HARB and in carbonate soils. Reported barium concentrations are below the Federal MCL of 2,000 µg/L in both total and dissolved analyses.

There are no Federal MCLs or Florida Drinking Water Standards for calcium. The dissolved calcium concentrations (78,800 and 82,200 μ g/L) are within the range reported by Causaras, 1987, for concentrations of dissolved inorganic constituents in the Biscayne Aquifer. These high calcium concentrations contribute interference in analytical determinations for other metals.

Total iron detected at 1,560 and 1,950 μ g/L is within or near the range detected as a dissolved constituent in the Biscayne Aquifer (<10 to 1,900 μ g/L). These detected concentrations are higher than the Federal Secondary MCLs and Florida Secondary Drinking Water Standards of 300 μ g/L. However, the filtered samples contained 40.5 and 50.0 μ g/L dissolved iron for SP1-I-02 and SP1-I-902. The dissolved iron values are below the CRQL for the method and are also below the Federal Secondary MCL and the Florida Secondary Drinking Water Standard. As reported by Sonntag in 1997, concentrations of iron are naturally high and commonly exceed the Florida Secondary Drinking Water Standard.

Lead was detected in only the unfiltered groundwater metals analysis (9.3 and 9.1 μ g/L. Total (unfiltered) lead levels were below the Federal MCL of 15 μ g/L. Total manganese (26.7 and 28.5 μ g/L) did not exceed the Florida Secondary Drinking Water Standard or the Federal Secondary MCL of 50 μ g/L. Manganese was not detected in any of the filtered samples. No established guidelines for vanadium are available. Total vanadium values reported (5.5 and 6.5 μ g/L) are below the CRQL. Vanadium was not detected above the CRQL in the filtered samples.

Thallium has a Florida Primary Drinking Water Standard of $\mu g/L$, a Federal MCL of $2 \mu g/L$, and a Federal MCLG of 0.5 $\mu g/L$. Thallium was not detected above the CRQL of 5 $\mu g/L$ in any of the groundwater samples collected from SP1-I-02.

Though not detected at the attainable CRQLs, selenium and thallium results are qualified due to technical interferences caused by the high calcium typically observed in samples obtained in south Florida.

1996 Investigation. Groundwater samples wereobtained from the four wells associated with OU-5/Site WP-1 (Figure 2-7) during the 1996 investigation and analyzed for total metals and cyanide. The groundwater analytical results obtained during this investigation are provided in Table 2-10. No filtered (dissolved) analyses were performed on the 1996 groundwater samples. Instead, care was taken during groundwater sampling to minimize the turbidity of the samples. The groundwater was sampled once the turbidity levels were below 10 NTU. Cyanide, aluminum, beryllium, cadmium, lead, mercury, selenium, silver, and thallium were not detected above the detection limit in the four groundwater samples collected in 1996.

Five metals, arsenic, barium, calcium, magnesium, and sodium, were detected in each of the four monitoring wells at OU-5/Site WP-1. Additionally, copper, iron, lead, manganese, potassium, vanadium, and zinc were detected in one or more of the groundwater samples. Concentrations of these metals were all below the Federal MCLs and Florida Drinking Water Standards. The metals, chromium, cobalt, copper, and nickel werconly reported above detection limits in well SP1-I-03. Concentrations of these metals in well SP1-I-03 were reported below the Federal MCL and Florida Drinking Water Standards.

The calcium, magnesium, and sodium concentrations, which range from 83,900 μ g/L to 114,000 μ g/L to 2,740 μ g/L, and 4,860 μ g/L to 10,600 μ g/L are within the range of values for dissolved inorganics detected in the Biscayne Aquifer.

Arsenic, which ranged in concentration from $9 \mu g/L$ to $15.3 \mu g/L$, is above the range of $<1.2 \mu g/L$ and the mean of $1.2 \mu g/L$ for dissolved organic constituents detected in the Biscayne Aquifer. Arsenic was detected at concentrations below the Federal MCL and Florida Primary Drinking Water Standard of $50 \mu g/L$.

2.6.3.5 Summary Section for Groundwater.

Impacts to groundwater as a result of past operation do not indicate significant impacts as determined by groundwater samples obtained in 1991, 1993, and 1996. Comparison of the groundwater results collected in 1991 with those collected in 1993 and 1996 indicate elevated concentrations of constituents, primarily metals in the 1991 groundwater samples. This is likely due to the turbid conditions which were reported in the 1991 samples. Comparison of the range of the calcium concentrations in the 1991 groundwater samples, (130,000 μ g/L to 5.4E+06 μ g/L) with the 1993 (209,000 μ g/L to 220,000 μ g/L) and 1996

 $(83,400 \ \mu g/L \text{ to } 114,000 \ \mu g/L)$ samples further substantiates the fact that turbid groundwater samples were collected.

The groundwater compounds detected in OU-5/Site WP-1 wells include VOCs (chloroform and bromodichloromethane), BNAs (naphthalene, 2-methylnaphthalene, di-n-butyl phthalate and bis(2-ethylhexyl)phthalate)and various naturally occurring metals. The organic contaminants are observed at very low levels. Methyle chloride was detected in monitoring wells I-10 and I-02 during the 1991 and 1996 groundwater sampling events. However, methylene chloride is a common laboratory contaminant and may not be indicative of groundwater impacts. Although arsenic exceeded FederalMCLs and Florida Primary Drinking Water Standards in one sample collected in 1991, this has been attributed to the significant turbidity pepted in these samples. Aluminum and iron are the only compounds which exceed Federal MCLs and/or Florida Drinking Water Standards. Elevated concentrations of the metals are indicative of the Biscayne Aquifer in South Florida. Thallium was not detected, but the quantitation limit is above Federal MCLs and Florida Drinking Water Standards.

2.6.4 Sediment Investigations

This summary of sediment investigations is presented for the purpose of review only. Sediments have been fully evaluated in the investigation of OU-9, Boundary Canal.

Various sediment samples have beencollected from the drainage swale south of Building 159, and the unlined drainage swale after it exits the underground culvert south of OU-5/Site WP-1 south and west of the equipment storage area (Figures 2-2, 2-3, 2-4, 2-5, and 2-6).

Because background sediment samples were not collected during the 1984 Phase II investigation and the 1991 CERCLA investigations at OU-5/Site WP-1, concentrations of constituents detected in the sediment samples collected during the OU-5/Site WP-1 investigations were compared with the background sediment sample, BC-SD-0010, from the 1991 Boundary Canal sampling event. The Boundary Canal background sample was obtained from a location upgradient of all industrial (PSCs) sites at Homestead ARB. Constituents detected in this sample are assumed to not be the result of the past or present industrial waste-handling activities/practices. Sediment sample BC-SD-0010 was analyzed for TCL BNAs, organochlorine pesticides, and TAL metals in 1991. Two sediment samples were collected in 1984 from the drainage swale located south of Building 159 (Figure 2-2). These sediments were subsequently excavated during the 1995 Interim Action. The sediment samples were analyzed for cyanide and total metals including cadmium, total chromium, copper, lead, nickel, and zinc (Table 2-2). Low concentrations of these constituents were detected in the sediment samples as follows: cyanide concentrations ranged from 0.6 to 3.9 mg/kg, cadmium concentrations ranged from 0 to 0.01 mg/kg, total chromium concentrations ranged from 0.02 to 0.09 mg/kg, copper concentrations ranged from 0.01 to 0.04 mg/kg, lead concentrations ranged from 0.11 to 1.18 mg/kg, nickel concentrations ranged from 0 to 0.01 mg/kg, and zinc concentrations ranged from 0.8-0.13 mg/kg. Concentrations of total chromium, lead, zinc, and cyanide detected in the sediment samples were slightly higher than the concentrations detected in the surficial soil samplescollected during the Phase II investigation in 1984; however, concentrations of chromium, copper, lead and zinc detected in the OU-5/Site WP-1 sediment samples were well below the Boundary Canal background concentrations. Concentrations of cadmium, total chromium, copper, lead, nickel and zinc were well below their respective NOAA ER-L and ER-M values.

The 1991 CERCLA RI included the collection of two sediment and surface water samples in the unlined drainage swale approximately 400 ft south of Building 164 and just south of the equipment storage area (Figure 2-3).

The 1993 RI field activities included the collection of six sediment samples from the drainage areas: the grassy drainage swale east of Building 164, the grassy drainage swale south of Building 159, and the unlined drainage swale south of Building 164, which was also sampled in 1991. Samples SP1-SD-0007 and SP1-SD-0008 were collected from the same locations as SP1-SD-0005 and SP1-SD-0006 (1991). SP1-SD-0012 was collected at the south end of the north/south drainage swale east of Building 164 while SP1-SD-0009 through SP1-SD-0011 were collected from the east/west drainage swale south of Building 159. Sediment samples SP1-SD-0009 through SP1-SD-0012 were subsequently excavated during the 1995 Interim Action. All sediment samples collected in 1993 were analyzed for USEPA TCL organic compounds, TAL metals, and yanide. Figures 2-3 and 2-4 illustrates the locations of the 1991 and 1993 sediment samples.

In 1994, IT Corporation completed a confirmation sampling program which included the collection of sediment samples from the east/west drainage swale south of Building 159 and the unlined drainage swale which extends south from Building 164. Twenty-one sediment samples were collected from the locations depicted on Figure 2-5. These samples were

analyzed for BNAs, pesticides and TAL metals. Based on the sults of the 1994 investigation, an Interim Action was performed by IT Corporation in 1995 to excavate affected sediments from the North Area drainages. After excavation, three confirmation samples were collected from the base of the excavation area. Two of the samples, CS-02 and SC-03, were considered to be sediments based on their locations in the drainage system (Figure 2-6).

Interim Action activities involved excavating the sediments to a depth of approximately 1 ft bls at each location. The excavation was extending out 3 ft from the centerline of the drainage swale. During the excavation activities, 12 sediments and 2 soil analytical sampling points from previous investigations were removed (Table 2-4). The results from samples collected at these locations are no longer considered representative of current site conditions.

2.6.4.1 Volatile Organic Compounds. **1991** Investigation. Two VOCs, tetrachloroethene, and trichloroethene, were detected in the 1991 sediment samples (Table 2-7). Tetrachloroethene and trichloroethene were detected in sample SP1-SD-0006 at concentrations of 29 and 12 μ g/kg, respectively. These VOCs are not commonly used in the electroplating operations. A possible source for the tetrachloroethene and trichloroethene is the adjacent equipment storage area and motor pool. Tetrachloroethene and trichloroethene are solvents used during degreasing operations as well as other processes associated with automobile maintenance. These constituents may be present in waste oils in minor amounts because through historical waste oil handling practices solvent residue may have become incorporated into the waste oil. These compounds were not detected in the sample (SP1-SD-0008) collected from a similar location in 1993.

1993 Investigation. Six sediment samples plus one duplicate were collected and analyzed during the 1993 field investigations. The samples were identified as SP1-SD-0007 through SP1-SD-0012 plus duplicate sample SP1-SD-9012. Sediment sampling points SP1-SD-0008 through SP1-SD-00012 have been subsequently excavated. Two VOCs were detected: methylene chloride was detected in only one sample (SP1-SD-0010) while acetone was detected in two samples and the duplicate. Sample SP1-SD-0007 contained acetone outside the calibration range of the instrument in the initial analysis. The subsequent dilution showed <16,000 µg/L. The results are considered qualitative only for the positive presence of acetone although a quantity was not reported. Samples SP1-SD-0012 and SP1-SD-90012 reported acetone at 8 and 15 µg/L, respectively. These values are less than the CRQL. As previously discussed, acetone has been widely detected throughout Homestead ARB and has been identified as an artifact of the isopropyl alcohol used during field decontamination.

Acetone does not appear to be a soil/sediment contaminant at OU-5/Site WP-1. Additional discussion of the presence of acetone is provided in the QCSR provided to the USACE, Omaha-District.

The methylene chloride detected at 8 μ g/L is below the CRQL. As previously discussed, this compound is a common contaminant in laboratories. Though not reported in an associated blank, it is unlikely, at the reported concentration, to be a site contaminant. Table 2-11 summarizes constituents detected during the 1993 investigation.

1994 and 1995 Interim Action. The sediment samples collected during the 1994 Confirmation Sampling and 1995 Interim Action were not analyzed for VOCs.

2.6.4.2 Base Neutral/Acid Extractable Compounds. 1991 Investigation. Several BNA compounds (mainly PAHs), including acenaphthene, anthracene, benzo(a)anthracene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(g,h,i)perylene, benzo(k)fluoranthene, chrysene, dibenzo(a,h)anthracene, fluoranthene, fluorene, indeno(1,2,3-c,d)pyrene, phenanthrene, and pyrene, were detected in the two sediment samples collected south of the equipment storage area. For the most part, concentrations of PAHs detected in sediment samples SP1-SD-0005, SP1-SD-9005, and SP1-SD-0006 were above concentrations detected in background sample, BC-SD-0010 (Table 2-7).

The unlined drainage swale sampled contains surface-water intermittently, after rain storms, and it does not support aquatic life. To be consistent with other Homestead ARB site investigations, the PAH concentrations detected in these sediment samples were compared with National Oceanic and Atmospheric Administrative effects range-low and effects range-median (NOAA ER-L and ER-M) values and sediment quality criteria (SQC) values, which are indicators of risk to aquatic life. Concentrations of anthracene, benzo(a)pyrene, benzo(a)anthracene, chrysene, dibenzo(a,h)anthracene, fluoranthene, phenanthrene, and pyrene detected in sediment samples SP1-SD-0006 and SP1-SD-0005 (and/or its duplicate SP1-SD-9005) were above the NOAA ER-L and ER-M values for these constituents. The concentrations of acenaphthene and fluorene detected in sample SP1-SD-0005 were above both their respective NOAA ER-L and ER-M values and the concentrations detected in SP1-SD-0006 were above their respective NOAA ER-L value.

Based on USEPA guidance for sediment samples with an unknown organic carbon content, sediment quality criteria for sediment containing one percent organic carbon were compared to the unlined drainage swale sediment samples. SQC for benzo(a)anthracene and

benzo(a)pyrene were exceeded in samples SP1-SD-0005 and SP1-SD-9005 and the SQC for fluoranthene, pyrene, and phenanthrene were exceeded in samples SP1-SD-0005, SP1-SD-9005, and SP1-SD-0006. PAHs are not commonly associated with electroplating operations. Possible sources for the PAHs are the adjacent equipment storage area and motor pool and road runoff. PAH compounds are commonly present in waste automotive oil and asphalt.

Additionally, two non-PAHs were detected in the sediment samples including bis(2ethylhexyl)phthalate, which was detected in samples SP1-SD-0005 and SP1-SD-0006 at concentrations of 6100 and 1700 μ g/kg, respectively, and dibenzofuran which was also detected in sample SP1-SD-0005 at a concentration of 2000 μ g/kg. Concentrations of bis(2ethylhexyl)phthalate detected in sediment samples SP1-SD-0005 and SP1-SD-0006 were above background concentrations detected in BC-SD-0010. Dibenzofuran was not detected in the background canal sediment.

1993 Investigation. Six sediments and one duplicate sample were collected and analyzed for TCL BNA compounds. Four of the 1993 sediment sampling locations have undergone excavation during the 1995 Interim Action. Sediments from SP1-SD-0009 through SP1-SD-0012 are no longer considered representative of site conditions. Eighteen PAHs were detected, as well as four phthalate compounds, dibenzofuran, carbazole, naphthalene, 2-methylnaphthalene, and 2,6-dinitrotoluene. The BNA compounds listed in Table 2-11 are similar to the compounds detected in 1991 (Table 2-7). Sample SP1-SD-0007 and SP1-SD-0008 collected in 1993 correspond to the same locations sampled in 1991 and identified as SP1-SD-0005 and SP1-SD-0006. The compounds detected in 1991 were reported at much lower concentrations in 1993 data. Concentrations of PAHs detected in sample SP1-SD-0007 were lower than the PQLs of the background sample (BC-SD-0100) used for comparison by G&M. With the exception of SP1-SD-0007 the PAH quantities reported are above the NOAA Median Effect Range for sediment quality. The concentrations of individual PAH compounds range from 5 μ g/kg (2-methylnaphthalene) to in excess of 45,000 μ g/kg of chrysene; 49,000 μ g/kg fluoranthene; and 110,000 μ g/kg pyrene in sample SP1-SD-0010.

The highest concentrations of PAHs occur in the east/west drainage swale located south of Building 159 in samples SP1-SD-0010 and SP1-SD-0011. These samples have been subsequently excavated. SP1-SD-0012 contained concentrations of PAHs at values comparable to SP1-SD-0011. Sample SP1-SD-0012 was collected from the southern end of the open drainage swale east of Building 164. Samples SP1-SD-0009 contained the same

PAHs as SP1-SD-0010 but in lower quantities. SP1-SD-0009 is downstream of the directional flow in the drainage swale from SP1-SD-0010. Samples SP1-SD-0007 and SP1-SD-0008 were collected further downstream of the directional flow of all the previously discussed samples. The two sediments, SP1-SD-0007 and SP1-SD-0008, contained the lowest quantity of detected PAHs for OU-5/Site WP-1.

Observations made during the March 1, 1993 investigation, indicated that a large diesel powered generator was operating approximately 20 ft northeast of monitor well I-03. An apparent fuel spill was observed during this event, which is believed to have occurred during fueling of the 250-gallon diesel fuel tank used to run the generator. Fuel was observed in the north-south drainage swale to the east of Building 164. This area corresponds with sample point SP1-SD-0012.

Dibenzofuran was detected in all the 1993 sediment samples at quantities ranging from $3 \mu g/kg$ to 1,100 $\mu g/kg$ (SP1-SD-0010). All reported quantities were below the CRQL and therefore are estimated. Di-n-butyl phthalate was detected in two samples, SP1-SD-0007 (23 $\mu g/kg$) and SP1-SD-0011 (570 $\mu g/kg$). This compound was also reported in the associated lab blank. Both reported sample values are below the CRQL and are evaluated as non-detects in the QCSR. Two additional phthalates were detected: benzyl butyl phthalate and bis(2-ethylhexyl)phthalate, at levels ranging from 53 $\mu g/kg$ to 4,000 $\mu g/kg$ (SP1-SD-0012, bis(2-ethylhexyl)phthalate). Carbazole was detected in five of the six sediment locations ranging from 330 to 8,000 $\mu g/kg$ (SP1-SD-0010). Only the 8,000 $\mu g/kg$ result was above the CRQL.

Naphthalene was detected in two sediment samples SP1-SD-0009 and SP1-SD-0012 at 320 and 130 μ g/kg, respectively. 2-methylnaphthalene was detected in all sediments except SP1-SD-0008 at ranging from 6 μ g/kg to 79 μ g/kg. All reported values are below the CRQL. Naphthalenes were not reported as detected in the background canal sediment. Naphthalenes have not previously been detected at OU-5/Site WP-1 and are not an anticipated by-product of electroplating operations.

The duplicate sample, SP1-SD-9012 contained $540 \mu g/kg$ of 2,6-dinitrotoluene. No other samples collected at this site had reported 2,6-dinitrotoluene at detectable quantities.

Concentrations in the sediments are higher than those detected in the soil samples collected at OU-5/Site WP-1 during any of the previous investigations. However, 1993 sediment results

for samples collected from the drainage swale 400 ft south of OU-5/Site WP-1 were lower than at the same locations sampled in 1991.

1994 and 1995 Interim Action. Twenty-one sediment samples were collected and analyzed for BNAs during the 1994 confirmation sampling program (Figure 2-5). Nine PAH compounds were reported in most of the sediments collected during the 1994 sampling event. The PAHs include; phenathrene, anthracene, fluoranthene, pyrene, benzo(a)anthracene, chrysene, benzo(b)fluoranthene, benzo(k)fluoranthtene, and benzo(a)pyrene. Addition BNAs detected in one or more of the samples include; acenapthene, acenaphylene, fluorene, bis(2-ethylhexyl)phthalate, indeno(1,2,3-c,d)pyrene, benzo(g,h,i)perylene, and dibenz(A,H)anthrancene.

The 1994 sediment sample concentrations in the east/west drainage ditch east of Building 164 and north of Building 153 include samples EWA-8 through EWA-17. Concentrations of PAHs in this drainage are highest in the EWA-9, EWA-11, and EWA-12 samples (Table 2-11). This is likely a result of their location immediately adjacent to areas which receive runoff from roadways and asphalt pavement. Concentrations of PAHs at these locations ranged from 3,100 μ g/kg to 41,000 μ g/kg. Sediments associated with the samples EWA-1 and EWA-12 have subsequently been excavated. The two 1995, post excavation samples, CS-02 and CS-03 have as much as one to two orders of magnitude lower concentrations of PAH compounds than the pre-excavation samples. Benzo(a)anthracene, fluoranthene, and pyrene in the post excavation sediment sample CS-02 are greater than 2 times the background sediment concentration.

Similarly, in the South Area, the PAH sediment concentrations are slightly elevated in samples EWA-18, EWA-19, EWA-20, and EWA-22. These points are the uppermost points in the South Area, just below the culvert (Figure 2-5). PAH concentrations may be slightly more elevated at these points due to runoff passing through the culvert and potentially dissipating quickly into the underlying formations. Concentrations of PAHs in the South Area range from 3,300 μ g/kg to 55,000 μ g/kg. There were no excavation activities in the South Area. Sediment concentration from the 1994 Confirmation Sampling and 1995 Interim Action have been summarized on Table 2-11.

2.6.4.3 Organochlorine Pesticides/PCBs. 1991 Investigation. Sediment samples were not analyzed for organochlorine pesticides/PCBs in the 1991 investigation.

1993 Investigation. Of the six sediments collected during the 1993 investigation only SP1-SD-0008 showed no pesticide presence. This sample did, however, contain a reported 870 μ g/kg of PCB 1260 (Table 2-11). This is below any Toxic Substances Control Act (TSCA) PCB spill cleanup guidelines (40 CFR 761.120). Field observations of March 1,1993, noted a downed power line and pole east of well I-03. A transformer was not observed in the area.

The other five sediments plus the duplicate contained p'p'-DDT, p'p'-DDE, p'p'-DDD, and alphaand beta-chlordane. The highest concentrations of DDT (1,200 μ g/kg) and the chlordanes (2,400 and 2,800 μ g/kg), were at SP1-SD-0012. The downstream samples showed lower concentrations of all pesticides detected. The full extent and source of pesticide contamination was not determined during this investigation. It is possible that local use of these pesticides has contributed to the concentration in sediments via surface run-off.

