## FLATHEAD SOLE

by

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## Executive Summary

The following changes have been made to this assessment relative to the November 2000 SAFE:

## Changes in the input data

1) 2001 total catch and discards through 6 October, 2001.
2) 2001 trawl survey biomass estimate and standard error.
3) 2001 length composition of the survey abundance.
4) 2000 length composition of the fishery catch.
5) 2000 age composition of the survey abundance.
6) Re-estimated age-length transition matrices, based upon survey length at age data.

## Model results

1) Estimated 3+ total biomass for 2001 is $612,337 \mathrm{t}$.
2) Projected female spawning biomass for 2002 is $262,402 \mathrm{t}$.
3) Recommended ABC for 2002 is $82,572 t$ based on an $F_{40 \%}(0.30)$ harvest level.
4) 2002 overfishing level is $100,770 t$ based on a $F_{35 \%}(0.38)$ harvest level.

The following summarizes our recommendations for flathead sole fisheries conservation measures.

| 2000 Assessment | 2001 Assessment |
| :--- | :--- |
| recommendations | recommendations |
| for the 2001 harvest | for the 2002 harvest |


| ABC | $83,964 \mathrm{t}$ | $82,572 \mathrm{t}$ |
| :--- | :--- | :--- |
| Overfishing | $102,485 \mathrm{t}$ | $100,770 \mathrm{t}$ |
| $\mathrm{F}_{\mathrm{ABC}}$ | $\mathrm{F}_{0.40}=0.30$ | $\mathrm{~F}_{0.40}=0.30$ |
| $\mathrm{~F}_{\text {overfishing }}$ | $\mathrm{F}_{0.35}=0.38$ | $\mathrm{~F}_{0.35}=0.38$ |

## Introduction

The flathead sole (Hippoglossoides elassodon) is distributed from northern California, off Point Reyes, northward along the west coast of North America and throughout Alaska (Hart 1973). In the northern part of its range it overlaps with the related and morphologically similar Bering Flounder (Hippoglossoides robustus) whose range extends north to the Chukchi Sea and into the western Bering Sea. The two species are very similar morphologically and at-sea identification is extremely difficult on the production schedule of the annual trawl survey. However, we feel there has been increasing accuracy during recent years. The growth and distribution differences between the species were described in Walters and Wilderbuer (1997), which illustrated the possible ramifications of combining information. For the purposes of this section, these two species are combined under the heading, Hippoglossoides sp.

Hippoglossoides sp. are managed as a unit stock in the Bering Sea and Aleutian Islands and were formerly a constituent of the "other flatfish" SAFE chapter. In June 1994, the Council requested the Plan Team to assign a separate ABC for flathead sole (Hippoglossoides sp.) in the BSAI, rather than combining flathead sole (Hippoglossoides sp.) with other flatfish as in past assessments. This request was based on a change in the directed fishing standards to allow increased retention of flatfish.

## Catch History

Prior to 1977, catches of Hippoglossoides sp. were combined with the species of the "other flatfish" category, which increased from around $25,000 \mathrm{t}$ in the 1960 s to a peak of $52,000 \mathrm{t}$ in 1971. At least part of this apparent increase was due to better species identification and reporting of catches in the 1970s. After 1971, catches declined to less than 20,000 t in 1975. Catches from 1977-89 averaged 5,286 t increasing to an annual average of 17,946 t from 1990-2000 (Table 1). The resource remains lightly harvested as the 2001 catch through 6 October is only $47 \%$ of the 2001 TAC of $34,000 \mathrm{t}$. The catch of flathead sole taken in research surveys from 1979-2001 are shown in Table 2. The catch locations by quarter for 2000 for flathead sole hauls (defined by flathead sole contributing at least $20 \%$ of the total catch) are shown in the Appendix.

Although flathead sole (Hippoglossoides sp.) receive a separate ABC and TAC they are still managed in the same PSC classification as rock sole and "other flatfish" and receive the same apportionments and seasonal allowances of bycaught prohibited species. In recent years, the flathead sole fishery has been closed prior to attainment of the TAC due to the bycatch of halibut (Table 3).

Substantial amounts of flathead sole are discarded overboard in various eastern Bering Sea target fisheries. Retained and discarded amounts are estimated for recent years using observer estimates of discard rate applied to the "blend" estimate of observer and industry reported retained catch (including flathead sole prior to 1995) (Table 4). A substantial portion of the discards in 2000 occurred in the Pacific cod, pollock, and rock sole fisheries.

## Data

Fishery Catch and Catch-at-age Data
This assessment uses fishery catches from 1977 through 6 October, 2001 (Table 1), and estimates of number caught by length group and sex for the years 1977-2000 (Tables 5-6).