1994 and 1995 Interim Action. Organochlorine pesticides were analyzed for in the 1994 and 1995 sediment samples. However, PCBs were only analyzed for in the two 1995 post excavation sediment samples. The primary pesticides detected in the 1994 and 1995 sediment samples were DDT and its metabolites and chlordane. Concentrations of DDT and its metabolites ranged from 7.9 μ g/kg to 620 μ g/kg, while chlordane ranged from 82 μ g/kg to 1,500 μ g/kg. The highest reported concentrations of pesticide in the North Area were associated with sample EWA-11 and EWA-12. Pesticide concentrations in the 1995 post excavation samples are consistent with the pre-excavation concentrations in sample CS-02 and at reduced levels in sample CS-03.

In the South Area, higher concentrations were associated with sample EWA-18. As with the BNA compounds, pesticides levels appear to be concentrated at the exit point of the culvert which may be due to transport of contaminants through the culvert and then once it exits rapidly infiltrate into the underly formation. Pesticides have been used throughout the Homestead-ARB. The concentrations observed at the Base are indicative of these orthopogenic sources. A summary of pesticide results from the 1994 Confirmation Sampling and 1995 Interim Action are presented in Table 2-11.

2.6.4.4 Metals and Cyanide. 1991 Investigation. During the 1991 investigation, two sample points were collected from approximately 400 ft south of the defined area of OU-5/Site WP-1. TAL metals detected in the sediment samples included aluminum, barium, cadmium, calcium, chromium, cobalt, copper, iron, magnesium, sodium, vanadium, zinc, mercury, arsenic, and lead. As previously discussed, these metals are commonly detected in limestone. Concentrations of most constituents detected in the sediment samples were below

average carbonate concentrations except for the trace metals, barium, cadmium, chromium, cobalt, copper, zinc, arsenic, mercury, and lead, which were above concentrations given for the average carbonate composition (Hem, 1989). In addition, concentrations of aluminum, barium, cadmium, chromium, cobalt, copper, iron, magnesium, manganese, vanadium, zinc, mercury, arsenic, lead and cyanide detected in at least one of the sediment samples collected were above background soil concentrations detected for OU-5/Site WP-1.

Concentrations of barium, chromium, iron, lead, mercury, vanadium and zinc detected in sediment samples SP1-SD-0005, SP1-SD-9005, and SP1-SD-0006 were greater than the concentrations of these metals detected in the background sample, BC-SD-0010. In addition, concentrations of cadmium detected in samples SP1-SD-0006 and SP1-SD-9005, copper in sample SP1-SD-9005, and concentrations of arsenic detected in samples SP1-SD-0005 and SP1-SD-0005 were above background concentrations of these metals detected in sample BC-SD-0010. Concentrations of barium, lead, mercury, and zinc detected in samples SP1-SD-0005 and its duplicate, SP1-SD-9005, located closest to the equipment storage area contained the highest concentrations above background.

Concentrations of chromium detected in sample SP1-SD-0005 exceeded the NOAA ER-L and ER-M values. Concentrations of zinc in samples SP1-SD-0005, SP1-SD-9005 and SP1-SD-0006 exceeded the NOAA ER-L values and concentrations detected in SP1-SD-0005 and SP1-SD-9005 exceeded the NOAA ER-M value. Sediment samples SP1-SD-0005, SP1-SD-9005 and SP1-SD-0006 contained mercury concentrations which exceeded the NOAA ER-L value, and additionally, the concentration of mercury detected in sample SP1-SD-0005 exceeded the NOAA ER-M value. Lead concentrations detected in samples SP1-SD-0005 and SP1-SD-9005 exceeded the NOAA ER-M and ER-L values and sample SP1-SD-0006 exceeded the NOAA ER-M value. Lead concentrations detected in samples SP1-SD-0005 and SP1-SD-9005 exceeded the NOAA ER-M and ER-L values and sample SP1-SD-0006 exceeded the NOAA ER-M value. Table 2-7 summarizes metals detected in the 1991 sediment samples.

Cyanide was detected in sample SP1-SD-9005 at a concentration of 2.2 mg/kg. Because of the distance from OU-5/Site WP-1 and the time elapsed since the Electroplating Facility was in operation, OU-5/Site WP-1 is an unlikely source for the cyanide. Additional sources of cyanide include herbicides, rodenticides, insecticides, and fungicides.

1993 Investigation. Metals analyses were performed on the six sediment samples collected in 1993 (Figure 2-4). Four of the sampling point were subsequently excavated during the 1995 Interim Action. The metals detected during the 1993 sampling activities are the same metals detected in 1991 with the addition of antimony and potassium. However, the

antimony values reported in 1993 are all below the PQLs reported for the 1991 samples. The sampling in 1993 was more extensive and include drainage swales within the OU-5/Site WP-1. Table 2-11 summarizes the 1993 investigation results for sediments.

Aluminum, calcium, cobalt, magnesium, and sodium values are similar to background concentrations in the Boundary Canal. The iron, manganese, nickel, potassium, and vanadium concentrations reported are less than the average carbonate compositions for those analytes.

Barium is below the carbonate composition (30 mg/kg) at SP1-SD-0012 and SP1-SD-0010, but considerably above the average at SP1-SD-0007 (635 mg/kg), and SP1-SD-0009 (201 mg/kg). Cadmium was detected at all points in the range of 1.6 to 4.6 mg/kg with the highest concentration at SP1-SD-0011. Chromium was detected in all samples (15-810 mg/kg) with the highest concentration at SP1-SD-0009). All values for chromium were above background and average carbonate concentrations. Copper was detected in all samples (14.6 to 61.1 mg/kg) with only 1 sample result below the Boundary Canal background value.

Lead values ranged from 87.6 to 1,180 mg/kg. Mercury was detected at five of the six sampling locations. Sample SP1-SD-0008 contained <0.11 mg/kg mercury. Concentrations at the other sampling points ranged from 0.25 to 4.4 mg/kg mercury with SP1-SD-0009 the highest reading. Zinc concentrations ranged from 58.9 to 612 mg/kg.

The reported values for antimony, arsenic, barium, chromium, iron, lead, and zinc are qualified as qualitative data due to technical difficulties encountered during analysis. The primary source of these technical interferences is the high calcium inherent in these samples from south Florida. Additional discussion of these technical interferences and the qualification of this data is presented in the associated QCSR.

The concentration reported for cobalt, potassium, sodium, and most of the nickel values are below the CRQL.

In summary, the highest metals concentrations generally occur in the grassy east/west drainage swale south of Building 159. The elevated lead concentration (1,180 mg/kg) is considered to be representative of a limited area within the swale and is considered an isolated detection. This sample point has been subsequently excavated. Sample point SP1-SD-0007 also shows higher concentrations of metals than downstream sample SP1-SD-0008.

Cyanide was not detected in any of the sediment samples collected during the 1993 investigation.

There were no significant differences in the sediment analytical results for samples collected in 1991 from samples collected in 1993 at sites SP1-SD-0005 and SP1-SD-0006, and SP1-SD-0007 and SP1-SD-0008.

1994 and 1995 Interim Action. Twenty-one sediment samples were collected and analyzed for TAL metals in 1994 and two sediments in 1995 from the locations depicted in Figures 2-5 and 2-6, respectively. Metal analytical results for the 1994 sediment samples indicated concentration above the method detection limit for each of the TAL metals except for cadmium, cobalt, nickel potassium, thallium, and vanadium. The metals arsenic, barium, chromium, lead, mercury, and zinc were detected at higher concentrations in sediment samples than soil samples. Arsenic concentrations ranged from 2 mg/kg: to 34 mg/kg. Six of the 1994 sediment samples exceed the soil CAL of 10 mg/kg established for Homestead ARB. Of those 6 sediments, 5 sample locations were subsequently excavated. Arsenic concentrations from the two post excavation samples are 25 mg/kg (CS-02) and 22 mg/kg (CS-03).

The ranges of concentrations for the remaining elevated metal compounds were 20.9 mg/kg to 5,290 mg/kg for barium; 44.3 mg/kg to 116 mg/kg for chromium; 157 mg/kg to 1210 mg/kg for lead; 0.13 mg/kg to 4.2 mg/kg for mercury; and 152 mg/kg to 954 mg/kg for zinc. Sediment concentrations exceeded CAL in only one sample for barium and all samples for lead. However, only one sample exceeded the State of Florida Health-Based Soil Target Level for lead. This sample (EWA-10) was subsequently excavated. With the exception of arsenic and lead, 1995 post excavation sediments were below the CAL and State of Florida Health-Based Soil Target Levels. Sediments analytical results for the 1994 Confirmation Sampling and 1995 Interim Action are summarized in Table 2-11.

2.6.4.5 Summary for Sediment. The two chlorinated VOCs detected at low concentrations in 1991 were probably associated with an adjacent equipment storage area. They were not detected in 1993. However, acetone and methylene chloride were detected, but are related to the field decontamination solvent and laboratory contamination, respectively.

Phthalates were detected in 1991 and 1993. These compounds are commonly observed when water has come in contact with plastics. No criteria are proposed for phthalates in sediment.

PAHs were detected in sediments. The most likely source of PAHs detected in the 1991 through 1995 sediment samples is run-off from the roadways and asphalt parking lots which are part of OU-5/Site WP-1. Also, in 1993 a leak from a diesel generator was observed following into one of the drainage swabs. PAHs and phthalates are not associated with electroplating wastes.

The metals detected in 1991 were comparable to the Boundary Canal background sample with the exception of mercury and zinc. Metals detected in 1993 were significantly higher then in 1991. Similarly, the metals arsenic, barium, chromium, lead, mercury, and zinc were detected at higher concentrations in 1994 and 1995 than in 1993. Arsenic and lead were the only metals detected which exceed CAL or State of Florida Health-Based Soil Target Levels. Cyanide was not detected in any of the sediment samples collected in 1991, 1993, or 1995.

Pesticides and PCBs were analyzed during the 1993 investigation only. PCB 1260 was detected at very low levels at one location. DDT, DDT metabolites, and chlordane were detected in sediment samples collected from 1994 through 1995. Pesticides have been observed in soil and sediment samples throughout Homestead ARB.

The significant and potential human health and environmental impacts of occurrences of constituents detected in drainage ditch sediments and surface water have been fully evaluated in the *Final OU-9 Remedial Investigation Report* (Woodward - Clyde, November 1995).

2.6.4 Surface Water Investigation

Due to the presence of surface water in the OU-5/Site WP-1 drainage ditches observed only during periods of heavy rain, surface water samples were only collected during the 1991 Investigation. In addition, surface water impacts of the Base ditches and canals have been evaluated further in the OU-9 Boundary Canal RI/RA. During the 1991 investigation, two surface-water samples were collected at the same locations as the sediment samples (Figure 2-3). Table 2-8 provides a summary of the compounds detected in the surface waters collected in 1991.

1991 Investigation. Two VOCs, acetone and methylene chloride, were detected in surface-water samples below the CRQL.

Several BNAs, mainly PAHs, were detected in the surface-water samples. Benzo(b)fluoranthene was detected in samples SP1-SW-0006 and SP1-SW-9005 at

concentrations of 22 and 2.0 μ g/L, respectively. Pyrene was detected in samples SP1-SW-0005, SP1-SW-90005, and SP1-SW-0006 at concentrations of 1.8, 2.4, and 27 μ g/L. Anthracene, benzo(a)anthracene, benzo(a)pyrene, benzo(k)fluoranthene, chrysene, and phenanthrene were detected in sample SP1-SW-0006 at concentrations of 0.96, 15, 14, 17, 26, and 6.8 μ g/l, respectively. Fluoranthene was detected in samples SP1-SW-0005, SP1-SW-90005, and SP1-SW-0006 at concentrations of 3.6, 3.6, and 41 μ g/L. PAHs are not commonly associated with electroplating operations but are present in waste automobile oils and fuels. The runoff from the equipment storage area and motor pool area or from the road would be a likely source of these contaminants.

Although the unlined drainage swale contains surface water intermittently, only during and after rain storms, the Class III Florida Surface-water Quality Standards for recreation and fish and wildlife were used to be consistent with investigations conducted at other Homestead ARB sites. TAL metals detected in the surface-water samples includes aluminum, barium, cadmium, calcium, chromium, copper, iron, magnesium, manganese, potassium, silver, sodium, vanadium, zinc, mercury, and lead. Aluminum, barium, calcium, magnesium, manganese, potassium, sodium, and vanadium are nutrients and there are no surface-water quality standards available for these constituents. The cadmium, copper, iron, silver, and mercury concentrations detected in sample SP1-SW-0006 exceeded their respective Florida Surface-water Quality Standards and the Federal Water Quality Criterion for these constituents. The concentrations of lead and zinc detected in samples SP1-SW-0005, SP1-SW-9005, and SP1-SW-0006 exceeded the Florida Surface-water Quality Criterion. Additionally, the detection limit for cyanide exceeded the Florida Surface-water Quality Standard and Federal Water Quality Criterion of 5 and 5.2 μ g/L, respectively.

2.6.4.1 Summary for Surface Water. Several PAHs were detected at low levels in surface water during the 1991 investigations. These concentrations are most likely the result of dissolution from the sediment when surface water is present primarily during the wet season or from surface runoff during frequent rainfall. The VOCs detected at very low levels are most probably laboratory or field contaminants.

Zinc, lead and cyanide were the only inorganics detected which exceeded water quality standards and/or criterion. The source of these constituents is most likely local runoff associated with frequent rainfall.

The significant and potential human health and environmental impacts of occurrences of constituents detected in drainage ditch sediments and surface water will be fully evaluated in the *Final OU-9 Remedial Investigation Report* (Woodward - Clyde, November 1995).

2.6.6 **Potential Routes of Migration**

Contaminants may migrate from a source area through a variety of processes. Volatile contaminants may be released into air and migrate in the vapor phase. Liquid or aqueous-phase contaminants may migrate to both soils and groundwater through direct infiltration. Erosion related to surface runoff or wind may transport contaminants sorbed to surface soils. Infiltrating precipitation may dissolve contaminants and carry them into deeper soils where they can be adsorbed, or into groundwater in the dissolved phase. Dissolved phase contaminants may be carried in the down gradient direction by groundwater flow in an aquifer.

Although other contaminated media are present at OU-5/Site WP-1, the principal route of migration of contaminants is through shallow groundwater. The impacts associated with the surface water and sediment samples have been further evaluated in the OU-9, Boundary Canal RI/RA. Past activities allowed contaminants to enter soil and surface water, which eventually migrated to shallow groundwater. Migration of contaminants via surface water occurs intermittently, during storm events.

OU-5/Site WP-1 and its drainages are situated on a developed portion of the Base which includes buildings, roads, and parking areas. The cycle of water through the site begins with precipitation. During rainfall events, water percolates rapidly through the limestone and weathered limestone bedrock underlying the site. Surface water runoff is over land to one of the drainage swales or ditches located in the immediate area of the site. The drainage swales and canals provide adequate surface water drainage for this site and are typically dry during non-storm events. Given the highly transmissive underlying formation, rainwater and surface water will typically infiltrate rapidly into the shallow aquifer system. It is estimated that horizontal groundwater movement can be on the order of tens of feet during a single rainfall event. Once the rainfall ceases, the water table returns to near static conditions and groundwater movement decreases dramatically.

Between rainfall events, evaporation from the surface soils returns water from the aquifer to the atmosphere. The rate of loss is greatest with open water bodies and decreases with increasing distance from the water table.

The natural concentrations of chemicals in the soil, rock, and water have a controlling effect on the fate and transport mechanisms. Soils at the site exist primarily as a veneer on the bedrock surface. A considerable amount of the OU-5/Site WP-1 area is covered by asphalt, roads, or buildings. The soil has both organic and iron precipitants. Nevertheless the calcium carbonate from the underlying oölite is the primary mineral present. The site drainage swales also receive runoff from the asphalted parking area located east of Building 164.

2.6.7 Exposure Assessment

This section of the risk assessment identifies and describes potential human receptors, reviews possible pathways of exposure for compounds of concern at OU-5/Site WP-1, and presents estimates of exposure doses resulting from identified pathways at OU-5/Site WP-1. An exposure assessment is conducted to identify potential sources and mechanisms of release, transport pathways (e.g., groundwater, surface water, soil, and air), routes of exposures (ingestion, inhalation, dermal contact), and potential on-site and off-site receptor populations (current users of the site, as well as adjacent populations which may be exposed to chemicals that have been transported off-site). This information provides the basis for constructing site-specific exposure scenarios.

Two environmental media were considered in this document - groundwater and surface soil. It should be noted that guidance on what depth range should be used for surface soil differs between the USEPA (0 to 12 inches) and the Florida DEP (0 to 24 inches). Samples taken between 0 and 24 inches below land surface (bls) were considered surface soil samples, so receptor exposure during gardening or landscaping activities could be evaluated in this assessment. This choice seems reasonable for south Florida, as the year-round, mild climate would permit possible residential gardening and frequent landscaping activities on base. No subsurface soil sampling was conducted because most soil layers at OU-5/Site WP-1 are only one to two inches deep and the underlying layers are composed of limestone and bedrock. Furthermore, the sediment and surface water samples collected at OU-5/Site WP-1 from the area canals and drainage ditches are not evaluated in this document. The potential human health effects due to exposures associated with the canal system are addressed in the BRA for OU-9, Boundary Canal Evaluation, which will be submitted as a separate report.

Other information considered in the development of present and future exposure scenarios includes: physical characteristics of the site and surrounding area such as climatology,

groundwater hydrology, location and description of surface water and surrounding land use and available state-specific guidelines relevant to exposure and risk assessments.

A critical step in assessing the potential risk to public health is to identify the pathways through which exposure could occur. A typical transport pathway consists of four necessary elements: 1) a source and mechanism of chemical release, 2) an environmental transport medium, 3) a point of potential contact with the contaminated medium, and 4) an exposure route (inhalation of vapors, ingestion of groundwater, etc.). All four of these elements must be present for a pathway to be complete.

Exposure Point Concentration. In accordance with USEPA methodology (1989a), the medium-specific 95 percent UCL on the arithmetic mean concentrations for the COPCs will be used as exposure point concentrations (EPCs) to estimate reasonable maximum exposure (RME). The RME approach is suggested by the USEPA (1989a) to provide an estimate of the maximum exposure (and therefore risk) that might occur. The RME corresponds to a duration and frequency of exposure greater than is expected to occur on an average basis. In those instances where the calculated 95 percent UCL exceeds the maximum detected concentration, the maximum detected concentration was used as the EPC for a more accurate estimate of RME concentration (USEPA, 1999a).

The total number of samples collected, as well as the sources of the data used in the risk assessment and included in the database for the calculation of each COPC exposure point concentration, varied by medium.

Once the database for each medium was developed, the 95 percent UCL concentration on the arithmetic mean concentration (one-tailed test, assuming a lognormal distribution) was calculated and compared to the maximum COPC concentration to determine the EPC for each COPC. The results of these analyses for the sampled media are presented in Tables 2-13 and 2-14. The information presented in these tables is discussed in the following subsections. An example of the data reduction used to calculate the arithmetic mean and UCL for each COPC is shown in Table 2-15.

Exposure Scenarios. Exposure pathways identified at OU-5/Site WP-1 are shown in Table 2-16 and are associated with soils or groundwater. With the exception of the VOCs, the chemicals detected at the site have low environmental mobility.

EXPOSURE POINT CONCENTRATIONS FOR GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

	Number of Samples in Groundwater Database ¹			tabase ¹		Maximum		Value Used
Constituent	Geraghty & Miller 1991	Mongomery Watson 1993	OHM 1996	Total Number Samples Averaged 1991-93	UCL Concentration ² (Fg/l)	Detected Concentration (Fg/l)	Mean Concentration (Fg/l)	in Risk Calculation s ³
VOCs								
Bromodichloromethane	4	1	4	9	4.8	2	3.6	3.6
Chloroform	4	1	4	9	4.8	2	3.6	3.6
Methylene Chloride	4	1	4	9	5.5	7	3.3	3.3
BNAs								
Bis(2-Ethylhexyl)phthalate	4	1	4	9	1500	320	38.5	38.5
METALs								
Aluminum	4	1	4	9	202,228,118	24000	4266.3	4266.3
Arsenic	4	1	4	9	83.2	92	26.3	26.3
Barium	4	1	4	9	203.7	150	37.2	37.2
Chromium	4	1	4	9	486.7	130	29.2	29.2
Lead	4	1	4	9	259.8	30	7.3	7.3
Manganese	4	1	4	9	170133.5	200	44	44
Nickel	4	1	4	9	42.3	8	10.8	10.8
Vanadium	4	1	4	9	151.7	82	16.4	16.4

Fg/L micrograms per Liter ¹ Count does not include duplicates; a duplicate sample was collected atone well (SP1-1-02) for each sampling program.

² The UCL concentration was calculated assuming a lognormal distribution of the data.

As per Region IV Guidance, the arithmetic mean of the wells in the plume (assumed to be the entire site) was used as the exposure point concentration. 3

EXPOSURE POINT CONCENTRATIONS IN SURFACE SOIL SAMPLES SITE WP-1/OUT-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Chemical	Geraghty & Miller 1991 ¹	Montgomery Watson 1993 ²	IT Corp 1994 ³	IT Corp 1995 ⁴	Total No. of Samples	UCL Concentration ⁵	Maximum Detected Concentration	Value Used in Risk Calculations ⁶
BNAs (Fg/kg)								
Benzo(a)pyrene	4	1	5	1	11	2.05E+06	460	460
Pesticides (Fg/kg)								
4,4'-DDD	0	4	3	1	8	87989	1,400	1,400
Chlordane Isomers	0	4	3	1	8	1.47E+06	1,400	1,400
TRPS (TICs) (mg/kg)	0	1	0	0	1		3,322	3,322
Metals (mg/kg)								
Aluminum	4	1	5	1	11	6,453	9,270	6,453
Arsenic	4	1	5	1	11	7.8	9.7	7.8
Cadmium	4	1	5	1	11	26.4	1.4	1.4
Chromium	4	1	5	1	11	18	23.4	18.0
Copper	4	1	5	1	11	226	160	160
Lead	4	1	5	1	11	543	120	120
Manganese	4	1	5	1	11	184	230	184
Mercury	4	1	5	1	11	0.15	0.4	0.2
Nickel	4	1	5	1	11	509	300	300
Vanadium	4	1	5	1	11	12.4	14.7	12

Micropams per kilogram Fg/kg

Miligram per kilograrn mg/kg

Not calculated due to low number of samples. ---

When a location was sampled in duplicate, the data is combined for risk assessment and is reported as one sample collected.

⁽¹⁾ Geraghty & Miller, 1991 Data Points: SP1-SL-0001, SP1-SL-0002, SP1-SL-0003, SP1-MW-0001-S.
 ⁽²⁾ Montgomery Watson 1993 Data Points: SP1-SL-0004, SP1-SL-0005, SP1-SL-0007.

⁽³⁾ IT Corporation 1994 Data Points; EWA1, EWA2, EWA3, EWA4, EWA5.

⁽⁴⁾ IT corporation 1995 Data Points: CS-01.

⁽⁵⁾ The UCL concentration was calculated assuming a lognormal distribution of the data.

(6) The UCL concentration is used as the exposure point concentration unless it is greater than the maximum detected concentration, in which case the maximum detected concentration is used.

(7) A UCL concentration was determined for lead in surface soil for use in the IEUBK model which was necessary since lead is a COPC in groundwater. In the dataset, 11 samples were collected,

the maximum concentration is 120 mg/kg, and the UCL is 543 mg/kg. Therefore the maximum concentration was used as the exposure point concentration (120 mg/kg).

EXAMPLE DATA REDUCTION CALCULATION FOR ARSENIC IN SURFACE SOIL SAMPLES AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Sample Designation	Analytical Result (mg/kg)	Value Used (1) (mg/kg)	Log Transformed Data
EWA1 (1 ft), 1994	0.43B	0.43	-0.844
EWA2 (1 ft), 1994	< 0.45	0.23	-1.492
EWA3 (1 ft), 1994	<1.5	0.75	-0.288
EWA4 (1 ft), 1994	0.56B	0.56	-0.580
EWA5 (1 ft), 1994	<0.41	0.21	-1.585
CS-01 (1 ft), 1995	6.9	6.9	1.932
SP1-SL-0007 (0-1 ft), 1993	9.7N	9.7	2.272
SP1-SL-0001 (0-1 ft), 1991	0.26	0.26	-1.347
SP1-SL-0002 (0-1 ft), 1991	0.61	0.61	-0.494
SP1-SL-0003 (0-1 ft), 1991	1.5	1.5	0.405
SP1-SL-MW-0001A (0-1 ft), 1991	0.83	0.83	-0.186

 $UCL = e^{\overline{x} + \left(0.5s^2 + \frac{sH}{\sqrt{n-1}}\right)}$

where:

Arithmetic mean of transformed data	$\bar{\chi} = -0.201$
Total number of samples	n = 11
Degrees of freedom	n - 1 = 10
Standard Deviation	s = 1.287
H-statistic of transformed data (∝=0.05)	H = 3.5
Upper Confidence Limit (in mg/kg)	UCL = 7.79

⁽¹⁾ All statistics were calculated using one-half the detection limit for non-detects, where applicable.

POTENTIAL PATHWAYS OF EXPOSURE TO CHEMICALS PRESENT IN SITE SAMPLES AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Medium	Pathway/Route	Potentially-Exposed Population	Comments
Groundwater (potable use)	Ingestion, dermal contact, and inhalation of constituents in groundwater.	None currently identified. Hypothetical future on-site residents unlikely due to nature and history of site.	No potable wells are located between Site WP-1 and the groundwater discharge point (drainage ditches or Boundary Canal). No active potable wells are located within a 1-mile radius of the site. Future potable use of groundwater is unlikely due to high total dissolved solids associated with salt-water intrusion.
Surface Soil	Incidental ingestion of and dermal contact with affected surface soils/dust and inhalation of affected dust.	Current base workers accessing site to cut the grass. Future construction workers excavating soils.	Most of the site is covered with sparse grass or gravel, so contact with soil, dust, or volatilized constituents is possible.
Surface Soil	Incidental ingestion of and dermal contact with affected surface soils/dust and inhalation of affected dust.	Hypothetical future residents could be exposed to constituents in soils.	The potential for future development of the site is limited due to the surrounding land use.

Under present conditions, access to the site is limited primarily to base workers performing duties that might require site access, such as cutting the grass. COPCs detected in the surface soils include one BNA, two pesticides, nine metals, and TRPH. Base workers cutting the grass at the site could be exposed to the soils via direct contact with exposed arms and face, incidental ingestion of soils that might adhere to the hands, and inhalation of dusts or vapors generated while cutting the grass. The site is sparsely vegetated, therefore, potential exposure rates will not be reduced by a vegetation factor.

If the operation of Homestead ARB is to continue in the future, exposure pathways at the site are unlikely to change. Given the location of OU-5/Site WP-1 in the midst of base administration activity, future development of this area for intensive base operations is unlikely. Foreseeable future use conditions at the site would result in potential exposure pathways similar to those discussed for present site conditions. However, it should be noted, that damage from the hurricane would necessitate construction activities regardless of future land use.

Under current reuse plans, OU-5/Site WP- 1 will be under cantonment of the US Air Force Reserve. Therefore, reuse of OU-5/Site WP-1 for residential purposes is unlikely in the foreseeable future. However, for risk characterization, hypothetical future residents were considered as receptors. Exposure pathways for hypothetical future residents would include direct contact with the surface soils, incidental ingestion of the surface soils, and inhalation of fugitive dust or vapors.

Hypothetical future construction workers were included in the risk characterization. Exposure pathways for future hypothetical construction workers would include ingestion and inhalation of soil. This receptor is evaluated for only surface soils (less than two feet) as no subsurface soil (greater than two feet) is expected at the site since most soil layers are usually one to two inches deep and the underlying layers are composed of limestone and bedrock.

Although it is unlikely that potable wells would be installed in the vicinity of the site, a conservative assumption made in this risk assessment is that a potable well is installed in the groundwater plume, downgradient of the site. Exposure of hypothetical future residents to affected groundwater via ingestion, inhalation, and dermal contact are considered potential exposure pathways.

In summary, workers cutting the grass on the site will be used to represent current exposure to the on-site soils. In the unforeseen event that the site is no longer under cantonment of the

US Air Force Reserve, hypothetical future exposure pathways considered included residential development of the site. Table 2-16 and Figure 2-9 summarize the potential exposure pathways for OU-5/Site WP-1.

2.7 SUMMARY OF SITE RISKS

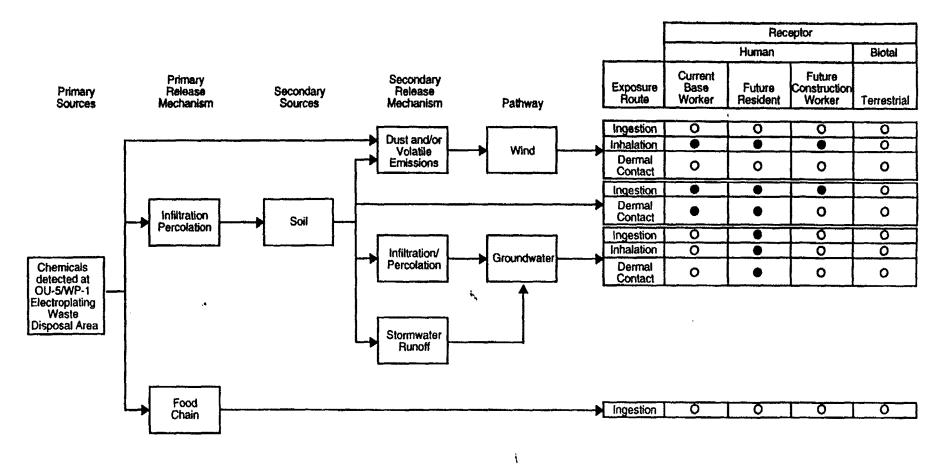
In order to evaluate whether existing or future exposure to contaminated media at OU-5/Site WP-1 could pose a risk to people or the environment, USAF completed a Baseline Risk Assessment (BRA) in October 1996 with USEPA oversight of this process. This evaluation then served as a baseline for determining whether cleanup of each site media was necessary. In the BRA, USAF evaluated site risks for several environmental media. This ROD addresses the risks attributable to chemicals in the soil and groundwater at OU-5/Site WP-1. Sediment and surface water will be addressed as part of OU-9, Boundary and Military Canal investigation. The risk assessment included the following major components: selection of chemicals of potential concern, exposure assessment, toxicity assessment, risk characterization, development of remedial goal options, ecological risk, and uncertainties. The USAF estimated potential site risk in the absence of any future remediation.

2.7.1 Selection of Chemicals of Potential Concern

This section presents an analysis of the site data to determine which chemicals present in site samples are potentially responsible for the greatest risks at the site. These chemicals are designated chemicals of potential concern (COPCs). The selection of COPCs allows the risk assessment to focus on a manageable list of the most important chemicals, which in turn permits concise analysis and presentation of information during the remainder of the risk assessment.

2.7.1.1 Criteria For Selection. The process of selecting the COPCs involves four criteria. These criteria are outlined in Figure 2-10. The first criterion involves determining whether a chemical is present within its range of natural background concentrations. Chemicals present at background levels are not selected as COPCs. Tables 2-17 and 2-18 present groundwater and soil background data, respectively.

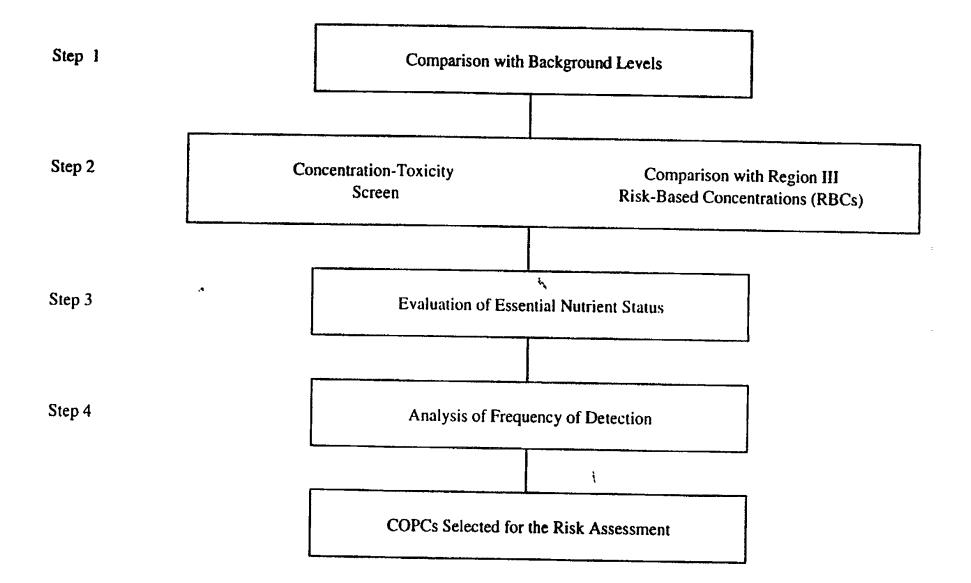
The second criterion is whether a chemical represents at least one percent of the risk in a given media, based on a screening method that involves the concentration and toxicity of the chemical. Factors other than concentration and toxicity are considered to potentially modify this criterion to include additional chemicals that account for less than one percent of the risk.



Pathway complete, further evaluation recommended

 $O \approx$ Pathway evaluated and found incomplete, no further evaluation recommended

Footnote: (1) Based on the non-volatile nature of the COPCs, it was considered that the potential for exposure to particulant, but not volatile emissions exists.



Identification of Chemicals of Potential Concern (COPCs): Criteria for Section

Figure 2-10

CONCENTRATIONS OF DISSOLVED INORGANIC CHEMICALS DETECTED IN THE BISCAYNE AQUIFER IN DADE COUNTY, FLORIDA Homestead Air Reserve Base, Florida

Constituent	Range (Fg/l)	Mean (F g/l)
Calcium	55,000 - 140,000	90,000
Magnesium	1,700 - 19,000	5,600
Sodium	7,400 - 77,000	26,600
Potassium	200 - 6,500	2,400
Chloride	13,000 - 110,000	42,000
Fluoride	100 - 500	200
Sulfate	100 - 45,000	14,600
Alkalinity (as CaCO3)	157,000 - 624,000	263,000
Arsenic	<1 - 2	1.2
Barium	<100 - 100	100
Cadmium	<1 - 3	1.0
Chromium ^a	<10 - 10	
Iron	<10 - 1,900	560
Lead	<1 - 6	1.9
Manganese	<10 - 30	9.7
Mercury	<0.1 - 0.3	0.10
Zinc	<10 - 30	7.5
TDS ^b	196 - 478	333
Hardness (as CaCO3) ^b	150 - 370	249

Source: Causaras, C.R., 1987, Geology of the Surficial Aquifer System, Dade County, Florida. U.S. Geological Survey Water Resources Investigation Report 86-4126.

Notes:

^a All detected observations had the same value.

^b In milligrams per liter

BACKGROUND SOIL CHEMICAL CONCENTRATIONS Homestead Air Reserve Base, Florida

Compound	Average Carbonate Composition Hem (1989)	Homestead ARB Background Soil ^{ton} 0-2 ft bls	Homestead ARB Background Soik ^a 4-6 ft bls	Typical Values for Uncontaminated Soils ⁴¹ (mg/kg)	Common Range" (mg/kg)	Average (mg/kg)	
Volatile Organic Compounds (µg/kg)		•					
Acesone		119.2	~	***	••		
Chlorobenzene		3.8	••			•-	
Methylene Chloride		4		~~	••	-	
Fotal PAHs (ug/kg)		738.55	-	0.01 - 1.3 forest " 0.01 - 1.01 rural 0.06 - 5.8 urban 8 - 336 road dust			
ase/Neutral and Acid Extractable Organ	ic Compounds (µg/kg/	dw)					
Acenaphthene		ND				**	
Benzo(a)anthracene		67					
Benzo(a)pyrene		66		+-		**	
Benzo(b)fluoranthene		69					
Benzo(g.h.i)perylene		44					
Benzo(k)fluoranthene		66					
bis(2-Ethylhexyl)phthalate		100	**				
Chrysene		79					
Dibenzofuran		ND					
Fluoranthene		52.4					
Fluorene		ND	-				
2-Methylnaphthalene		54					
Naphthalene		50				-	
Phenanthrene		50					
Рутеле		49.15		-			
1,2-Dichlorobenzene		ND		-		-	
1.4-Dichlorobenzene		NŐ		-	_		
otal Phthalates (µg/kg)		126	515				
fetals (mg/kg)							
Aluminum	8970	2400	425	-	700->10,000	57000	
Antimony		<28 - 30	<7.4 - <160	0 - 30	2 - 10"		
Arsenic	1.8	1.6	<1.4 - <2.9	0 - 30	<0.1 - 73	7.4	
Banum	30	42.9	5	0 - 500	10 - 1.500	420	
Beryllium	••	<2.8 - 2.9	<0.63 - <0.74	0-5	<1 - 7	0.85	
Cadmium	0.048	<2.8 - 3.0	<0.74 - <16	0 - 1	0.01 - 0.1*	0.06 ^m	
Calcium	272.000	345.000	400.000		10 - 28,000	630	
Chromium	>0.1	11.5	3.9	0 - 100	1 - 1.000	52	
Cobalt	0.12	<1.1 - 1.2	<1.3 - <1.5	7	⊲0.3 - 70	9.2	
Copper	4.4	<2.7 - 3.0	<3.2 - <3.7	30	<1 - 700	22	
Iron	8,190	1650	260		10 - 10,000	2,500	
Lead	16	4.05	1.4	0 - 500	<10 - 300	17	
Magnesium	45,300	1050	875	0 - 500	5 - 5,000	460	
Manganese	842	23	5.4	0 - 500	-2 - 7,000	640	
Mercury	0.046	0.014	<0.013 - <0.014	0 - 1	≪0.01 - 3.4	0.12	
Nickel	13	<4.5 - 4.7	5.1 - 5.9	15	<5 - 700	18	
Potassium	2,390	<110 - 120	<130 - <150	-	5 - 3,700		
Scienium		<5.6 - 5.7	<2.9 - <7.1	0 - 1	<0.01 - 3.9	0.45	
Silver	-	<1.1 - 1.2	<1.3 - <1.5	0.15	0.01 - 5.0*	0.05**	
Sodium	398	555	910	-	<500 - 50,000	7,800	
Thallium		<1.1 - 5.6	<1.3 . <6.8		2.2 - 23	8.6	
Vanadium	13	<5.7 - 5.9	2.3	0 - 100	<7 - 300	66	
		~	4	0 - 100		~~	

(a) Source: Based on 5 background samples as reported in Geraghty & Millet, 1992.

(b) Source: Based on 2 background samples as reported in Geraghty & Miller, 1992.

(c) Source: Gas Research Institute, 1987.

(d) U.S. Geological Survey Professional Paper 1270, Element Concentrations in Suils and Other Surficial Material of the Conterminous

United States Page 4, Table 1 (unless indicated otherwise).

(e) Source: Menzie, et al., 1992.

(f) Data for these metals were not included in the USGS Paper. Concentrations were obtained from the USEPA

Office of Solid Waste and Emergency Response, Hazardous Waste Land Treatment, SW-874, April 1983. Page 273, Table 6.45

(g) Average not established.

These factors include physical and chemical properties of a given chemical, environmental persistence, medium-specific mobility, the potential to bioaccumulate, potential routes of exposure, the spatial extent of the chemical, and the range and magnitude of concentrations detected. This screening method is consistent with toxicity screening guidance available during the preparation of previous drafts of this document.

Changes in COPC screening guidance have occurred. At the request of regulators, this change in guidance was incorporated into this document by screening chemicals detected in site samples using an additional method based on USEPA Region III Risk-Based Concentrations (RBCs). This additional screening is further discussed in Section 2.7.1.4.

The third criterion is whether a chemical is an essential human nutrient that is only toxic at very high doses (i.e., at doses that are both much higher than beneficial levels and much higher than could be associated with contact at the site). Chemicals typically considered under this criterion include calcium, iron, magnesium, potassium, and sodium.

The fourth criterion is to determine frequency of detection in a given medium. When chemicals are detected in less than five percent of the site samples for a given medium, they are not selected as a chemical of potential concern. This criterion was only used when at least 20 site samples had been collected: for a particular medium. The following paragraphs discuss the four criteria above in greater detail.

Background levels have been estimated for groundwater, surface soil, and subsurface soil. As per Region IV risk assessment guidance (USEPA, 1992b), inorganic chemicals which have maximum detected concentrations less than twice the background concentration are considered to be present at background levels. Exceptions to this rule have been made for known human carcinogens such as arsenic and chromium (assumed to present in the hexavalent state, or Cr(VI)). For these metals, the maximum detected concentration has been required to be less than background to assume that the metal is present at background levels.

The results of COPC screening groundwater and surface soil are summarized in Tables 2-19, and 2-20, respectively.

Groundwater. For groundwater, United States Geological Survey (USGS) data on the Biscayne Aquifer have been used for comparison with site groundwater samples (Causarus, 1987). The USGS data are summarized in Table 2-17. While it is generally considered preferable to determine background concentrations with wells immediately upgradient of the

SUMMARY OF CHEMICALS PRESENT IN SITE SAMPLES - GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

(Page 1 of 2)

Constituent	Minimum Detected Concentration (µg/l)	Maximum Detected Concentration (µg/l)	No. of Samples With Detects/ Total No. Samples	Preliminary Screening Summary
VOCs (µg/l)				
Bromodichloromethane	2	2	1/9	Included ¹
Chloroform	2	2	1/9	Included ¹
Methylene chloride	1	7	5/9	Included ¹
BNAs (µg/l)				
bis-(2-Ethylhexyl)phthalate	0.3	320	4/9	Included
Di-n-butyl phthalate	0.5	0.5	1/9	Excluded, low score ²
2-Methylnaphthalene	2	2	1/9	Excluded, low score ²
Naphthalene	1	1	1/9	Excluded, low score ²
Metals (µg/l)				
Aluminum	2610	24,000	5/9	Included
Arsenic	3.9	92	9/9	Included
Barium	4.5	150	9/9	Included
Calcium	10,100	5,400,000	9/9	Included, qualitative, high conc essential nutrient
Chromium	4.1	130	7/9	Included
Cabalt	3.5	3.5	1/9	Excluded, low score 2
Copper	3.7	11.9	3/9	Excluded, low score 2
Iron	4.9	18,000	7/9	Included, qualitative, high conc essential nutrient
Lead	5.4	30	5/9	Included

SUMMARY OFCHEMICALS PRESENT IN SITE SAMPLES - GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

(Page 2 of 2)

Constituent	Minimum Detected Concentration (µg/l)	Maximum Detected Concentration (µg/l)	No. of Samples With Detects/ Total No. Samples	Preliminary Screening Summary
Metals (µg/l) (cont	inued)			
Magnesium	1490	14,000	9/9	Excluded, essential nutrient, below site background
Manganese	0.81	200	7/9	Included
Nickel	8	8	1/9	Included
Potassium	642	4,950	9/9	Excluded, essential nutrient, below site background
Sodium	3,870	26,000	9/9	Excluded, essential nutrient, below site background
Vanadium	2.9	82	6/9	Included
Zinc	6.7	33.5	5/9	Excluded, low score ²

¹ Chemical was included as a COPC based on additional screening using benchmarks based on USEPA Region III Risk-Based Concentrations (See Section 2.6).

² Low score indicates <1% result for concentration-toxicity screen (USEPA, 1989) for the RfD and/or SF calculation (see Table 2-5).

SUMMARY OF CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida (Page 1 of 3)

Constituent	Minimum Detected Concentration	Maximum Detected Concentration	No. of Samples With Detects/ Total No. Samples	Preliminary Screening Summary
Volatile Organics (µg/kg)				
Acetone	1268	27,000	3/5	Excluded, low score ¹
BNAs (µg/kg)				
Acenaphthene	17	17	1/11	Excluded, low score ¹
Anthracene	39	39	1/11	Excluded, low score ¹
Benzo(a)Anthracene	25	691	3/11	Excluded, low score ¹
Benzo(a)Pyrene	30	460	3/11	Included ²
Benzo(b)Fluoranthene	37	540	3/11	Excluded, low score ¹
Benzo(g,h,i)Perylene	8.1	240	3/11	Excluded, low score ¹
Benzo(k)Fluoranthene	44	460	3/11	Excluded, low score ¹
Bis(2-Ethylhexyl)Phthalate	93	110	2/11	Excluded, low score ¹
Butyl Benzyl Phthalate	24	24	1/11	Excluded, low score ¹
Carbazole	62	62	1/5	Excluded, low score ¹
Chrysene	46	540	3/11	Excluded, low score ¹
Di-n-Butyl Phthalate	17	460	2/11	Excluded, low score ¹
Di-n-Octyl Phthalate	7	7	1/11	Excluded, low score ¹
Dibenz(a,h)Anthracene	87	87	1/11	Excluded, low score ¹
Dibenzofuran	11	11	1/11	Excluded, low score ¹
Fluoranthene	55	1,200	3/11	Excluded, low score ¹
Fluorene	20	20	1/11	Excluded, low score ¹
Indeno(1,2,3-c,d)Pyrene	230	230	1/11	Excluded, low score ¹
Phenanthrene	10	490	3/11	Excluded, low score ¹
Pyrene	58	810	3/11	Excluded, low score ¹
TRPHs (TICs) (mg/kg)	3322	3322	1/1	Included

SUMMARY OF CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL SITE 2–1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida (Page 2 of 3)

Constituent	Minimum Detected Concentration	Maximum Detected Concentration	No. of Samples With Detects/ Total No. Samples	Preliminary Screening Summary
Pesticides/PCBs (µg/kg)				
4,4'-DDD	1.4	1,400	5/8	Included
4,4'-DDE	3.9	240	7/8	Excluded, low score 1
4,4'-DDT	0.83	34	5/8	Excluded, low score 1
Chlordane	4.2	1,400	4/8	Included
Endosulfan sulfate	0.77	8.8	2/8	Excluded, low score 1
Metals (mg/kg)				
Aluminum	728	9,270	11/11	Included
Arsenic	0.26	9.7	8/11	Included
Barium	43.6	21.2	11/11	Excluded, low score 1
Beryllium	0.31	0.31	1/11	Excluded, low score 1.4
Cadmium	0.42	1.4	2/11	Included
Calcium	30,600	646,716	11/11	Excluded, essential nutrient, below site background
Chromium	7.2	23.4	6/11	Included
Cobalt	0.34	1.4	4/11	Excluded, low score ¹
Copper	1.7	160	7/11	Included
Iron	302	5,800	1/11	Included, qualitative, high cone essential nutrient
Lead	2.4	120	10/11	Exluded, USEPA ⁵
IEUBK Magnesium	970	1,500	11/11	Excluded, essential nutrient, below site background
Manganese	5.1	230	11/11	Included
Mercury	0.011	0.4	2/11	Included
Nickel	8.87	300	6/11	Included
Potassium	28	2,160	7/11	Included, qualitative, high cone essential nutrient
Sodium	212	513	9/11	Excluded, essential nutrient, below site backgroun

SUMMARY OF CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

(Page 3 of 3)

Constituent	Minimum Detected Concentration	Maximum Detected Concentration	No. of Samples With Detects/ Total No. Samples	Preliminary Screening Summary
Meatal (mg/kg) (continu				
Vanadium	8.2	14.7	6/11	Included
Zinc	2.2	300	9/11	Excluded, low score ¹

¹ Low score indicates <1% result for toxicity-concentration screen (USEPA, 1989) for the RfD and/or SF calculation (Refer to Table 2-6).

² Chemical was included as a COPC, despite low score, because the maximum detected concentration exceeded the benchmark based on USEPA Region III Risk-Based Concentration (RBC) (see Section 2.6).

³ Included as a COPC, despite low score, because there is no Region III RBC available for n-nonane (see Section 2.6).

⁴ Beryllium exceeded benchmark based on USEPA Region III RBCs (see Section 2.6), but did not exceed site background.

⁵ This concentration of lead in soil has been shown by the USEPA IEUBK model to result in no significant additional uptake of lead in blood for children or adults.

site, the monitoring well SP1-I-01 designated by Geraghty & Miller as a background well in 1991 had concentrations of several metals which indicate it is not representative of background conditions. Therefore, this well has subsequently been included in the risk assessment dataset for groundwater. The USGS data for the Biscayne Aquifer are more likely to represent undisturbed groundwater conditions and have been used to place the site data in perspective.

Soil. For surface soil, five Base-wide background samples were collected by Geraghty & Miller in 1991. These samples include SP1-SL-0028-2, P3-SL-0023, P2-SL-0023-2, SP3-SL-0004-1, and SP3-SL-0004-2. For subsurface soil, two background samples (SP11-SL-0028-6 and SP7-SL-0002) were collected. Soil background values are summarized in Table 2-18. In addition, data concerning typical chemical concentration ranges in soil arc used to place the site data in perspective (Shacklette and Boerngen, 1984).

2.7.1.2 Concentration-Toxicity Screen. The concentration-toxicity screen is used to calculate indices that rank the chemicals according to their relative potentials to create health risks at the site. One index is used to rank chemicals according to their potential for initiating or promoting cancers, and a second index ranks chemicals-according to their potential for chronic non-cancer effects. The first index applies only to carcinogens, while the latter index applies to noncarcinogens. These indices used for ranking purposes only, and do not represent actual risk values.

The index used for ranking carcinogens involves the use of a cancer slope factor (CSF). Studies of carcinogenicity tend to focus on identifying the slope of the linear portion of a curve of dose versus response. A plausible upper-bound value of the slope is called the slope factor.

The index used to rank chemicals according to their potential to cause noncarcinogenic effects involves the use of a reference dose (RfD). A chronic RfD is an estimate of a daily exposure level for which people, including sensitive populations, do not have an appreciable risk of suffering significant adverse health effects. Most CSFs and RfDs were obtained from the *Integrated Risk Information System* (IRIS), or, if not available there, from the *Health Effects Assessment Summary Tables* (HEAST).

The index for carcinogenic effects is calculated by taking the maximum detected concentration of each contaminant and multiplying by the oral slope factor. The inhalation CSF is used for chemicals that are only carcinogenic by inhalation (chromium and cadmium).

The index for noncarcinogenic effects is calculated by taking the maximum detected concentration of each contaminant and dividing by the oral RfD. Chemicals making up at least one percent of the total index for all chemicals have been selected as COPCs (unless the chemical has been eliminated based on background or essential nutrient considerations). Concentration toxicity screening results for groundwater and surface soils are presented in Tables 2-21 and 2-22, respectively.

Due to changes in guidance during the development of this document, an additional toxicityscreening method, based on USEPA Region III RBCs, was also used to screen for COPCs. This method is described in Section 2.7.1.4.

2.7.1.3 Data Analysis. This subsection is organized according to media (groundwater and surface soil). Within each medium, the data are presented in the order of volatile organic compounds (VOCs), semi-volatile organic compounds, pesticides/polychlorinated biphenyls (PCBs), and inorganics. Comparisons are made to the four criteria listed in Section 2.7.1.1, and then COPCs are selected. The summary Tables 2-19 and 2-20 present chemical concentration ranges, frequencies of detection, and whether a chemical has been selected as a COPC.

The analytical data for this risk assessment were collected by Geraghty & Miller during an investigation in 1991, Montgomery Watson during 1993, IT Corporation in 1994 and 1995 (soil sampling only), and OHM in 1996 (groundwater sampling only). An in-depth discussion of the sample collection and analytical methodology is presented in Section 2.0 of the *Final Remedial Investigation Report for Operable Unit 5/Site WP-1, Electroplating Waste Disposal Area (Former Site SP-1)* (Montgomery Watson, 1996).

The soil and groundwater analytical data were reduced and analyzed for use in the risk assessment according to guidelines provided by USEPA (1989a, 1991a). Geraghty & Miller, IT Corporation, and OHM performed laboratory analyses and data validation for their field samples; Montgomery Watson performed its own data validation, which is reported in a *Draft Quality Control Summary Report*, while Savannah Laboratories performed the laboratory analyses. All data collected by Geraghty & Miller in 1991, Montgomery Watson in 1993, IT Corporation in 1994 and 1995, and OHM in 1996 were reviewed for this risk evaluation, This includes a review of detects, detection limits for non-detects, and estimated (J-qualified) data. Detection limits reported for Montgomery Watson samples were in compliance with CLP SOW contract required quantitation limits (CRQL). However, sample

TOXICITY - CONCENTRATION SCREEN FOR CHEMICALS PRESENT IN SITE SAMPLES - GROUNDWATER WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homstead Air Reserve Base, Florida

Constituent	Maximum Concentration mg/l			Slope Factor (mg/kg/day) ¹		Non-Carcinogen Index (conc/RfD)	Carcinogen Index (conc x SF)	% RfD	% SF	
VOCs										
Bromodichloromethane	0.002	2.0E-02	(a)	6.2E-02	(a)	1.0E-01	1.2E-04	0.03%	0.00%	
Chloroform	0.002	1.0E-02	(a)	6.1E-03	(a)	2.0E-01	1.2E-05	0.05%	0.00%	
Methylene chloride	0.007	6.0E-02	(e)	7.5E-03	(e)	1.2E-01	5.3E-05	0.03%	0.00%	
BNAs										
Bis(2-Ethylhexyl)phthalate	0.32	2.0E-02	(a)	1.4E-02	(a)	1.6E+01	4.5E-03	4.04%	0.08%	
Di-n-Butyl Phthalate	0.0005	1.0E-01	(a)	NA		5.0E-03	NC	0.00%	NC	
2-Methylnaphthalene ⁽¹⁾	0.002	4.0E-02	(c)	NA		5.0E-02	NC	0.01%	NC	
Naphthalene	0.001	4.0E-02	(c)	NA		2.5E-02	NC	0.01%	NC	
Metals										
Aluminum	24	1.0E+00	(c)	NA		2.4E+01	NC	6.06%	NC	
Arsenic	0.092	3.0E-04	(a)	1.5E+00	(a)	3.1E+02	1.4E-01	77.39%	2.46%	
Barium	0.15	7.0E-02	(a)	NA		2.1E+00	NC	0.54%	NC	
Calcium	5400	NA		NA		NC	NC	NC	NC	
Chromium ⁽²⁾	0.13	5.0E-03	(a)	4.2E+01	(a)	2.6E+01	5.5E+00	6.56%	97.34%	
Cabalt	0.0035	6.0E-02	(c)	NA		5.8E-02	NC	0.01%	NC	
Copper	0.0119	3.7E-02	(h)	NA		3.2E-01	NC	0.08%	NC	
Iron	18	NA		NA		NC	NC	NC	NC	
Lead	0.03	NA		NA		NC	NC	NC	NC	
Magnesium	14	NA		NA		NC	NC	NC	NC	
Manganese (3)	0.2	2.4E-02	(a)	NA		8.3E+00	NC	2.10%	NC	
Nickel ⁽⁴⁾	0.008	2.0E-02	(e)	8.4E-01	(a)	4.0E-01	6.7E-03	0.10%	0.12%	
Potassium	4.95	NA		NA		NC	NC	NC	NC	
Sodium	26	NA		NA		NC	NC	NC	NC	
Vanadium	0.082	7.0E-03	(h)	NA		1.2E+01	NC	2.96%	NC	
Zinc	0.0335	3.0E-01	(a)	NA		1.1E-01	NC	0.03%	NC	

Notes:

Toxicity values quoted in this table are for the oral pathway unless otherwise noted

 $\operatorname{conc} = \operatorname{concentration}$

NA = Not Available

NC = Non-Carcinogenic

RfD = Reference Dose

SF = Slope Factor

^(a) IRIS, 1995

^(b) HEAST, 1995

(c) ECAO

 $^{(1)}$ $\,$ Naphthalene RfD used as surrogate for 2-Methylnaphthalene, RfD $\,$

⁽²⁾ Slope factor is for inhalation pathway
 ⁽³⁾ RfD for manganese is calculated based on the NOAEL of 10 mg/day in food, using a modifying factor of 3 for non-dietary intake.

⁽⁴⁾ Nickel refinery dust inhalation slope factor used as surrogate for Nickel slope factor

TOXICITY - CONCENTRATION SCREEN FOR CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL (0-2 FT) SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

(Page 1 of 2)

Constituent	Maximum RfD Slope Factor tituent Concentration mg/kg/day mg/kg/day			Non-Carcinogen Index (conc/RfD)	Carcinogen Index (conc x SF)	% RfD	% SF		
VOCs									
Acetone	27	1.0E-01	(a)	NA		2.7E+02	NC	0.24%	NC
BNAs									
Acenaphthene	0.017	6.0E-02	(a)	NA		2.8E-01	NC	0.00%	NC
Anthracene	0.039	3.0E-01	(a)	NA		1.3E-01	NC	0.00%	NC
Benzo(a)anthracene ⁽¹⁾	0.691	3.0E-02	(a)	7.3E-01	(8) (a)	2.3E+01	5.0E-01	0.02%	0.04%
Benzo(a)pyrene ⁽¹⁾	0.46	3.0E-02	(a)	7.3E+00	(a)	1.5E+01	3.4E+00	0.01%	0.27%
Benzo(b)fluoranthene	0.54	3.0E-02	(a)	7.3E-01	(8) (a)	1.8E+01	3.9E-01	0.02%	0.03%
Benzo(g,h,i)perylene ⁽¹⁾	0.24	3.0E-02	(a)	NA		8.0E+00	NC	0.01%	NC
Benzo(k)fluoranthene ⁽¹⁾	0.46	3.0E-02	(a)	7.3E-02	(8) (a)	1.5E+01	3.4E-02	0.01%	0.00%
bis(2-Ethylhexyl)phthalate	0.11	2.0E-02	(a)	1.4E-02	(a)	5.5E+00	1.5E-03	0.00%	0.00%
Butylbenzyohthalate	0.024	2.0E-01	(a)	NA		1.2E-01	NC	0.00%	NC
Carbazole ⁽¹⁾	0.062	3.0E-02	(a)	2.0E-02	(b)	NC	1.2E-03	NC	0.00%
Chrysene ⁽¹⁾	0.54	3.0E-02	(a)	7.3E-03	(8) (a)	1.8E-01	3.9E-03	0.02%	0.00%
Di-n-Butyl Phthalate	0.46	1.0E-01	(a)	NA		4.6E+00	NC	0.00%	NC
Di-n-octyl phthalate	0.007	2.0E-02	(a)	NA		3.5E-01	NC	0.00%	NC
Dibenzo(a,h)anthracene ⁽¹⁾	0.087	3.0E-02	(a)	7.3E+00	(8) (a)	2.9E+00	6.4E-01	0.00%	0.05%
Dibenzofuran	0.011	4.0E-03	(a)	NA		2.8E+00	NC	0.00%	NC
Fluoranthene	1.2	4.0E-02	(a)	NA		3.0E+01	NC	0.03%	NC
Fluorene	0.02	4.0E-02	(a)	NA		5.0E-01	NC	0.00%	NC
Indeno(1,2,3-c,d)pyrene ⁽¹⁾	0.23	3.0E-02	(a)	7.3E-01	(8) (a)	7.7E+00	1.7E-01	0.01%	0.01%
Phenathrene ⁽¹⁾	0.49	3.0E-02	(a)	NA		1.6E+01	NC	0.01%	NC
Pyrene	0.81	3.0E-02	(a)	NA		2.7E+01	NC	0.02%	CN
TRPHs (TICs) (as n-Nonane) ⁽²⁾	3322	6.0E-01	(d)	NA		5.5E+03	NC	4.85%	NC
Pesticides/PCPs									
4,4'-DDD ⁽³⁾	1.4	5.0E-04	(a)	2.4E-01	(a)	2.8E+03	3.4E-01	2.44%	0.03%
4,4'-DDE ⁽³⁾	0.24	5.0E-04	(a)	3.4E-01	(a)	4.8E+02	8.2E-02	0.42%	0.01%
4,4'-DDT	0.034	5.0E-04	(a)	3.4E-01	(a)	6.7E+01	1.2E-02	0.06%	0.00%
Chlordane	1.4	6.0E-05	(a)	1.3E+00	(a)	2.3E+04	1.8E+00	20.46	0.15%
Endosulfan Sulfate ⁽⁴⁾	0.0088	6.0E-03	(a)	NA		1.5E+00	NC	% 0.00%	NC
<u>Metals</u>									
Aluminum	9270	1.0E+00	(C)	NA		9.3E+03	NC	8.13%	NC
Arsenic	9.7	3.0E-04	(a)	1.5E+00	(a)	3.2E+04	1.5E+01	28.34	1.17%
Barium	21.2	7.0E-02	(a)	NA		3.0E+02	NC	%	NC
Beryllium	0.31	5.0E-03	(a)	4.3E+00	(a)	6.2E+01	1.3E+00	0.27%	0.11%
Cadmium (food) ⁽⁵⁾	1.4	1.0E-03	(a)	6.3E+00	(a)	1.4E+03	8.8E+00	0.05%	0.71%
Calcium	646716	NA		NA		NC	NC	1.23% NC	NC

TOXICITY - CONCENTRATION SCREEN FOR CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL (0-2 FT) SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

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Constituent	Maximum Concentration mg/kg	RfD mg/kg/day	Slope Factor mg/kg/day	Non-Carcinogen Index (conc/RfD)	Carcinogen Index (conc x SF)	% RfD	% SF
Metals (continued)							
Chromium ⁽⁵⁾	2.34	5.0E-03 ^(a)	4.1E+01 ^(b)	4.7E+03	9.6E+02	4.10%	77.16%
Cobalt	1.4	6.0E-02 ^(c)	NA	2.3E+01	NC	0.02%	NC
Copper	160	3.7E-02 ^(c)	NA	4.3E+03	NC	3.79%	NC
Iron	5800	NA	NA	NC	NC	NC	NC
Lead	120	NA	NA	NC	NC	NC	NC
Magnesium	1500	NA	NA	NC	NC	NC	NC
Manganese (6)	230	2.4E-02 ^(a)	NA	9.6E+03	NC	8.40%	NC
Mercury	0.4	3.0E-04 ^(b)	NA	1.3E+03	NC	1.17%	NC
Nickel ⁽⁷⁾	300	2.0E-02 ^(a)	8.4E-01 ^(a)	1.5E+04	2.5E+02	13.15%	20.27%
Potassium	2,160	NA	NA	NC	NC	NC	NC
Sodium	513	NA	NA	NC	NC	NC	NC
Vanadioum	14.7	7.0E-03 ^(b)	NA	2.1E+03	NC	1.84%	NC
Zinc	300	3.0E-01 ^(a)	NA	1.0E+03	NC	0.88%	NC

NOTES:

Toxicity values quoted in this table are for the oral pathway unless otherwise noted

NA = Not Available

- NC = Not calculated
- RfD = Reference Dose
- SF = Slope Factor
- ^(a) IRIS, 1996
- ^(b) HEAST, 1995
- (c) ECAO

^(d) Massachusetts DEP, October 1994

⁽¹⁾ Pyrene RfD used as a surrogate for RfD of various PAHs

⁽²⁾ n-Nonane RfD used as surrogate for TRPHs RfD

⁽³⁾ DDT RfD used as surrogate for DDD and DDE RfDs

⁽⁴⁾ Endosulfan RfD used as surrogate for endosulfan sulfate RfD

⁽⁵⁾ Slope factor is for inhalation pathway

⁽⁶⁾ RfD for manganese is calculated based on the NOAEL of 10 mg/day in food, using a modifying factor of 3 for non-dietary intake.

⁽⁷⁾ Nickel refinery dust inhalation slope factor used as surrogate for Nickel slope factor

⁽⁸⁾ Toxicity equivalence factor (TEF) was applied to the benzo(a)pyrene slope factor, based on the relative potency of this chemical to be

quantitation limits (SQL) at levels suitably low for risk assessment use were not consistently achieved.

2.7.1.4 Screening Using Risk-Based Concentrations. Guidance on COPC selection changed during the development of this document. Therefore, a Risk-Based Concentration (RBC)-based benchmark screening method was added after input from regulators. Note that the use of both the toxicity-concentration screening method described in Section 2.7.1.2 and the RBC method described below results in a greater number of COPCs than use of each method singly. Therefore, selection of COPCs in this document is more conservative.

Risk-Based Concentrations. Current USEPA Region IV guidance recommends using the USEPA Region III RBCs as guidance for screening. RBCs are published periodically by USEPA Region III to act as guidance in risk management, risk assessment, and remediation decisions. RBCs are generated using default exposure parameters for chemicals in a specific media. Concentrations quoted in the USEPA Region III RBC Table that represent risk levels of 1E-06 (for carcinogens) or a hazard quotient of 1 (for non-carcinogens). USEPA Region IV suggests that screening values for non-carcinogenic chemicals be adjusted to represent a hazard quotient of 0.1.

Maximum concentration values of all chemicals detected in a particular environmental medium are compared to the appropriate RBCs in Tables 2-23 and 2-24. Chemicals whose maximum concentration exceeded the benchmark value were added as COPCs. The results of this process are summarized below.

Groundwater. Chemicals detected in groundwater were compared to the Tap Water RBCs. The results of this comparison are shown in Table 2-23. The comparison resulted in bromodichloromethane, chloroform, and methylene chloride being added to the list of COPCs for groundwater. All other chemicals that exceeded the RBCs had been already selected as COPCs, based on previous screening described in Sections 2.7.1.2 and 2.7.1.3, and Table 2-21.

Surface soil. Chemicals detected in surface soil were compared to RBCs for residential soil. The results of this comparison are shown in Table 2-24. The comparison resulted in benzo(a)pyrene being added to the list of COPCs for surface soils. Although the maximum concentration of beryllium in surface soil exceeded its RBCs, the concentrations detected were within background, so beryllium was not considered a COPC in surface soil.

RBC-BASED BENCHMARK SCREENING FOR CHEMICALS PRESENT IN SITE SAMPLES - GROUNDWATER SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

	COPC from previous Maximum		Toxicity Values			Œ	RCBs Region III	RBC-based	Exceeds Benchmark	
Constituent	screening? ^(b) (yes = +)	Concentration mg/l	RfD mg/kg/day		ope Factor ng/kg/day) ⁻⁴	ì	ap Water) mg/l	Benchmark mg/l	(yes = +)	COPC
VOCs		_								
Bromodichloromethane	-	0.002	2.0E-02	(a)	6.2E-02	(a)	0.00017	0.00017	+	+
Chloroform	-	0.002	1.0E-02	(a)	6.1E-03	(a)	0.00015	0.00015	+	+
Methylene Chloride	-	0.007	6.0E-02	(a)	7.5E-03	(a)	0.0041	0.0041	+	+
BNAs										
Bis(2-Ethylhexyl)phthalate	+	0.32	2.00E-02	(a)	1.40E-02	(a)	0.0048	0.0048	+	+
Di-n-Butyl Phthalate	-	0.0005	1.00E-01	(a)	NA		3.7	0.37	-	-
2-Methylnaphthlene ⁽²⁾	-	0.002	4.00E-02	(c)	NA		1.5	0.15	-	-
Naphthalene	-	0.001	4.00E-02	(c)	NA		1.5	0.15	-	-
Metals										
Aluminum	+	24	1.00E+00	(c)	NA	(a)	37	3.7	+	+
Arsenic	+	0.092	3.00E-04	(a)	1.50E+00	(a)	0.000045	0.000045	+	+
Barium	+	0.15	7.00E-02	(a)	NA	(a)	2.6	0.26	-	+
Chromium ⁽³⁾	+	0.13	5.00E-03	(c)	4.20E+01		0.18	0.18	-	+
Cobalt	-	0.0035	6.00E-02	(c)	NA		2.2	0.22	-	-
Copper	-	0.0119	3.70E-02	(b)	NA		1.5	0.15	-	-
Lead	+	0.03	NA		NA		NA	NA	NA	+
Manganese	+	0.2	2.40E-02	(a)	NA		0.18	0.018	+	+
Nickel ⁽⁴⁾	+	0.008	2.00E-02	(a)	8.40E-01	(a)	0.73	0.073	-	+
Vanadium	+	0.082	7.00E-03	(b)	NA		0.26	0.026	+	+
Zinc	-	0.0335	3.00E-01	(a)	NA		11	1.1	-	-

Notes:

s table are for the oral pathway unless otherwise noted.

Essential nutrients (calcium, iron, magnesium, potassium and sodium) are not considered in this table. See Sections 2.4 and 2.5 for full discussion of essential nutrients. Chemicals which did not have RBC values were carried as COPCs in the risk assessment.

NA = Not Available

RfD = Reference Dose

^(a) IRIS, 1996

(b) HEAST, 1995(c) ECAO

Lene

 $^{(1)}$ $\,$ Based on screening carried out in Table 2-5 and Section 2.5 $\,$

 $^{\scriptscriptstyle (2)}$ $\,$ N2phthalene RID used as surrogate for 2-Methylnaphthalent RfD $\,$

⁽³⁾ Slope factor is for inhalation pathway

(4) Nickel refinery dust inhalation slope factor used as surrogate for Nickel slope factor

RBC-BASED BENCHMARK SCREENING FOR CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL (0-2 FT) SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida (Page 1 of 2)

	COPC from previous	Maximum	Toxicit	y Values	RCBs		Exceeds		
Constituent	screening? ⁽¹⁾ (yes = +)	Concentration mg/l	RfD mg/kg/day	Slope Factor (mg/kg/day)	 (Region III Resid Soil) mg/kg 	RBC-based Benchmark mg/kg	Benchmark (yes = +)	COPC	
VOCs				214					
Acetone	-	27	1.0E-01 ^(a)	NA	7,800	780	-	-	
BNAs									
Acenaphthene	-	0.017	6.0E-02 ^(a)	NA	4,700	470	-	-	
Anthracene	-	0.039	3.0E-01 ^(a)	NA	23,000	2,300	-	-	
Benzo(a)anthracene ⁽²⁾	-	0.691	3.0E-02 ^(a)	7.3E-01 ^{(9) (a)}	0.88	0.88	-	-	
Benzo(a)pyrene ⁽²⁾	_	0.46	3.0E-02 ^(a)	7.3E+00 ^(a)	0.088	0.088	+	+	
Benzo(b)fluoranthene	_	.54	3.0E-02 ^(a)	7.3E-01 ^{(9) (a)}	0.88	0.88	-	-	
Benzo(g,h,i)perylene ⁽²⁾	_	0.24	3.0E-02 ^(a)	NA	2,300	230	_	_	
Benzo(k)fluoranthene ⁽²⁾	_	0.46	3.0E-02 ^(a)	7.3E-02 ^{(9)(a)}	8.8	8.8	-	-	
bis(2-Ethylhexyl)phthalate	-	0.40	2.0E-02 ^(a)	1.4E-02 ^(a)	46	46	-	-	
Butylbenzyohthalate	-	0.024	2.0E-02 (a)	NA	73	0.73	-	-	
Carbazole ⁽²⁾	-	0.024		2.0E-02 ^(b)	32	32	-	-	
Chrysene ⁽²⁾	-		$3.0E-02^{(a)}$	7.3E-03 ^{(9) (a)}	32 88		-	-	
	-	0.54	$3.0E-02^{(a)}$			88	-	-	
Di-n-Butyl Phthalate	-	0.46	$1.0E-01^{(a)}$	NA	7,800	780	-	-	
Di-n-octyl phthalate	-	0.007	$2.0E-02^{(a)}$	NA	1,600	160	-	-	
Dibenzo(a,h)anthracene ⁽²⁾	-	0.087	3.0E-02 ^(a)	7.3E+00 ^{(9)(a)}	0.088	0.088	-	-	
Dibenzofuran	-	0.011	4.0E-03 (a)	NA	310	31	-	-	
Fluoranthene	-	1.2	4.0E-02 ^(a)	NA	3,100	310	-	-	
Fluorene	-	0.02	4.0E-02 ^(a)	NA	3,100	310	-	-	
Indeno(1,2,3-c,d)pyrene ⁽²⁾	-	0.23	3.0E-02 ^(a)	7.3E-01 ^{(9) (a)}	0.88	0.88	-	-	
Phenathrene ⁽²⁾	-	0.49	3.0E-02 ^(a)	NA	2,300	230	-	-	
Pyrene	-	0.81	3.0E-02 ^(a)	NA	2,300	230	-	-	
TRPHs (TICs) (as n-Nonane) ⁽³⁾		3322	6.0E-01 ^(a)	NA	NA	NA	NA		
Pesticides/PCPs									
4,4'-DDD (4)	+	1.4	5.0E-04 ^(a)	2.4E-01 (a)	2.7	2.7	-	+	
4,4'-DDE ⁽⁴⁾	-	0.24	5.0E-04 ^(a)	3.4E-01 ^(a)	1.9	1.9	-	-	
4,4'-DDT	-	0.034	5.0E-04 ^(a)	3.4E-01 ^(a)	1.9	1.9	-	-	
Chlordane	+	1.4	6.0E-05 ^(a)	1.3E+00 (a)	0.49	0.49	+	+	
Endosulfan Sulfate ⁽⁵⁾	-	0.0088	6.0E-03 ^(a)	NA	47	47	-	-	
Metals									
Aluminum	+	9270	1.0E+00 (c)	NA	78,000	7,800	+	+	
Arsenic	+	9.7	3.0E-04 ^(a)	1.5E+00 (a)	0.43	0.43	+	+	
Barium	-	21.2	7.0E-02 ^(a)	NA	550	550	-	-	
Beryllium ⁽⁶⁾	-	0.31	5.0E-03 ^(a)	4.3E+00 (a)	0.15	0.15	+	+	
Cadmium (food) ⁽⁷⁾	+	1.4	1.0E-03 ^(a)	6.3E+00 ^(a)	3.9	3.9	-	+	
Chromium ⁽⁷⁾	+	23.4	5.0E-03 ^(a)	4.1E+01 ^(a)	39	39	-	+	
Cobalt	-	1.4	6.0E-02 ^(c)	NA	470	470	-	-	
Copper	+	1.4	3.7E-02 ^(b)	NA	310	310		+	

RBC-BASED BENCHMARK SCREENING FOR CHEMICALS PRESENT IN SITE SAMPLES - SURFACE SOIL (0-2 FT) SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida (Page 2 of 2)

Constituent	COPC from previous	Maximum	Toxicity Values				RCBs		Exceeds	
	screening? ⁽¹⁾ (yes = +)	Concentration mg/kg RfD mg/kg/day			Slope Factor (mg/kg/day)	(Region III Resid Soil) mg/kg		RBC-based Benchmark mg/kg	Benchmark (yes = +)	COPC
etals (continued)					NA		NA	NA	NA	-
Lead	-	120	NA	(a)	NA		390	39	+	+
Manganese	+	230	2.4E-02	(a)	NA		23	2.3	-	+
Mercury	+	0.4	3.0E-04	(a)	8.4E-01	(a)	1,600	160	+	+
Nickel ⁽⁸⁾	+	300	2.0E-02	(a)	NA		550	55	-	+
Vanadium	+	14.7	7.0E-03	(a)	NA		23,000	2,300	-	-
Zinc	-	300	3.0E-01	(a)						

NOTES:

Toxicity values quoted in this table are for the oral pathway unless otherwise noted.

Essential nutrients (calcium, iron, magnesium, potassium and sodium) are not considered in this table. See Sections 2.4

and 2.5 for full discussion of essential nutrients.

- NA = Not Available
- NC = Not calculated
- RFD = Reference Dose
- SF = Slope Factor (a) IRIS, 1995
- (b)
- HEAST, 1995 (c) ECAO
- (d) Massachusetts DEP, October 1994.
- ⁽¹⁾ Based on screening carried out in Table 2-6 and Section 2.5.
- $^{\scriptscriptstyle (2)}$ $\,$ Pyrenc RfD and RBC used as a surrogate for RfD and RBC of various PAHs.
- ⁽³⁾ n-Nonane RfD used as surrogate for TICs RfD. As no RBC is available for n-nonane. TRPH was carried as a COPC.
- (4) DDT RID used as surrogate for DDD and DDE RfDs
- ⁽⁵⁾ Endosulfan RFD and RBC used as surrogates for endosulfan sulfate RID and REC, respectively.
- (6) Although beryllium exceeds RBC-based benchmarks, the maximum concentration does not exceed background, so it is not considered a COPC.
- ⁽⁷⁾ Slope factor is for inhalation pathway.
- ⁽⁸⁾ Nickel refinery dust inhalation slope factor used as surrogate for Nickel slope factor.
- (9) Toxicity equivalence factor (TEF) was applied to the benzo(a)pyrene slope factor, based on the relative potency of this chemical to benzo(a)pyr

At the suggestion of USEPA Region IV, toxicity values for n-nonane were used as surrogates for TRPH. As no RBC was available for n-nonane, TRPH was added to the list of COPCs for surface soil. All other chemicals detected in surface soil whose maximum concentrations exceeded the RBCs had been already selected as COPCs based on previous screening described in Sections 2.7.1.2 and 2.7.1.3, and Table 2-22.

2.7.1.5 Chemicals of Potential Concern Selection Process. The chemicals of potential concern (COPCs) selection process determines those chemicals that are the most toxic and that are anticipated to create the greatest potential risk. As stated previously, Figure 2-10 illustrates the criteria used to select COPCs in this risk assessment

Identification of the COPCs for the risk assessment was accomplished in accordance with USEPA (1989a) guidance. All detected constituents were included as COPCs for the risk assessment with the following exceptions:

- As per USEPA Region IV risk assessment guidance (USEPA, 1992b), inorganic constituents present at concentrations less than twice background concentrations were excluded from the list of COPCs. Only those constituents for which the maximum detected concentration was greater than twice the background concentration were retained as COPCs.
- Chemicals detected in less than 5% of the samples analyzed per media.
- Chemicals represented in less than 1% of the potential overall risk via the concentration-toxicity screen (USEPA, 1989a), and whose maximum concentration detected did not exceed a benchmark based on USEPA Region III RBCs (USEPA, 1995a).

Based on the above evaluation, a group of COPCs was carried through the quantitative risk assessment for each of the environmental media, groundwater and soil. This selection is summarized in Table 2-25.

Tentatively Identified Compounds (TICs) and TRPH. Where it was appropriate, TICs were included within the quantitative risk analysis as COPCs for soil and groundwater. Tentatively identified chemicals in the Montgomery Watson 1993 dataset associated with petroleum products were summed for quantification. Categories of TICs included in this evaluation were: alkanes, unknown hydrocarbons, substituted benzenes, PAHs,

CHEMICALS OF POTENTIAL CONCERN IN ENVIRONMENTAL MEDIA AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

	Affected	Media
Constituents	Groundwater	Surface Soil (0.2 ft.)
<u>VOCs</u>		
Bromodichloromethane	Х	
Chloroform	X	
Methylene chloride	Х	
BNAs		
Bis(2-ethylhexyl)phthalate	Х	
Benzo(a)pyrene		Х
TRPHs (TICs)		Х
Pesticides/PCBs		
4,4'-DDD		Х
Chlordane		Х
Metals		
Aluminum	Х	Х
Arsenic	Х	Х
Barium	Х	
Cadmium		Х
Calcium	Х	
Chronium	Х	Х
Copper		Х
Iron	Х	Х
Lead	Х	
Manganese	Х	Х
Mercury		Х
Nickel	Х	Х
Potassium		Х
Vanadium	Х	Х

BNAs Base/neutral and acid	l extractable compounds
PCBs	Polychlorinated biphenyls
TRPHs	Total recoverable petroleum hydrocarbons.
VOCs	Volatile organic compounds.

cycloalkanes, and aromatics. The summed petroleum-related TICs were treated as TPH in screening and the risk characterization

Unknown and other partially identified TICs were not included for further analysis due to the lack of information on these chemicals. Organic acids detected in soil and groundwater were not included in the quantitative risk assessment as these chemicals are the result of natural processes by biological organisms (bacteria) in the breaking down or "weathering" of petroleum product at the site.

USEPA Region IV has adopted an approach to TPH developed by the State of Massachusetts DEP (Massachusetts DEP, 1994). This approach uses the toxicity values of certain hydrocarbon compounds (e.g. n-hexane, n-nonane, eicosane) for fractions of TPH. The toxicity of hydrocarbons tends to decrease with increasing carbon chain length, n-Hexane has an RfD of 0.06, n-nonane an RfD of 0.6, and eicosane an RfD of 6.

After review and discussion with USEPA Region IV, toxicity values for n-nonane (C9) were used as surrogate values for TPH/TRPH and fuel-associated TICs. Use of n-nonane as a surrogate was felt to be more representative of the TPH present at the site than use of n-hexane, as volatile fractions of TPH (C4-C7) would be expected to attenuate by weathering more rapidly than heavier components. Also, certain of the soil samples (such as those taken by Geraghty and Miller) were analyzed for hydrocarbons solely in the C8-C20 carbon range.

2.7.2 Potential Routes of Migration

Contaminants may migrate from a source area through a variety of processes. Volatile contaminants may be released into air and migrate in the vapor phase. Liquid or aqueous-phase contaminants may migrate to both soils and groundwater through direct infiltration. Erosion related to surface runoff or wind may transport contaminants sorbed to surface soils. Infiltrating precipitation may dissolve contaminants and carry them into deeper soils where they can be adsorbed, or into groundwater in the dissolved phase. Dissolved phase contaminants may be carried in the down gradient direction by groundwater flow in an aquifer.

Although other contaminated media are present at OU-5/Site WP-1, the principal route of migration of contaminants is through shallow groundwater. The impacts associated with the surface water and sediment samples have been further evaluated in the OU-9, Boundary Canal RI/RA. Past activities allowed contaminants to enter soil and surface water, which

eventually migrated to shallow groundwater. Migration of contaminants via surface water occurs intermittently, during storm events.

OU-5/Site WP-1 and its drainages are situated on a developed portion of the Base which includes buildings, roads, and parking areas. The cycle of water through the site begins with precipitation. During rainfall events, water percolates rapidly through the limestone and weathered limestone bedrock underlying the site. Surface water runoff is over land to one of the drainage swales or ditches located in the immediate area of the site. The drainage swales and canals provide adequate surface water drainage for this site and are typically dry during non-storm events. Given the highly transmissive underlying formation, rainwater and surface water will typically infiltrate rapidly into the shallow aquifer system. It is estimated that horizontal groundwater movement can be on the order of tens of feet during a single rainfall event. Once the rainfall ceases, the water table returns to near static conditions and groundwater movement decreases dramatically.

Between rainfall events, evaporation from the surface soils returns water from the aquifer to the atmosphere. The rate of loss is greatest with open water bodies and decreases with increasing distance from the water table.

The natural concentrations of chemicals in the soil, rock, and water have a controlling effect on the fate and transport mechanisms. Soils at the site exist primarily as a veneer on the bedrock surface. A considerable amount of the OU-5/Site WP-1 area is covered by asphalt, roads. or buildings. The soil has both organic and iron precipitants. Nevertheless the calcium carbonate from the underlying oolite is the primary mineral present. The site drainage swales also receive runoff from the asphalted parking area located cast of Building 164.

2.7.3 Exposure Assessment

This section of the risk assessment identifies and describes potential human receptors, reviews possible pathways of exposure for compounds of concern at OU-5/Site WP-1, and presents estimates of exposure doses resulting from identified pathways at OU-5/Site WP-1. An exposure assessment is conducted to identify potential sources and mechanisms of release, transport pathways (e.g., groundwater, surface water, soil, and air), routes of exposures (ingestion, inhalation, dermal contact), and potential on-site and off-site receptor populations (current users of the site, as well as adjacent populations which may be exposed

to chemicals that have been transported off-site). This information provides the basis for constructing site-specific exposure scenarios.

Two environmental media were considered in this document - groundwater and surface soil. It should be noted that guidance on what depth range should be used for surface soil differs between the USEPA (0 to 12 inches) and the Florida DEP (0 to 24 inches). Samples taken between 0 and 24 inches below land surface (bls) were considered surface soil samples, so receptor exposure during gardening or landscaping activities could be evaluated in this assessment. This choice seems reasonable for south Florida, as the year-round, mild climate would permit possible residential gardening and frequent landscaping activities on base. No subsurface soil sampling was conducted because most soil layers at OU-5/Site WP-1 are only one to two inches deep and the underlying layers are composed of limestone and bedrock. Furthermore, the sediment and surface water samples collected at OU-5/Site WP-1 from the area canals and drainage ditches are not evaluated in this document. The potential human health effects due to exposures associated with the canal system are addressed in the BRA for OU-9, Boundary Canal Evaluation, which will be submitted as a separate report.

Other information considered in the development of present and future exposure scenarios includes: physical characteristics of the site and surrounding area such as climatology, groundwater hydrology, location and description of surface water and surrounding land use and available state-specific guidelines relevant to exposure and risk assessments.

A critical step in assessing the potential risk to public health is to identify the pathways through which exposure could occur. A typical transport pathway consists of four necessary elements: 1) a source and mechanism of chemical release, 2) an environmental transport medium, 3) a point of potential contact with the contaminated medium, and 4) an exposure route (inhalation of vapors, ingestion of groundwater, etc.). All four of these elements must be present for a pathway to be complete.

Exposure Point Concentration. In accordance with USEPA methodology (1989a), the medium-specific 95 percent UCL on the arithmetic mean concentrations for the COPCs will be used as exposure point concentrations (EPCs) to estimate reasonable maximum exposure (RME). The RME approach is suggested by the USEPA (1989a) to provide an estimate of the maximum exposure (and therefore risk) that might occur. The RME. corresponds to a duration and frequency of exposure greater than is expected to occur on an average basis. In those instances where the calculated 95 percent UCL exceeds the maximum detected

concentration, the maximum detected concentration was used as the EPC for a more accurate estimate of RME concentration (USEPA, 1989a).

The total number of samples collected, as well as the sources of the data used in the risk assessment and included in the database for the calculation of each COPC exposure point concentration, varied by medium.

Once the database for each medium was developed, the 95 percent UCL concentration on the arithmetic mean concentration (one-tailed test, assuming a lognormal distribution) was calculated and compared to the maximum COPC concentration to determine the EPC for each COPC. The results of these analyses for the sampled media are presented in Tables 2-13 and 2-14. The information presented in these tables is discussed in the following subsections. An example of the data reduction used to calculate the arithmetic mean and UCL for each COPC is shown in Table 2-15.

Exposure Scenarios. Exposure pathways identified at OU-5/Site WP-1 are shown in Table 2-16 and are associated with soils or groundwater. With the exception of the VOCs, the chemicals detected at the site have low environmental mobility.

Under present conditions, access to the site is limited primarily to base workers performing duties that might require site access, such as cutting the grass. COPCs detected in the surface soils include one BNA, two pesticides, nine metals, and TRPH. Base workers cutting the grass at the site could be exposed to the soils via direct contact with exposed arms and face, incidental ingestion of soils that might adhere to the hands, and inhalation of dusts or vapors generated while cutting the grass. The site is sparsely vegetated, therefore, potential exposure rates will not be reduced by a vegetation factor.

If the operation of Homestead ARB is to continue in the future, exposure pathways at the site are unlikely to change. Given the location of OU-5/Site WP-1 in the midst of base administration activity, future development of this area for intensive base operations is unlikely. Foreseeable future use conditions at the site would result in potential exposure pathways similar to those discussed for present site conditions. However, it should be noted, that damage from the hurricane would necessitate construction activities regardless of future land use.

Under current reuse plans, OU-5/Site WP-1 will be under cantonment of the US Air Force Reserve. Therefore, reuse of OU-5/Site WP-1 for residential purposes is unlikely in the

foreseeable future. However, for risk characterization, hypothetical future residents were considered as receptors. Exposure pathways for hypothetical future residents would include direct contact with the surface soils, incidental ingestion of the surface soils, and inhalation of fugitive dust or vapors.

Hypothetical future construction workers were included in the risk characterization. Exposure pathways for future hypothetical construction workers would include ingestion and inhalation of soil. This receptor is evaluated for only surface soils (less than two feet) as no subsurface soil (greater than two feet) is expected at the site since most soil layers are usually one to two inches deep and the underlying layers are composed of limestone and bedrock.

Although it is unlikely that potable wells would be installed in the vicinity of the site, a conservative assumption made in this risk assessment is that a potable well is installed in the groundwater plume, downgradient of the site. Exposure of hypothetical future residents to affected groundwater via ingestion, inhalation, and dermal contact are considered potential exposure pathways.

In summary, workers cutting the grass on the site will be used to represent current exposure to the on-site soils. In the unforeseen event that the site is no longer under cantonment of the US Air Force Reserve, hypothetical future exposure pathways considered included residential development of the site. Table 2-16 and Figure 2-9 summarize the potential exposure pathways for OU-5/Site WP-1.

2.7.4 Toxicity Assessment

This section of the baseline risk assessment provides information on the human health effects of site-specific COPCs. The information presented in this section provides a basis for the dose-response assessment carried out in the quantitative risk assessment.

Evaluation of the toxic potential of a chemical involves the examination of available data that relate observed toxic effects to doses. Generally, there are two categories of information that are considered in this part of a quantitative risk assessment:

- Information on the potential acute or chronic non-cancer effects of chemicals, and
- Information on the potential for chemicals to initiate or promote cancers.

A wide variety of factors must be considered in using health effects data in qualitative or quantitative assessments. As discussed in the following subsections, there may be a variety of relationships between dose and effects. Also, the fact that some chemicals display thresholds (i.e., there are doses below which the chemical does not cause an effect) must be considered.

Non-Carcinogenic Effects. In general, non-carcinogenic (acute or chronic systemic) effects are considered to have threshold values, while carcinogenic effects are considered to not have thresholds. Toxicity studies for the former focus on identifying where this threshold occurs. The threshold can be related to a reference dose (RfD). A chronic RfD is an estimate of a daily exposure level for which people, including sensitive individuals, do not have an appreciable risk of suffering significant adverse health effects. Exposure doses above an RfD could possibly cause health effects.

Carcinogenic Effects. Studies of carcinogenicity tend to focus on identifying the slope of the linear portion of a curve of dose versus response. A plausible upper-bound value of the slope is called the cancer slope factor (CSF) or cancer potency factor (CPF). The product of the CSF and the exposure dose is an estimate of the risk of developing cancer. In accordance with current scientific policy concerning carcinogens, it is assumed that any dose, no matter how small, has some associated response. This is called a non-threshold effect. In this assessment, the non-threshold effect was applied to all probable carcinogens.

Toxicological Properties. The risks associated with exposure to COPCs at OU-5/Site WP-1 are a function of the inherent toxicity (hazard) of each chemical and the exposure dose. This section addresses the inherent toxicological properties of the COPCs. The exposure doses are estimated in the Exposure Assessment section which follows.

A distinction is made between carcinogenic and non-carcinogenic effects, and two general criteria are used to describe these effects: excess lifetime cancer risk for constituents which are thought to be potential human carcinogens and the hazard quotient (HQ) for constituents that cause non-carcinogenic effects. For potential carcinogens, the current regulatory guidelines (USEPA, 1989a) use an extremely conservative approach in which it is assumed that any level of exposure to a carcinogen could hypothetically cause cancer. This is contrary to the traditional toxicological approach to toxic chemicals, in which finite thresholds are identified, below which toxic effects are not expected to occur. This traditional approach still is applied to non-carcinogenic chemicals.

Toxicity Values. In general, CSFs, cancer classifications, RfDs, and RfCs are taken from IRIS (1996) or, in the absence of IRIS data, the USEPA Health Effects Assessment Summary Tables (HEAST) (USEPA, 1995). Because toxicity values for dermal exposure are rarely available, several adjustments were made to toxicity values for use in calculating dermal dose as per Region IV supplemental guidance to RAGS issued in March of 1994. The PAH CSFs were not adjusted to assess dermal exposure since the portal of entry differs in the outcome of tumors from oral and dermal exposure (USEPA, 1989a). Additionally, oral toxicity constants (RfDs and CSFs) were adjusted for dermal use via the application of oral absorption efficiency values obtained from Region IV supplemental guidance to RAGS issued in March of 1994. The factors used to correct both exposure dose calculations for dermal absorption from soil and the factors used to adjust oral toxicity constants (RfDs and CSFs) for use in calculating risks and hazard indices via dermal exposure are provided in Table 2-26. Unadjusted oral and inhalation RfDs are provided in Table 2-27. CSFs, cancer type or tumor sites, and carcinogen classifications for the COPCs at the site are presented in Table 2-28. Derivation of the adjusted RfDs and CSFs is shown in Table 2-29.

There are no USEPA-verified acceptable doses (i.e., RfDs) for lead. Considerable controversy currently exists concerning the appropriate acceptable doses for lead. The best method for evaluating exposure to lead is through the measurement of lead in blood or blood lead levels. Lead was evaluated in this risk assessment based on acceptable blood lead levels for young children using the USEPA (1994a) IEUBK model (LEAD 0.99d).

USEPA Region IV has adopted an approach to TPH developed by the State of Massachusetts DEP (Massachusetts DEP, 1994). This approach uses the toxicity values of certain hydrocarbon compounds (e.g. n-hexane, n-nonane, eicosane) as surrogate toxicity values for fractions of TPH (Andrews and Snyder, 1991). The toxicity of hydrocarbons tends to decrease with increasing carbon chain length. n-Hexane has an RfD of 0.06, n-nonane an RfD of 0.6, and eicosane an RfD of 6.

After review and discussion with USEPA Region IV, n-nonane was used to calculate non-cancer risks associated with exposure to Total Recoverable Petroleum Hydrocarbons (TRPHs) and tentatively identified compounds (TICs) shown to be petroleum related. The toxicity of hydrocarbons generally decreases as chain length increases (Andrews and Snyder, 1991). The light-end hydrocarbons (e.g., n-hexane) present in TPH tend to attenuate by weathering faster than heavier components, leaving the long-chain, less toxic components of TPH. Thus, use of n-nonane as a toxicity surrogate for the TPH represents a conservative (protective) approach.

DERMAL AND ORAL ABSORPTION EFFICIENCIES FOR COMPOUNDS OF CONCERN AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Constituents	Absorption	Efficiencies
	Dermal (a)	Oral (b)
VOCs		
Bromodichloromethane	0.01	0.80
Chloroform	0.01	0.80
Methylene chloride	0.01	0.80
BNAs		
Benzo(a)pyrene	0.01	0.50
Bis(2- ethylhexyl)phthalate	0.01	0.50
TRPHs (as n-nonane)	0.01	0.50
Pesticides		
Chlordane	0.01	0.50
4,4'-DDD	0.01	0.50
Metals		
Aluminum	0.001	0.20
Arsenic	0.001	0.95
Barium	0.001	0.20
Cadmium	0.001	0.20
Chromium	0.001	0.20
Copper	0.001	0.20
Lead	0.001	0.20
Manganese	0.001	0.20
Mercury	0.001	0.20
Nickel	0.001	0.20
Vanadium	0.001	0.20

(a) Used to adjust dermal dose calculation for absorption from soil as per Region IV Supplemental Guidance to RAGS Bulletin, Vol. 1 No. 1, USEPA, Atlanta, Georgia, March 1994.

(b) Used to adjust oral toxicity constants (RfDs and CPFs) to estimate effects via dermal exposure. Values as per Region IV Supplemental Guidance to RAGS Bulletin, Vol. 1 No. 1. USEPA. Atlanta, Georgia, March 1994.

REFERENCE DOSES FOR COMPOUNDS OF CONCERN AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

stituent	Chronic Oral RfD (mg/kg/day)		Subchronic Oral RfD (mg/kg/day)		Chronic Inhalation RfC (mg/kg/day)		Subchronic Inhalation RfC (mg/kg/day)	
VOCs								
Bromodichloromethane	2.00E-02	a	2.00E-02	b	NA		NA	
Chloroform	1.00E-02	a	1.00E-02	b	NA		NA	
Methylene chloride	6.00E-02	a	6.00E-02	b	8.57E-01	b	8.57E-01	b
BNAs								
Benzo(a)pyrene ⁽¹⁾	3.00E-02	а	3.00E-01	b	NA		NA	
Bis(2-ethylhexyl)phthalate	2.00E-02	а	NA		NA		NA	
TRPHs (as n-nonane) ⁽²⁾	6.00E-01	с	NA		NA		NA	
Pesticides								
Chlordane	6.00E-05	а	6.00E-05	b	NA		NA	
4,4'-DDD ⁽³⁾	5.04E-04	a	5.04E-04	b	NA		NA	
Metals								
Aluminum	1.00E+00	d	NA		NA		NA	
Arsenic	3.00E-04	а	3.00E-04	b	NA		NA	
Barium	7.00E-02	a	7.00E-02	b	1.00E-04	b	1.00E-03	b
Cadmium (water)	5.00E-04	а	NA		NA		NA	
Cadmium (food)	1.00E-03	а	NA		NA		NA	
Chromium VI	5.00E-03	a	2.00E-02	b	NA		NA	
Copper	3.70E-02	a	3.70E-02	а	NA		NA	
Lead	NA		NA		NA		NA	
Manganese	2.40E-02	a	NA		1.43E-05	а	NA	
Mercury	3.00E-04	b	3.00E-04	b	8.60E-05	а	8.60E-05	b
Nickel	2.00E-02	а	2.00E-02	b	NA		NA	
Vanadium	7.00E-03	b	7.00E-03	b	NA		NA	

a IRIS, 1996

b USEPA, 1995

c Massachusetts DEP, 1994

d ECAO

(1) The pyrene RfD was used as a surrogate for the benzo(a)pyrene RfD

(2) The n-nonane RfD was used as a surrogate for TRPHs RfD

(3) The DDT RfD was used as a surrogate for the DDD RfD

CANCER SLOPE FACTORS, TUMOR SITES, AND USEPA CANCER CLASSIFICATIONS FOR COMPOUNDS OF CONCERN AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Constituent	CSF (r	ng/kg	/day)-1		Tu	USEPA Classification	
-	Oral		Inhalation		Oral		
VOCs							
Bromodichloromethane	6.20E-02	а	NA		kidney	NA	B2
Chloroform	6.10E-03	а	8.10E-02	b	kidney	liver	B2
Methylene chloride	7.50E-03	а	1.65E-03	a	liver	liver and lung	B2
BNAs							
Benzo(a)pyrene	7.30E+00	а	6.10E+00	с	stomach	respiratory tract	B2
Bis(2-ethylhexyl)phthalate	1.40E-02	a	NA		liver	NA	B2
Pesticides							
Chlordane	1.30E+00	а	1.30E+00	а	liver	liver	B2
4,4'-DDD	2.40E-01	а	NA		liver	NA	B2
Metals							
Arsenic	1.50E+00	а	1.50E+01	а	skin	respiratory tract	А
Cadmium	NAP		6.30E+00	а	NAP	respiratory tract	B1
Chromium VI	NAP		4.10E+01		NAP	lung	А
Lead	NA		NA		NA	NA	B2
Nickel ⁽¹⁾	NAP		8.40E-01	а	NA	NA	а

mg/kg/day milligrams per kilogram per day

NA Not available

NAP Not applicable since it is considered carcinogenic via inhalation only

a IRIS, 1996

b USEPA, 1995

c ECAO

(1) Nickel refinery dust inhalation slope factor used as surrogate for nickel slope factor

ADJUSTED TOXICITY VALUES USED TO ASSESS DERMAL EXPOSURE AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Constituent	C	Oral Toxic	city Values		Oral Absorption		Dermal Toxicity Value (Adjusted Oral)	
	RfDo	Source	CSFo	Source	Efficiency	Source	RfDa	CSFa
VOCs								
Bromodichloromethane	2.00E-02	а	6.20E-02	а	0.80	e	1.6E-02	7.8E-02
Chloroform	1.00E-02	а	6.10E-03	а	0.80	e	8.0E-03	7.6E-03
Methylene chloride	6.00E-02	а	7.50E-03	а	0.80	e	4.8E-02	9.4E-03
BNAs								
Benzo(a)pyrene ^f	3.00E-02	а	7.30E+00	а	0.50	e	NA	NA j
Bis(2-ethylhexyl)phthalate	2.00E-02	а	1.40E-01	a	0.50	e	1.0E-02	2.8E-02
TRPHs (as n-nonane) ^g	6.00E-01	d	NA		0.50	e	3.0E-01	NA
Pesticides								
Chlordane	6.00E-05	а	1.30E+00	а	0.50	e	3.0E-05	2.6E+00
4,4'-DDD ^h	5.04E-04	а	2.40E-01	а	0.50		2.5E-04	4.8E-01
Metals								
Aluminum	1.00E+00	с	NA		0.20	e	2.0E-01	NA
Arsenic	3.00E-04	а	1.50E+00	а	0.95	i	2.9E-04	1.6E+00
Barium	7.00E-02	а	NA		0.20	e	1.4E-02	NA
Cadmium (water)	5.00E-04	а	NAP		0.20	e	1.0E-04	NA
Cadmium (food)	1.00E-03	а	NAP		0.20	e	2.0E-04	NA
Chromium	5.00E-03	а	NAP		0.20	e	1.0E-03	NA
Copper	3.70E-02	а	NA		0.20	e	7.4E-03	NA
Lead	NA	а	NA		0.20	e	NA	NA
Manganese	2.40E-02	а	NA		0.20	e	4.8E-03	NA
Mercury	3.00E-04	b	NA		0.20	e	6.0E-05	NA
Nickel	2.00E-02	а	NAP		0.20	e	4.0E-03	NA
Vanadium	7.00E-03	b	NA		0.20	e	1.4E-03	NA

CSFa Adjusted cancer slope factor (mg/kg/day)⁻¹.

CSFo Oral cancer slope factor $(mg/kg/day)^{-1}$.

NA Not available.

NAP Not applicable, carcinogenic only by inhalation route.

- RfDa Adjusted reference dose (mg/kg/day).
- RfDo Oral reference dose (mg/kg/day).
 - a IRIS, 1996
 - b USEPA, 1995
 - c ECAO
 - d Massachusetts DEP, 1994
 - e Default value as per Region IV Supplemental Guidance to RAGS Bulletin, Vol. 1, No. 1, USEPA, Atlanta, Georgia, March 1994.
 - f Pyrene RfD used as surrogate for benzo(a)pyrene RfD
 - g The n-nonane RfD was used as a surrogate for TRPHs RfD
 - h DDT RfD used as surrogate for DDD RfD
 - i National Research Council, 1982
 - j PAH slope factors were not adjusted to assess dermal exposure since the portal of entry differs in the outcome of tumors from oral and dermal exposure (USEPA, 1989a).

The CSFs for benzo(a)pyrene were used to calculate cancer risks associated with exposure to all carcinogenic PAHs at the site. In accordance with USEPA Region IV guidance (USEPA, 1992b), the oral CSF and inhalation CSF for benzo(a)pyrene were converted using toxicity equivalency factors (TEFs) for each individual carcinogenic PAH. This approach is based on the relative potency of each compound to the potency of benzo(a)pyrene (USEPA, 1992b). There are a limited number of RfDs available for the PAHs detected at the site. The following PAHs have USEPA-verified RfDs: fluoranthene and pyrene. The RfD for pyrene was used to calculate non-cancer risks associated with exposure to detected non-carcinogenic PAHs not having individual RfDs.

2.7.5 Risk Characterization

This section of the baseline risk assessment describes how calculated exposure doses are converted into health risks. This section characterizes risks as part of a quantitative risk assessment for the site. Risk characterization involves the integration of health effects information developed as part of the dose-response assessment with exposure estimates developed as part of the exposure assessment. The result is a quantitative estimate of chronic and non-carcinogenic risks based on the presumption that a threshold dose is required to elicit a response, as well as a quantitative estimate of carcinogenic risks presumed to exist regardless of the dose. These estimates are usually presented in either probabilistic terms (e.g., one-in-one-million), or with reference to specific benchmark or threshold levels. Because risk estimates are based on a combination of measurements and assumptions, it is important to provide information on sources of uncertainty in risk characterization. The key elements of risk, a presentation included in this section are: an estimation of human dose, an estimation of risk, a presentation of risk, and an uncertainty analysis.

2.7.5.1 Carcinogenic Risks. Public health risks are evaluated separately for carcinogenic and non-carcinogenic effects. The excess lifetime cancer risk is an estimate of the increasedrisk of cancer which results from lifetime exposure, at specified average daily dosages, to chemicals detected in media at the site. Excess lifetime cancer risk, equal to the product of the exposure dose and the slope factor, is estimated for each known, probable, or possible carcinogenic chemical in each medium. The risk values provided in this report are an indication of the increased risk, above that applying to the general population, which may result from the exposure scenarios described in the Exposure Assessment Section 2.7.3. The risk estimate is considered to be an upperbound estimate; therefore, it is likely that the true risk is less than that predicted by the model. Current regulatory methodology assumes that

excess lifetime cancer risks can be summed across routes of exposure and constituents to derive a "Total Site Risk" (USEPA, 1989a). The USEPA, OSWER Directive 9355.0-30, Role of the Risk Assessment in Superfund Remedy Selection Decisions (1991e) has stated that sites with an excess lifetime cancer risk less than 10^4 (1 in 10,000) generally do not warrant remedial action. However, the state of Florida's target cancer risk is 10^{-6} .

The incremental risk is calculated for each exposure scenario based on the following basic equation:

where the slope factor (SF) is in units of (mg/kg/day)⁻¹ based on a compound specific cancer bioassay dose response curve.

The exposure dose is adjusted over a 70-year lifetime. The summation of dose is in keeping with the concept that for genotoxic agents there exists no threshold dose and implies that total, lifetime exposure is of greater importance than the actual dose during the exposure event(s). Ingestion and inhalation risks are calculated separately since chemicals often have differentSFs for differing routes of exposure. The different SFs relate to the pharmacokinetics inherent in each chemical/organ and the specific routes of uptake.

Slope factors are derived by USEPA in an intentionally conservative way, that is, the actual risk is not expected to exceed the predicted risk, and could be considerably lower. Cancer risks calculated using these conservative slope factors and reasonable maximum exposure estimates are upper bound estimates of excess cancer risk potentially arising from exposure to the chemicals in question. A number of assumptions have been made in the derivation of these values, many of which are likely to overestimate exposure and toxicity. The actual incidence of excess cancers is likely to be lower than these estimates and may be zero.

Lifetime daily intakes, using an averaging time of up to 70 years, effectively prorates the total cumulative dose over a lifetime. This approach is based on the assumption for carcinogens that a high dose received over a short period of time at any age is equivalent to a corresponding low dose received over a lifetime (USEPA, 1989a). This assumption is unlikely to be true for all carcinogens, and introduces uncertainty into the assessment of potential risk. This assumption may also lead to an overestimate or underestimate of potential risk, depending upon the actual timing of exposure and the mechanism of action of individual carcinogens.

The magnitude of cancer risk relative to Superfund site remediation goals in the National Contingency Plan ranges from 10^{-4} (one-in-ten-thousand) to 10^{-6} (one-in-one-million) depending on the site, proposed usage, and chemicals of concern (USEPA, 1989a). Within this range, the level of risk which is considered to be acceptable at a specific site is a risk management decision and is decided on a case-specific basis. It is generally accepted that risks above this range require attention. The one-in-a-million level of risk (expressed as 1E-06) is often referred to as the de minimis level of risk; risks calculated below this range would not require attention. The 1E-06 risk level does not equate to an actual cancer incidence of one-in-a-million. For substances that may cause cancer, the risk assessment process uses animal data to predict the probability of humans developing cancer over a 70-year lifetime. The numbers are given as upper bounds; the real risk is expected to be less. The one-in-a-million risk level is a theoretical prediction that no more than one person out of a million lifetimes would contract cancer due to an environmental exposure. By the way of comparison, the average person in the U.S. incurs a background risk of cancer (from all causes) of about one chance in four (0.25). Adding a risk of 0.000001 to a background risk of 0.25 is of little significance to any single individual. These small risk levels may be of concern only if the exposed population includes many millions of people.

2.7.5.2 Chronic Health Risks. The HQ is the ratio of the estimated exposure dose to the RfD. This ratio is used to evaluate non-carcinogenic health effects due to exposure to a chemical. An HQ greater than 1 indicates that the estimated exposure dose for that chemical exceeds acceptable levels for protection against non-carcinogenic effects. Although an HQ of less than 1 suggests that non-carcinogenic health effects should not occur, an HQ of slightly greater than 1 is not necessarily an indication that adverse effects will occur. The sum of the HQs is termed the hazard index (HI). Current regulatory methodology assumes that HIs can be summed across exposure routes for all media at the site to derive a "Total Site Risk" (USEPA, 1989a). The USEPA, OSWER Directive 9355.0-30, Role of Risk Assessment in Superfund Remedy Selection Decisions (1991e) has stated that sites with a non-carcinogenic HQ less than 1 generally do not warrant remedial action.

The USEPA has developed a set of health based benchmark numbers, called reference doses, or RfDs, as guideposts in a risk assessment. Reference doses are an adaptation of the earlier toxicological measure of "acceptable daily dose" or ADI. The unit of a reference dose is mg contaminant/kg body weight/day. The potential for adverse effects on human health (other than cancer) is evaluated by comparing an intake over a specific time period with a reference dose derived for a similar exposure period.

The hazard index is the ratio (unitless) of the estimated exposure dose (D) of a compound to a reference dose (RfD) judged to be without adverse effects given long-term exposure. Thus, the index is used as a measure of potential noncarcinogenic health risks. Due to the margin of safety built into the RfD value, exceedence of the number has no immediate meaning with regard to specific health effects, the frequency of effects, or the magnitude of effects. However, exceedence of the number should serve as an indicator that the potential for unacceptable exposure does exist and further evaluation needs to be considered. The effects of noncarcinogens in the body vary greatly with regard to potential target organs, threshold dose, and "severity" of effect. Therefore, the individual toxicity for each compound needs to be assessed.

If the hazard index is less than 1.0, then no chronic health effects are expected to occur. If the hazard index is greater than 1.0, then adverse health risks are possible. In the case of noncarcinogenic effects, chronic exposure below a threshold dose results in a non-response or a diminished response.

2.7.5.3 Risks Associated With Exposure to Groundwater. Risks for a hypothetical future adult resident exposed to groundwater are shown in Table 2-30. The excess lifetime cancer risk and HI are 5E-04 and 3, respectively. The excess lifetime cancer risk level associated with hypothetical future resident conditions at the site is above the USEPA remediation-based risk benchmarks for carcinogens (10^{-4} to 10^{-6}) and above the state of Florida's criterion of 1E-06. The hazard index also exceeds the risk benchmark of one.

In accordance with current USEPA Region IV guidance (USEPA, 1995d), the inhalation and dermal exposure to VOCs during showering are assumed to be equivalent to the ingestion dose. This is based on a growing body of evidence that risk estimates from ingestion of VOCs in potable water, inhalation of volatiles from showering, and dermal exposure to volatiles during showering or bathing are similar (Andelman, 1985; Andelman, et.al., 1986, 1987; McKone, 1987, and Jo, et.al., 1990). Given this assumption, risks via the inhalation and dermal routes for groundwater contact can be calculated using the oral dose (mg/kg/day-1) and multiplying by the inhalation slope factor for carcinogens and dividing by the RfD for noncarcinogens. No inhalation RfCs were available for bromodichloromethane and chloroform, thus, oral RfDs are used for these compounds. Therefore, the total risk via groundwater contact including oral, dermal and inhalation exposures is 5E-04 for cancer risk and 3 for noncancer risk. Inorganics, including arsenic are not expected to volatilize from the water droplet, thus, the primary exposure routes via groundwater use would be ingestion

GROUNDWATER INGESTION EXPOSURE DOSES AND RISK CALCULATIONS FOR A HYPOTHETICAL FUTURE ADULT RESIDENT AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Constituent	Cgw (mg/L)	GWExD (mg/kg-day)	Toxicity Values	Calculated Risk
CANCER EFFECTS			CSFo	
VOCs				
Bromodichloromethane	0.004	4.2E-05	6.20E-02	2.6E-06
Chloroform	0.004	4.2E-05	6.10E-03	2.6E-07
Methylene Chloride	0.003	3.9E-05	7.5E-03	2.9E-07
BNAs				
Bis(2-Ethylhexyl)phthalate	0.04	4.5E-04	1.4E-02	6.3E-06
Metals				
Arsenic	0.026	3.1E-04	1.50E+00	4.6E-04
Chromium	0.0292	3.4E-04	NAP	NAP
Lead	0.0073	8.6E-05	-	-
Nickel	0.0108	1.3E-04	NAP	NAP
			ELCR =	5E-04
NON-CANCER EFFECTS			RfDo	
VOCs				
Bromodichloromethane	0.004	9.9E-05	2.0E-02	4.9E-03
Chloroform	0.004	9.9E-05	1.0E-02	9.9E-03
Methylene chloride	0.003	9.0E-05	6.0E-02	1.5E-03
BNAs				
Bis(2-Ethylhexyl)phthalate	0.0385	1.1E-03	2.0E-02	5.3E-02
Metals				
Aluminum	4.266	1.2E-01	1.0E+00	1.2E-01
Arsenic	0.026	7.2E-04	3.0E-04	2.4E+00
Barium	0.0372	1.0E-03	7.0E-02	1.5E-02
Chromium (VI)	0.0292	8.0E-04	5.0E-03	1.6E-01
Lead	0.0073	2.0E-04	-	-
Manganese	0.044	1.2E-03	2.40E-02	5.0E-02
Nickel	0.0108	3.0E-04	2.00E-02	1.5E-02
Vanadium	0.0164	4.5E-04	7.00E-03	6.4E-02
			HI =	3E+00

S	Insufficient data; USEPA-verified toxicity value not available.
NAP	Cancer slope factor and/or reference dose applies to inhalation pathway only, not to ingestion.
Cgw	Constituent exposure point concentration in groundwater in milligrams per liter (mg/L) (see Table 4-2).
GWExD	Ground-water exposure dose in milligrams per kilogram per day (mg/kg/day).
CSFo	Cancer Slope Factor, Oral
RfDo	Reference Dose, Oral
ELCR	Excess lifetime cancer risk.
HI	Hazard index (sum of the hazard quotients).

and to a small degree dermal. The dermal dose is expected to be two to three orders of magnitude less than oral dose.

The primary contributor to the carcinogenic risk estimate is arsenic (98% of the calculated risk). This compound was detected in nine of nine samples at concentrations ranging from $3.9 \ \mu g/l$ to $92 \ \mu g/l$. Two of the nine samples had arsenic concentrations (60 and $92 \ \mu g/l$) higher than the state and federal drinking water standard of $50 \ \mu g/l$. The arsenic risk level is based on unfiltered samples which exhibited high turbidity during the 1991 G&M sampling program. Therefore, this level probably overestimates concentrations in a hypothetical potable well. Further, the presence of arsenic in groundwater in southeast Florida is common and future use as a potable supply is unlikely due to high levels of dissolved solids associated with the salt-water intrusion.

The definition of an EPC representing the groundwater plume area for this site is difficult and is complicated by the turbid sampling conditions noted in 1991. Although, all four wells were re-sampled in 1996, no removal actions affecting groundwater has occurred at the site. Therefore, the selection of data points best representing the arsenic concentrations in groundwater may need additional examination.

As previously stated in Section 2.7.1 of this document, the total unfiltered groundwater sampling results for arsenic at OU 5/Site WP-1 are as follows:

(1) Four samples were collected by G&M in 1991 from wells SP1-I-01 (92 ug/L), SPI-I-02 (60 ug/L), SP1-I-03 (19 ug/L) and SP1-MW-0001 (11 ug/L). All of the water sampling logs from the 1991 G&M data indicated turbid conditions;

(2) One sample was collected by MW in 1993 from SP1-I-02 (18.4 ug/L, 18.1 ug/L in the duplicate). MW also collected a filtered sample and duplicate from SP1-I-02 and detected concentrations of 16.1 ug/l and 16.7 ug/l, respectively; and

(3) Four samples were collected from all of the wells by OHM in 1996, SP1-I-01 (9.6 ug/L), SP1-I-02 (15.3 ug/L, 14.8 ug/L in the duplicate), SP1-I-03 (3.9 ug/L), and SP1-MW-0001 (7.8 ug/L).

Although the 1991 G&M data is not considered to represent actual groundwater conditions, in order to be conservative, a groundwater EPC of 26 ug/L was used in the risk calculations. This represents the arithmetic average of the four wells which are assumed to constitute the

concentrated area of the plume. Therefore, this EPC is considered a reasonably conservative number.

However, if the construction of the groundwater EPCs was changed based on the apparent sampling method discrepancies in 1991, risk estimates would be changed as well. The following information shows the range of risk estimates via exposure to groundwater for different EPCs:

If all data points (G&M, 1991; MW, 1993: and OHM, 1996) are considered -For the maximum concentration of 92ug/L, the ELCR = 2E-03 and the HI = 9E+00 For the average concentration of 26 ug/L, the ELCR = 2E-04 and the HI = 3E+00(26 ug/L represents the EPC used in the risk calculations presented in this assessment)

If only the MW, 1993, and OHM, 1996 data are considered -For the maximum concentration of 18 ug/L, the ELCR = 3E-04 and the HI = 2E+00For the average concentration of 11 ug/L, the ELCR = 2E-04 and the HI = 1E+00

If only the most recent sampling (OHM, 1996) is considered -For the maximum concentration of 15 ug/L, the ELCR = 3E-04 and the HI = 2E+00For the average concentration of 9 ug/L, the ELCR = 2E-04 and the HI = 1E+00

Therefore the potential risks associated with groundwater exposures could range from 2E-03 to 2E-04 for cancer and from 9E+00 to 1E+00 for noncancer risks. These values are above the USEPA remediation-based risk benchmarks for carcinogens (10^4 to 10^{-6}) and above the state of Florida's criterion of 1E-06. The hazard indices also exceed the risk benchmark of one.

A lesser, secondary contributor to the carcinogenic risk estimate is bis(2-ethylbexyl)phthalate (1.4% of the calculated risk). Bis(2-ethylhexyl)phthalate was detected in four of nine samples at concentrations ranging from $0.3 \mu g/L$ to $320 \mu g/L$. The maximum value for this chemical was used as the exposure point concentration in the risk calculations. The cancer risk (6E-06) is within the US EPA remediation-based risk benchmark, but above the state criteria.

However, these exceedances are of limited significance due to the unlikely use of this groundwater as a potable supply. The use of potable wells on-base has been replaced with the use of off-base wells because of high dissolved solids due to salt-water intrusion. Finally,

the future military land reuse of the site makes potable uses of groundwater even more remote.

2.7.5.4 Risks Associated With Exposure to Soils. Base Worker. Risks for a potential current base worker who regularly accesses OU-5/Site WP-1 are calculated in Table 2-31. The ELCR and HI are 2E-07 and 4E-03, respectively. These risk levels are below the USEPA remediation-based risk benchmarks.

Hypothetical Future Residents. The risks for hypothetical future residents exposed to onsite soils are calculated in Tables 2-32 (adult, 24-year exposure period) and 2-33 (young child, 6-year exposure period). For an adult, the ELCR and HI are 9E-06 and 2E-01, respectively. The ELCR and HI for the child are 2E-05 and 1E+00, respectively. Both the adult and child cancer risk estimates and the adult hazard index are below the USEPA remediation-based risk benchmarks. The hazard index for the child is equal to the benchmark of 1.0.

Hypothetical Future Construction Worker.Risks for future construction workers who would access OU-5/Site WP-1 are calculated in Table 2-34. The risks are estimated for construction worker exposure to surface soils via inhalation and ingestion routes of exposure. The ELCR and HI are 1E-06 and 5E-01, respectively. The cancer risk estimate is equal to the USEPA remediation-based risk benchmarks of IE-06, and HI is below the benchmarks of 1.

2.7.5.5 Lead. The USEPA has identified a 10 to 15 μ g/dL blood lead level as a range of potential concern for health effects in children (Federal Register, 1988b). The results from the IEUBK model using soil and groundwater data are listed in Table 2-35. The model predicted that 94% of children exposed to lead at concentrations at OU-5/Site WP-1 would have blood lead concentration below the 10 μ g/dL acceptable blood lead level. For this site, the model assumes the child is exposed to a concentration of 120 mg/kg of lead (represents the maximum concentration) in surface soil and 30 μ g/l of lead (represents the maximum concentration) in groundwater. The model used USEPA default exposure assumptions and used the EPCs calculated from the site data, conservatively assuming a lognormal distribution.

Although the maximum concentration of lead detected in unfiltered groundwater samples (30 μ g/l) is greater than the federal treatment technique level in drinking water (15 μ g/l), this concentration is not anticipated to be the delivered concentration in drinking water, as water

SOIL EXPOSURE DOSES AND RISK CALCULATIONS FOR A POTENTIAL CURRENT BASE WORKER AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Cs SexDo SexDd SExDi Calculated **Toxicity Values** Risk/HI (mg/kg) (mg/kg-day) (mg/kg-day) (mg/kg-day) Constituent **CANCER EFFECTS** CSFo CSFd CSFi BNAs 0.46 1.7E-08 1.1E-08 2.3E-13 7.3E+00 7.3E+00 6.1E+00 2.0E-07 Benzo(a)pyrene Pesticides/PCBs Chlordane Isomers 1.4 5.1E-08 3.2E-08 7.0E-13 1.3E+00 2.6E+00 1.3E+00 1.5E-07 4,4'-DDD 1.4 5.1E-08 3.2E-08 7.0E-13 2.4E-01 4.8E-01 2.4E-01 2.8E-08 Metals Arsenic 7.8 2.8E-07 1.8E-08 3.9E-12 1.5E+00 1.6E+00 1.5E+01 4.5E-07 5.1E-08 3.2E-09 7.0E-13 NAP NAP 6.3E+00 4.4E-12 Cadmium 1.4 Chromium (VI) 6.5E-07 4.1E-08 9.0E-12 4.1E+01 3.7E-10 18 NAP NAP 300 1.1E-05 6.9E-07 1.5E-10 NAP NAP 8.4E-01 1.3E-10 Nickel ELCR 8E-07 NON-CANCER EFFECTS RfDo RfDd RfDi **BNAs** Benzo(a)pyrene 0.46 4.7E-08 3.0E-08 6.4E-13 3.0E-02 1.5E-02 3.0E-02 3.5E-06 TRPHs (as n-nonane) 3322 3.4E-04 2.1E-04 4.7E-09 6.0E-01 3.0E-01 6.0E-01 1.3E-03 Pesticides/PCBs 1.4E-07 9.0E-08 2.0E-12 6.0E-05 3.0E-05 6.0E-05 5.4E-03 Chlordane Isomers 1.4 4,4'-DDD 1.4 1.4E-07 9.0E-08 2.0E-12 5.05-04 2.5E-04 5.0E-04 6.5E-04 Metals 6453 6.6E-04 4.2E-05 9.0E-09 1.0E+00 2.0E-01 1.0E+00 8.6E-04 Aluminum 7.9E-07 5.0E-08 1.1E-11 3.0E-04 2.9E-04 3.0E-04 2.8E-03 Arsenic 7.8 Cadmium 1.4 1.4E-07 9.0E-09 2.0E-12 1.0E-03 2.0E-04 1.0E-03 1.9E-04 Chromium (VI) 1.8E-06 4.8E-04 18 1.2E-07 2.5E-11 5.0E-03 1.0E-03 5.0E-03 Copper 160 1.6E-05 1.0E-06 2.2E-10 3.7E-02 7.4E-03 3.7E-02 5.8E-04 Manganese 1.9E-05 1.2E-06 2.6E-10 2.4E-02 4.8E-03 1.4E-05 1.0E-03 184 Mercury 0.2 2.0E-08 1.3E-09 2.8E-43 3.0E-04 6.0E-05 8.6E-05 8.9E-05 300 3.1E-05 1.9E-06 4.2E-10 2.0E-02 4.0E-03 2.0E-02 2.0E-03 Nickel Vanadium 12 1.2E-06 7.7E-08 1.7E-11 7.0E-03 1.4E-03 7.0E-03 2.3E-04 2E-02 HI

ELCR	Excess lifetime cancer risk.	CSFO	Cancer Slope Factor. Oral
HI	Hazard index (sum of the hazard quotients)	CSFd	Cancer Slope Factor. Dermal
Cs	Concentration of chemical in soil (mg/kg)	CSFi	Cancer Slope Factor. Inhalation
SExDo	Soil exposure does, oral route	RfDo	Reference Dose, Oral
SExDd	Soil exposure does, dermal route	RfDd	Reference Dose, Dermal
SExDo	Soil exposure dose, inhalation route	RfDi	Reference Does, Inhalation
NIAD	And		

NAP Not applicable, carcinogenic via inhalation pathway only

SOIL EXPOSURE DOSES AND RISK CALCULATIONS FOR A HYPOTHETICAL FUTURE ADULT RESIDENT AT SITE WP-1/OU-S, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Constituent	Cs (mg/kg)			SExDi (mg/kg-day)	1	Calculated Risk/HI		
CANCER EFFECTS					CSFo	CSFd	CSFi	
BNAs Benzo(a)pyrene	0.46	2.2E-07	6.8E-08	1.2E-11	7.3E+00	7.3E+00	6.1E+00	2.1E-06
Belizo(a)pyrelie	0.40	2.2E-07	0.8E-08	1.2E-11	7.3E+00	7.3E+00	0.1E+00	2.1E-00
Pesticides/PCBs								
Chlordane Isomers	1.4	6.6E-07	2.1E-07	3.6E-11	1.3E+00	2.6E+00	1.3E+00	1.4E-06
4,4'-DDD	1.4	6.6E-07	2.1E-07	3.6E-11	2.4E-01	4.8E-01	2.4E-01	2.6E-07
Metals								
Arsenic	7.8	3.7E-06	1.2E-07	2.0E-10	1.5E+00	1.6E+00	1.5E+01	5.7E-06
Cadmium	1.4	6.6E-07	2.1E-08	3.6E-11	NAP	NAP	6.3E+00	2.3E-10
Chromium (VI)	18	8.5E-06	2.7E-07	4.6E-10	NAP	NAP	4.1E+01	1.9E-08
Nickel	300	1.4E-04	4.5E-06	7.7E-09	NAP	NAP	8.4E-01	6.5E-09
						ELCR		9E-06
NON-CANCER EFFEC	CTS					<u></u>		
					RfDo	RfDd	RfDi	
<u>BNAs</u> Benzo(a)pyrene	0.46	6.3E-07	2.0E-07	3.5E-11	3.0E-02	1.5E-02	3.0E-02	3.4E-05
· / . ·								
TRPHs (as n-nonane)	3322	4.6E-03	1.4E-03	2.5E-07	6.0E-01	3.0E-01	6.0E-01	1.2E-02
Pesticides/PCBs								
Chlordane Isomers	1.4	1.9E-06	6.1E-07	1.1E-10	6.0E-05	3.0E-05	6.0E-05	5.2E-02
4,4'-DDD	1.4	1.9E-06	6.1E-07	1.1E-10	5.0E-04	2.5E-04	5.0E-04	6.3E-03
Metals								
Aluminum	6453	8.8E-03	2.8E-04	4.8E-07	1.0E+00	2.0E-01	1.0E+00	1.0E-02
Arsenic	7.8	1.1E-05	3.4E-07	5.9E-10	3.0E-04	2.9E-04	3.0E-04	3.7E-02
Cadmium	1.4	1.9E-06	6.1E-08	1.1E-10	1.0E-03	2.0E-04	1.0E-03	2.2E-03
Chromium (VI)	18	2.5E-05	7.8E-07	1.4E-09	5.0E-03	1.0E-03	5.0E-03	5.7E-03
Copper	160	2.2E-04	6.9E-06	1.2E-08	3.7E-02	7.4E-03	3.7E-02	6.9E-03
Manganese	184	2.5E-04	8.0E-06	1.4E-08	2.4E-02	4.8E-03	1.4E-05	1.3E-02
Mercury	0.2	2.7E-07	8.7E-09	1.5E-11	3.0E-04	6.0E-05	8.6E-05	1.1E-03
Nickel	300	4.1E-04	1.3E-05	2.3E-08	2.0E-02	4.0E-03	2.0E-02	2.4E-02
Vanadium	12	1.6E-05	5.2E-07	9.0E-10	7.0E-02	1.4E-03	7.0E-02	2.7E-03
						111		2E-01
						HI		∠E-01
				30E-	C 61		-1	
LCR Excess lifetime				CSFo		pe Factor, Ora		
	(sum of the haz			CSFd		pe Factor Der		
	of chemical in			CSFi		pe Factor, Inh	alation	
1	dose, Oral route			RfDo	Reference	· ·		
	dose, dermal ro			RfDd		Dose, Dermal		
1	dose, inhalatior	i route via inhalation pathy		RfDi	Reference	Dose, inhalati	on	

Not applicable, carcinogenic via inhalation pathway only NAP

SOIL EXPOSURE DOSES AND RISK CALCULATIONS FOR A HYPOTHETICAL FUTURE CHILD RESIDENT AT SITE WP-1/OU-S, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Best, Florida

Constituent	Cs (mg/kg)	SexDo (mg/kg-day)	SexDd (mg/kg-day)	SExDi (mg/kg-day)1	Foxicity Valu	es	Calculated Risk/HI
CANCER EFFECTS					CSFo	CSFd	CSFi	
BNAs								
Benzo(a)pyrene	0.46	5.0E-07	9.2E-08	1.4E-11	7.3E+00	7.3E+00	6.1E+00	4.4E-06
Pesticides/PCBs								
Chlordane Isomers 4,4'-DDD	1.4 1.4	1.5E-07 1.5E-07	2.8E-07 2.8E-07	4.2E-11 4.2E-11	1.3E+00 2.4E-01	2.6E+00 4.8E-01	1.3E+00 2.4E-01	2.7E-06 5.0E-07
Metals								
Arsenic	7.8	8.5E-06	1.6E-07	2.0E-10	1.5E+00	1.6E+00	1.5E+01	1.3E-05
Cadmium	1.4	1.5E-06	2.8E-08	4.2E-11	NAP	NAP	6.3E+00	2.7E-10
Chromium (VI)	18	2.0E-05	3.6E-07	5.4E-10	NAP	NAP	4.1E+01	2.2E-08
Nickel	300	3.3E-04	6.0E-06	9.0E-09	NAP	NAP	8.4E-01	7.6E-09
						ELCR		2E-05
NON-CANCER EFFEC	CTS				RfDo	RfDd	RfDi	
BNAs	0.44	5 0 5 0 4		4 45 40	0.07.00		0.05.00	
Benzo(a)pyrene	0.46	5.9E-06	1.1E-06	1.6E-10	3.0E-02	1.5E-02	3.0E-02	2.7E-04
TRPHs (as n-nonane)	3322	4.2E-02	7.8E-03	1.2E-06	6.0E-01	3.0E-01	6.0E-01	9.7E-02
Pesticides/PCBs								
Chlordane Isomers	1.4	1.8E-05	3.3E-06	4.9E-10	6.0E-05	3.0E-05	6.0E-05	4.1E-01
4,4'-DDD	1.4	1.8E-05	3.3E-06	4.9E-10	5.0E-04	2.5E-04	5.0E-04	4.9E-02
Metals		0.05.00			4 05 00		4 05 00	
Aluminum	6453	8.3E-02	1.5E-03	2.3E-06	1.0E+00	2.0E-01	1.0E+00	9.0E-02
Arsenic	7.8	1.0E-04	1.8E-06	2.7E-09	3.0E-04	2.9E-04	3.0E-04	3.4E-01
Cadmium	1.4	1.8E-05	3.3E-07	4.9E-10	1.0E-03	2.0E-04	1.0E-03	2.0E-02
Chromium (VI)	18	2.3E-04	4.2E-06	6.3E-09	5.0E-03	1.0E-03	5.0E-03	5.0E-02
Copper	160	2.0E-03	3.7E-05	5.6E-08	3.7E-02	7.4E-03	3.7E-02	6.0E-02
Manganese	184	2.4E-03	4.3E-05	6.5E-08	2.4E-02	4.8E-03	1.4E-05	1.1E-01
Mercury	0.2	2.6E-06	4.7E-08	7.0E-11	3.0E-04	6.0E-05	8.6E-05	9.3E-03
Nickel	300	3.8E-03	7.0E-05	1.1E-07	2.0E-02	4.0E-03	2.0E-02	2.1E-02
Vanadium	12	1.5E-04	2.8E-06	4.2E-09	7.0E-03	1.4E-03	7.0E-03	2.7E-03
						HI		1E+00
				225	a a -			
LCR Excess lifetime		• • • • •		CSFo	Cancer Slope Fa			
	(sum of the haz			CSFd	Cancer Slope Fa			
	of chemical in			CSFi	Cancer Slope Fa		n	
1	does, Oral rout			RfDo	Reference Dose,			
ExDd Soil exposure (dom dermal roi	ite	1	RfDd	Reference Dose	Dermal		

RfDd

RfDi

Reference Dose, Dermal

Reference Dose, Inhalation

SExDd Soil exposure dom dermal route

SExDi Soil exposure dose, inhalation route

Not applicable, carcinogenic via inhalation pathway only NAP

SOIL EXPOSURE DOSES AND RISK CALCULATIONS FOR A HYPOTHETICAL FUTURE CONSTRUCTION WORKER AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

	Cs	Surfa	ice Soil			Surface Soi
	Surface SExDo		SExDi	Toxicit	Calculated	
Contstituent	(mg/kg-day)	(mg/kg-day)	(mg/kg-day)		Risk/HI	
CANCER EFFECTS				CSFo	CSFi	
BNAs						
Benzo(a)pyrene	0.46	3.1E-08	3.5E-13	7.3E+00	6.1E+00	2.3E-07
Pesticides/PCBs						
Chlordane Isomers	1.4	9.4E-08	1.1E-12	1.3E+00	1.3E+00	1.2E-07
l,4'-DDD	1.4	9.4E-08	1.1E-12	2.4E-01	2.4E-01	2.3E-08
Metals						
Arsenic	7.8	5.2E-07	6.0E-12	1.5E+00	1.5E+01	7.9E-07
Cadmium	1.4	9.4E-08	1.1E-12	NAP	6.3E+00	6.8E-12
Chromium (VI)	18	1.2E-06	1.4E-11	NAP	4.1E+01	5.7E-10
Vickel	300	2.0E-05	2.3E-10	NAP	8.4E-01	1.9E-10
			ELCR	osure)	1E-06	
NON-CANCER EFFECTS						
				RfDo	RfDi	
<u>BNAs</u>						
Benzo(a)pyrene	0.46	2.2E-06	2.5E-11	3.0E-01	3.0E-01	7.2E-06
TRPHs (as n-nonane)	3322	1.6E-02	1.8E-07	6.0E-01	6.0E-01	2.6E-02
Pesticides/PCBs						
Chlordane Isomers	1.4	6.6E-06	7.5E-11	6.0E-05	6.0E-05	1.1E-01
4,4'-DDD	1.4	6.6E-06	7.5E-11	5.0E-04	5.0E-04	1.3E-02
Metals						
Aluminum	6453	3.0E-02	3.5E-07	1.0E+00	1.0E+00	3.0E-02
Arsenic	7.8	3.7E-05	4.2E-10	3.0E-04	3.0E-04	1.2E-01
Cadmium	1.4	6.6E-06	7.5E-11	1.0E-03	1.0E-03	6.6E-03
Chromium (VI)	18	8.5E-05	9.7E-10	2.0E-02	2.0E-02	4.2E-03
Copper	160	7.5E-04	8.6E-09	3.7E-02	3.7E-02	2.0E-02
laganese	184	8.6E-04	9.9E-09	2.4E-02	1.4E-02	3.7E-02
Aercury	0.2	9.4E-07	1.1E-11	3.0E-04	8.6E-04	3.1E-03
Nickel	300	1.4E-03	1.6E-08	2.0E-02	2.0E-02	7.0E-02
Vanadium	12	5.6E-05	6.5E-10	7.0E-03	7.0E-03	8.1E-03
			HI (S	Surface Soil Expos	sure)	5E-01

ELCR	Excess lifetime cancer risk.	CSFo	Cancer Slope Factor, Oral
HI	Hazard index (sum of the hazard quotients)	CSFd	Cancer Slope factor, Dermal
Cs	Concentration of chemical in soil (mg/kg)	CSFi	Cancer Slope Factor, Inhalation
SExDo	Soil exposure dose, oral route	RfDo	Reference Dose Oral
SExDd	Soil exposure dose, dermal route	RfDd	Reference Dose. Dermal
SExDi	Soil exposure dose, inhalation route	RfDi	Reference Dose, Inhalation
NAP	Not applicable, carcinogenic via inhalation pathway only		

MODELED BLOOD LEAD LEVELS IN HYPOTHETICAL CHILDREN (AGED 0 TO 6), SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

				Blood Lead Level ^b	
Study Site	Medium	Concentration ^a	Geometric Mean µg/dL	Percent Below 10 μg/dL	Percent Below 15 μg/dL
WP-1/OU-5	Soil Air ^c Groundwater	120 mg/kg negligible 30 μg/L	4.8	94.3	99.2

a Lesser of 95 percent UCL on the mean or maximum detected concentration.

b Calculated using the USEPA model (version 0.99d) (USEPA, 1994a).

c Air concentration = SPM x Cs x UC1 x UC2.

where:

- Cs Soil concentration (mg/kg).
- dL Deciliter.
- Kg Kilogram.
- m³ Cubic meter.
- mg Milligram.
- μg Microgram.
- SPM Suspended particulate matter (0.075 mg/m³) (Federal Register, 1988a).
- UCI Unit conversion 1 (10^{-6} kg/mg).
- UC2 Unit conversion 2 ($10^3 \mu g/mg$).

treatment prior to use would be expected to remove the metal in particulate form from water. Lead was detected in five of nine groundwater samples in concentrations ranging from 5.4 to 30 μ g/l. At present, the shallow groundwater is not used as a drinking water supply. Further, the use of the shallow groundwater in the future as a potable supply is highly improbable. Saltwater intrusion under the base has caused the replacement of on-base supply wells with off-base supply wells. So it is likely that saltwater intrusion would preclude the use of groundwater at OU-5/Site WP-1 for drinking water.

In addition, the low lead concentrations in surface soil (maximum value of 120 mg/kg) are not expected to present a significant contribution to blood lead levels in the base worker or construction worker (USEPA, 1994a). In both cases the potential routes of exposure to site soils (dermal, ingestion and inhalation), combined with the limited exposure duration for these receptors compared to the child receptor, minimize the expected dose received from the soil. Further, the IUEBK model assumes that the child is the most sensitive potential receptor. Based on this premise, if child blood lead levels do not exceed risk-based benchmarks, given the conditions at the site, then adult blood lead levels would also not be expected to exceed the risk-based benchmarks.

The levels of lead in the soil at OU-5/Site WP-1 are not unusual. Soil surveys have found soils within 25 meters of roadway to have from 30 to 2,000 mg/kg lead above background soil concentrations.

In summary, the lead concentrations in soils and groundwater are not expected to be of concern for the hypothetical future child resident, the current base worker, nor the future construction worker at OU-5/Site WP-1.

2.7.5.6 Total Site Risk. A summary of the total site risk estimates for OU-5/Site WP-1 is presented in this section. Table 2-36 includes the hazard indices and cancer estimates for all scenarios. Potential current total site risk is equivalent to the risk estimates calculated for a potential current on-site worker exposed to surface soil. This scenario is evaluated in Table 2-31 with an ELCR of 2E-07 and an HI of 0.004.

Total hypothetical future site risk for residential use was estimated by assuming that a future child resident could live on the site (6-year period), grow up, and continue to live there as an adult (24-year period), for a total residency period of 30 years. This total site risk is obtained by summing all of the residential exposures considered in the risk assessment: groundwater ingestion by an adult resident, soil exposure by a child resident (6-year period), and soil

TABLE 2.36

SUMMARY TABLE OF HAZARD INDICES AND CANCER RISKS FOR ALL SCENARIOS SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA Homestead Air Reserve Base, Florida

Scenario	Cancer Effects	Hazard Index
Groundwater Exposure for Future Adult Resident (Table 5-1, Section 5.1)	5E-04	3
Soil Exposure for Current Worker (Table 5-2, Section 5.2)	8E-07	0.02
Soil Exposure for Future Adults Resident (Table 5-3, Section 5.2)	9E-06	0.2
Soil Exposure for Future Child Resident (Table 5-4, Section 5.2)	2E-05	1
Surface Soil Exposure for Future Construction Worker (Table 5-5, Section 5.2)	1E-06	0.5
Total Risk to Future Resident (Child and Adult) (Tables 5-1, 5-3, and 5-4, Section 5.4)	5E-04	4

Note: all risks estimates are rounded to one significant figure.

exposure by adult (24-year period) residents. These scenarios are evaluated in Tables 2-30, 2-32, and 2-33. The combined risk across all pathways (groundwater, soils, surface water, and sediment) for a hypothetical future resident results in a total site excess lifetime cancer risk of 5E-04 and an HI of 4.

For the hypothetical construction worker, the total future site risk would be based on exposure to a combination of surface and subsurface soils. However, most soil layers at OU5/Site WP-1 are only one to two inches deep and the underlying layers are composed of limestone and bedrock. The construction worker scenario is evaluated in Table 2-34 with an excess lifetime cancer risk of 1E-06 and HI of 0.5.

Uncertainties in the Risk Assessment. The uncertainty associated with a risk estimate is primarily the combination of the uncertainties associated with the exposure estimates and the uncertainties in the toxicity evaluation. Additional uncertainty is inherent in environmental sampling, which itself introduces uncertainty, largely because of the potential for uneven distribution of chemicals in environmental media and the use of estimated data, such as J-qualified data. The rest of the discussion presented here focuses on the uncertainties in the exposure assessment and toxicity evaluation. It also presents a perspective on the overall effect of uncertainties on the risk estimates for OU-5/Site WP-1.

Risks associated with the future exposure pathways are only meaningful if the pathways are completed. For pathways, such as using shallow groundwater for drinking water, the probability is very low. It is expected that saltwater intrusion in this area already precludes the use of wells in this zone for potable supplies. Thus, use of groundwater at the site by the hypothetical future resident appears remote.

The exposure doses generally represent the reasonable maximum exposure that can be expected to occur. Most of the parameter values used in calculating the exposure, including the exposure point concentrations, were selected so that there was only a five to ten percent probability that the resulting exposure would be underestimated due to an error in an individual value. The analytical data used to estimate risks from groundwater contaminants probably do not lead to significant errors. These same conclusions can be made for soil samples. In cases where contaminated soil acts as a continuing source of groundwater contamination or where contaminants may be produced by biodegradation, the risk may be underestimated. Likewise, exposure doses are calculated based on the assumption that the current conditions would remain constant throughout the exposure period. If the source is

eliminated, natural attenuation processes will reduce constituent concentrations and the likelihood of exposure, thus reducing risks for the hypothetical future exposure scenarios.

Exposure point concentrations were calculated assuming a lognormal distribution of concentrations. The entire site was used as an exposure unit. Differing ranges of different receptors were not considered in the calculation of exposure point concentrations, if a receptor had a smaller range than the size of the site. However, the assumption of a lognormal distribution of data, and the use of maxima in many cases for the exposure point concentrations, means that the exposure point concentration used for COPCs in this document are conservative.

The most important uncertainties associated with the toxicity evaluation are the absence of a quantitative dose-response relationship for developmental and reproductive effects, and the absence of slope factors and reference doses for some chemicals of potential concern. The developmental and reproductive toxicity of the indicator chemicals has not been quantitatively accounted for in performing the risk assessment, because this dose-response relationship has generally not been characterized for the chemicals of potential concern. Another factor which could lead to an underestimate of the-total potential risk at the site is the lack of RfDs or SFs for several chemicals of potential concern. A review of the chemicals of potential concern without RfDs or SFs indicates the following: calcium, iron, and potassium are all essential nutrients and unless present in high doses, would have low toxic potential.

The slope factors are upper bound values for a fit of carcinogenicity data to a specific mathematical function (of which the function selected is in itself generally conservative with respect to other mathematical functions that fit the data equally well). Both the slope factors and reference doses incorporate safety factors when extrapolating from animal data to humans (including sensitive individuals), although animals may be more sensitive to a given compound than people. Slope factors and reference doses typically have safety factors of 100 to 1,000. There are some notable exceptions to this, especially when there is human toxicity data available. The uncertainty factor for the RfD for arsenic is 1, implying that the chronic dose necessary to cause a toxic effect is well known (IRIS, 1991). On the other hand, it is possible that some compounds (such as the VOCs) have minimum threshold doses associated with a carcinogenic response in humans that are not observed in animal experiments, due to the differences between rodent and human metabolism. If this is true, the slope factors would be overestimates by one or more orders of magnitude.

Toxicity values derived from the IRIS database system were accompanied with a qualitative description of their "strength of evidence" as determined by the CRAVE Work Group; the corresponding confidence in each toxicity value added to the uncertainty.

The evaluation of health effects associated with arsenic exposure is presently a very controversial area. While existing toxicological models attempt to relate exposure levels to quantifiable carcinogenic and toxic risk, there is no general consensus that all arsenic exposure has negative consequences or that a threshold level of effect does not exist. For example, recent research indicates that arsenic may be nutritionally essential for humans, a requirement that has been demonstrated for four other mammalian species. The presently available technology for estimating cancer risks to humans at low levels may not be appropriate for evaluating arsenic exposure risks.

For purposes of this risk assessment, it was assumed that all of the chromium detected in media at the site was in the hexavalent form. Under most natural conditions in soils and water containing reducing agents, the majority of chromium is in the trivalent oxidation state. Hexavalent chromium is more toxic than trivalent chromium. Thus, the risk estimates calculated in this report for potential exposure to chromium likely overestimate the actual risk.

The non-carcinogenic risks associated with potential lead exposure were not evaluated in a manner similar to other constituents in this risk assessment (for lack of an RfD). However, the integrated exposure biokinetic/uptake (IEUBK) model developed by the USEPA (version 0.99d) was used to predict blood lead levels in young children. Although any pharmacokinetic model is subject to uncertainties, the predicted blood lead levels (which indicate potential hypothetical future lead exposure at the site is not a major concern) are believed to be a reasonable estimate.

There is also considerable uncertainty associated with the toxicity of mixtures. For the most part, data on the toxicity of chemical mixtures are unavailable. Rather, toxicity studies generally are performed using a single chemical; such is the case for the carcinogenic PAHs. Chemicals present in a mixture can interact to yield a new chemical or one can interfere with the absorption, distribution, metabolism, or excretion of another. Chemicals may also act by the same mechanism at the same target organ or can act completely independently. The risk assessment assumes that toxicity is additive; the excess lifetime cancer risks and HQ were each summed across chemicals. This assumes that the mixture of chemicals present at OU5/Site WP-1 has neither synergistic nor antagonistic interactions and that all of the chemicals

have the same mechanism of action in the same target organ to produce the same toxic endpoints.

The toxicity of all chemicals in groundwater and soil has been assumed to be the same as the sum of the individual effects from each chemical. Neither synergistic nor antagonistic effects resulting from the interaction of the contaminants have been considered. In addition, transformation products with greater or less severe toxic effects than chemicals discussed herein may form, and are not accounted for in this evaluation.

Because of the arguments presented in this section, it can be stated that for those exposure scenarios which have been quantitatively evaluated and for which the most toxic and prevalent compounds at OU-5/Site WP-1 have reference doses and slope factors, this risk assessment is expected to be conservative, and the actual risks are expected to be less than those calculated.

2.7.5.7 Development of Remedial Goal Options. As risk characterization indicated that the risk benchmarks of 1E-04 for ELCR and 1 for HI were exceeded for certain of the scenarios considered, remedial goal options (RGOs) have been generated for OU-5/Site WP-1

Remedial Goal Options (RGOs) are outlined in this document to assess potential cleanup levels if site cleanup is necessary. RGOs were generated for surface soil for the base worker scenario and the construction worker scenario, and for potable use of groundwater. Residential RGOs for the residential scenario were not generated for soil as residential development is unlikely at the site, given the planned future military use of the site.

In the calculation of RGOs, concentrations for each individual chemical corresponding to ELCRs of 1E-04, 1E-05, and 1E-06 (for carcinogenic effects) and HQS of 3, 1, and 0.1 (for noncarcinogenic effects) are calculated for each chemical that has an ELCR exceeding IE-06 or a HQ exceeding 0.1. RGOs are specific to a certain risk scenario. RGOs were calculated, as per Florida DEP and USEPA Region IV guidance, by rearranging the site specific risk equations and solving for the concentration term for the target risk. RGOs were generated for those chemicals that were significant contributors to hazard, i.e. chemicals with an individual risk contribution of greater than 1E-06 or HQ of greater than 0.1. The corresponding state and federal guidance and results of the RGO calculations are presented in Tables 2-37 through 2-40.

RISK-BASED REMEDIAL GOAL OPTIONS AND FDEP SOIL TARGET LEVELS HYPOTHETICAL FUTURE CONSTRUCTION WORKER AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA SURFACE SOIL (mg/kg)

	SITE SPECIFIC REMEDIAL GOAL OPTIONS HAZARD INDEX			GO	ECIFIC RE DAL OPTIC INOGENIC	FDEP Soil Target Levels	
COMPOUNDS	0.1	1.0 3.0 1E-06 1E-05 1E-04			Based on an ELCR of 1E-06 / HI of 1		
<u>Pesticides/PCBs</u> Chlordane Isomers	1.3E+00	1.3E+01	3.8E+01	NAP	NAP	NAP	3.0E+00
<u>Metals</u> Arsenic	6.4E+00	6.4E+00	1.9E+02	NAP	NAP	NAP	3.1E+00

NAP = Not Applicable ELCR = Excess Lifetime Cancer Risk HI = Hazard Index

RISK-BASED REMEDIAL GOAL OPTIONS AND FDEP SOIL TARGET LEVELS HYPOTHETICAL FUTURE ADULT RESIDENT AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA SOIL (mg/kg)

	SITE SPECIFIC REMEDIAL GOAL OPTIONS HAZARD INDEX			GC	ECIFIC RE DAL OPTIC INOGENIC	FDEP Soil Target Levels	
COMPOUNDS	0.1	0.1 1.0 3.0			1E-05	1E-04	Based on an ELCR of 1E-06 / HI of 1
<u>BNAs</u> Benzo(a)pyrene Pesticides/PCBs	NAP	NAP	NAP	2.2E+01	2.2E+00	2.2E+01	1E-01
Chlordane Isomers	NAP	NAP	NAP	1.0E+00	1.0E+01	1.0E+02	5E-01
<u>Metals</u> Arsenic	NAP	NAP	NAP	1.4E+00	1.4E+01	1.4E+02	7E-01

NAP = Not Applicable ELCR = Excess Lifetime Cancer Risk HT = Hazard Index

RISK-BASED REMEDIAL GOAL OPTIONS AND FDEP SOIL TARGET LEVELS HYPOTHETICAL FUTURE CHILD RESIDENT AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA SOIL (mg/kg)

	GO	ECIFIC REN DAL OPTION AZARD INDI	NS	G	PECIFIC REN OAL OPTIO CINOGENIC	FDEP Soil Target Levels	
COMPOUNDS	0.1 1.0 3.0		1E-06	1E-05	1E-04	Based on an ELCR of 1E-06 / HI of 1	
<u>BNAs</u> Benzo(a)pyrene	NAP	NAP	NAP	1.1E-01	1.1E+00	1.1E+01	1.1E-01
Pesticides/PCBs Chlordane Isomers	3.4E-01	3.4E+00	1.0E+01	5.1E-01	5.1E+00	5.1E+01	5.0E-01
<u>Metals</u> Arsenic Manganese Nickel	2.3E+00 1.7E+02 1.4E+02	2.3E+01 1.7E+03 1.4E+03	6.9E+01 5.0E+03 4.3E+03	6.0E-01 NAP NAP	6.0E+00 NAP NAP	6.0E+01 NAP NAP	7.0E-01 3.7E+02 1.5E+03

NAP = Not Applicable ELCR = Excess Lifetime Cancer Risk HI = Hazard Index

RISK-BASED REMEDIAL GOAL OPTIONS HYPOTHETICAL FUTURE ADULT RESIDENT AT SITE WP-1/OU-5, ELECTROPLATING WASTE DISPOSAL AREA GROUNDWATER (mg/L)

	SITE SPECIFIC REMEDIAL GOAL OPTIONS HAZARD INDEX			SITE SPECIFIC REMEDIAL GOAL OPTIONS CARCINOGENIC RISK			EPA Maximum	Florida Drinking
COMPOUNDS	0.1	1.0	3.0	1E-06	1E-05	1E-04	Contaminant Level	Water Standard
<u>VOCs</u>								
Bromodichloromethane	NAP	NAP	NAP	1.4E-03	1.4E-02	1.4E-01		
BNAs Bis(2) Etherlihouryl) shthe late	0.072	0.72	2 10	6 1E 02	6 1E 02	C 1E 01		
Bis(2-Ethylhexyl)phthalate	0.073	0.73	2.19	6.1E-03	6.1E-02	6.1E-01		
<u>Metals</u>								
Aluminum	3.65	36.5	109.5	NAP	NAP	NAP	5E-02 to 2E-01 ^a	0.2^{b}
Arsenic	0.001	0.011	0.03	5.7E-05	5.7E-04	5.7E-03	0.05	
Chromium (VI)	0.018	0.18	0.55	NAP	NAP	NAP	1E-01	
Manganese	0.088	0.88	2.6	NAP	NAP	NAP	0.05^{a}	
Vanadium	0.026	0.26	0.77	NAP	NAP	NAP	NA	

NAP = Not Applicable

ELCR = Excess Lifetime Cancer Risk

HI = Hazard Index

^a USEPA Secondary Drinking Water Standard

^b Florida Secondary Drinking Water Standard

For residential groundwater exposure, bromodichloromethane, bis(2-ethylhexyl)phthalate, and arsenic had ELCRs exceeding 10^{-6} and aluminum, arsenic, chromium, manganese, and vanadium had hazard indices above 0.1. In surface soils, the adult resident scenario exposure the chemicals benzo(a)pyrene, chlordane, and arsenic had ELCRs exceeding 10^{-6} but no hazard indices above 0.1. For the child soil exposure, benzo(a)pyrene, chlordane, and arsenic exceeded 10^{-6} , and chlordane, arsenic, manganese, and nickel had hazard indices above 0.1. The base worker exposure to soils scenario had no chemicals exceeding an ELCR of 10^{-6} or a HI or 0.1. The construction worker scenario for exposure to surface soils had no chemicals exceeding an ELCR of 10^{-6} but chlordane and arsenic exceeded HIs of 0.1.

2.7.6 Ecological Risk Assessment

Conditions at OU-5/Site WP-1 provide little usable or preferred habitat for terrestrial species. Little vegetation is available for food or cover, and the shallow depth of soil to bedrock is expected to restrict the activities of burrowing animals. Base personnel activity at OU-5/Site WP-1 likely inhibit the activities of animals. Although avian species may potentially visit the site, it is highly unlikely that they would derive a significant portion of their diet from the limited resources available at OU-5/Site WP-1. Therefore, while constituent concentrations detected at OU-5/Site WP-1 might potentially represent ecotoxicological hazard, it is unlikely that terrestrial biota would inhabit or frequent the site.

While there is limited vegetative cover at the site, groundwater may be a potential source of exposure to plants via their root systems. Possible uptake would be modified by a variety of factors such as alkalinity of soils, organic content of soils, possible synergistic or antagonistic effects of multiple compounds, and the individual chemical and physical characteristics of the COCs in groundwater. Comparison with literature toxicity information indicates that the concentrations at OU-5/Site WP- 1 should not be significant.

Additionally, the potential for animals to contact groundwater constituents would be possible if groundwater were to recharge the drainage swales/canals. The maximum detected concentrations of several metals in groundwater were greater than Florida and Federal MCLs and associated surface water criteria protective of freshwater and saltwater aquatic species. These exceedences of surface water quality criteria or MCLs do not indicate the potential for adverse impacts to aquatic biota or terrestrial animals, respectively, due to the inability to sustain a resident aquatic population in the intermittent ditch and the limited expected use of these canals by wildlife.

Uncertainties in Ecological Risk. Although the effects of constituents on ecological receptors are a concern, it is difficult to predict if observed effects on individual populations will result in any real damage to the ecosystem. Populations are dynamic; therefore, information concerning the normal range of variability within the populations needs to be known. Sublethal effects, which may be very important to overall ecosystem health, are difficult to detect, and constituents present at low concentrations may not kill organisms directly but may greatly diminish their ability to survive and reproduce. Finally, it is important to note that constituent contamination is not the only manner in which humans impact ecosystems. Habitat destruction from development, agriculture and recreation are likely the major ways in which humans cause ecological impacts (Moriarty, 1988).

In summary, there is no evidence of significant use of the site as habitat by ecological receptors. Urbanization and base operations have already replaced this ecosystem and rendered its current use and likely future use as poor quality habitat for wildlife.

2.8 DESCRIPTION OF THE "NO FURTHER INVESTIGATION" ALTERNATIVE

Under its legal authorities, USEPA's primary responsibility at Superfund sites is to undertake remedial actions that achieves adequate protection of human health and the environment. Based on <u>soil</u> and <u>groundwater</u> analytical results collected to date, and the Interim Action remedial activities, the Electroplating Waste Disposal Area has been cleaned to industrial standards and therefore must be controlled in the future <u>with Land Use</u> <u>Controls</u> to safeguard both human health and the environment and be subject to <u>a five year review</u> to assure that the remedial actions do not allow for contaminant migration.

Homestead Air Force Base, Florida Operable Unit No. 5, Site WP-1 5 Electroplating Waste Disposal Area

Responsiveness Summary for the Record of Decision

RESPONSIVENESS SUMMARY

FOR THE

RECORD OF DECISION

The responsiveness summary serves three purposes. First, it provides regulators with information about the community preferences regarding both the remedial alternatives and general concerns about Operable Unit No. 5, Homestead ARB. Second, the responsiveness summary documents how public comments have been considered and integrated into the decision making, process. Third, it provided the USEPA with the opportunity to respond to each comment submitted by the public on the record.

The Remedial Investigation/Baseline Risk Assessment Report and Proposed Plan (PP) for Homestead ARB, OU-5/Site WP-1 were released to the public in October and December 1996, respectively. These documents were made available to the public in both the administrative record and an information repository maintained at the Air Force Base Conversion Agency OL-Y office

A public comment period was held from March 16, 1997 to April 14, 1997 as part of the community relations plan for OU-5/Site WP-1. Additionally, a public meeting was held on Thursday, March 19, 1997 at 7:00 pm at South Dade Senior High School. A Public Notice was published in the Miami Herald and South Dade News Leader on February 21, 1997. At this meeting the USAF, in coordination with USEPA Region 4, FDEP and Dade County Environmental Resources Management (DERM) will be prepared to discuss the Remedial Investigation, the Baseline Risk Assessment and the Preferred Alternative as described in the proposed Plan.

The Air Force Reserves did not receive any public comment either during the public comment period or at the public meeting.