## Survey Data

Because Hippoglossoides sp. are often taken incidentally in target fisheries for other species, CPUE from commercial fisheries seldom reflect trends in abundance for these species. It is therefore necessary to use research vessel survey data to assess the condition of these stocks.

Large-scale bottom trawl surveys of the Eastern Bering Sea continental shelf have been conducted in 1975 and 1979-2001 by NMFS. Information on length at age for flathead sole are available from the 1982, 1985, 1992, 1994, 1995 and 2000 surveys. An evaluation of growth rates was made for each of the survey years above, as these rates form the basis of transition matrices that convert numbers at age to numbers at length. This examination of growth rates was motivated by the finding of temporal variability in growth for some flatfish species, such as rock sole and Pacific halibut. The von Bertalanffy growth parameters ( $L_{i n f}, t_{0}$, and $K$ ) and $\sigma_{g}$ (the standard deviation of length at age on a log scale) were estimated by minimizing the negative log-likelihood of the observed flathead sole size at age data. The Akaike Information Criterion (AIC), defined as the negative log likelihood plus twice the number of parameters (Hilborn and Mangel 1997), was used to evaluate two models for each sex: a "combined" model in which the size at age data for all years was fit with a single growth model, and a "separate year" model, in which a separate growth curve was fit for each year. The results (shown below) suggest the use of the separate year model, although the difference in AIC between the two models is not large (especially for males).
"Separate Year" Model (24 parameters)

| Sex | - In like | AIC | - In like | AIC |
| :--- | ---: | ---: | ---: | ---: |
| Male (n=1148) | -1873.62 | -1825.62 | -1830.42 | -1822.42 |
| Female (n=1371) | -2302.87 | -2254.86 | -2253.20 | -2245.21 |

Although the AIC analysis suggests basing the transition matrix upon the most recent survey data, the estimated growth curves and observed size at age are similar between years up to age 10 (Figure 1). Beyond age 10, the growth curves began to deviate as the size at age data becomes more sparse in several years. The similarity in the size at age data led to the conclusion that there is no practical difference in growth rates. The transition matrices were estimated from the growth curve derived from the "combined" model; the coefficient of variation (CV) of growth was assumed to increase from 0.08 at age 3 to 0.10 at age 21 .

A length (cm) - weight (g) relationship of the form $W=a L^{b}$ was also fit to Hippoglossoides sp., with the estimated parameters of $a=0.003965$ and $b=3.25912$ applying to both sexes.

Survey estimates of total biomass and numbers by length group and sex for the years 1982-2001 are shown in Tables 7-9 and Figure 2. The survey gear changed after 1981, and as in previous assessments (Spencer et al. 1999) only the data from 1982 to the present are used. Since the early 1980s, estimated Hippoglossoides sp. biomass has approximately quadrupled to the 1997 peak estimate of $807,800 \mathrm{t}$ (Figure 2). However, estimated biomass declined to $394,822 \mathrm{t}$ in 1999 and $399,023 \mathrm{t}$ in 2000, respectively, and increased to $514,023 \mathrm{t}$ in the 2001 survey.

In summary, the data available for flathead sole are

1) Total catch weight, 1977-2001;
2) Proportional catch numbers by length group, 1977-2000;
3) Survey biomass and standard error, 1982-2001;
4) Survey age composition $1982,1985,1992,1995$, and 2000;
5) Proportional survey numbers by length group, 1982-2001.

## Analytical Approach

## Model Structure

The assessment model has a length-based formulation, which is underlaid by an age-based model. A transition matrix (TR) is used to convert the selectivity at length to selectivity at age, and to convert the predicted catch and numbers at age to catch and numbers at length.

An age-structured, split-sex population dynamics model was used to obtain estimates of recruitment, numbers at age, and catch at age for each sex. Population size in numbers at age $a$ in year $t$ for sex $s$ was modeled as

$$
N_{s, t, a}=N_{s, t-1, a-1} e^{-Z_{s, t-1, a-1}} \quad 4 \leq a<A, \quad 2 \leq t \leq T
$$

where $Z$ is the sum of the instantaneous fishing mortality rate $\left(F_{s, t, a}\right)$ and the natural mortality rate $\left(M_{s}\right), A$ is the maximum number of ages in the population, and $T$ is the terminal year of the analysis (2001). The numbers at age $A$ are a "pooled" group consisting of fish of age $A$ and older, and are estimated as

$$
N_{s, t, A}=N_{s, t-1, A-1} e^{-Z_{s, t-1, A-1}}+N_{t-1, A} e^{-Z_{s, t-1, A}}
$$

The total numbers of age 3 fish over all years are estimated as parameters in the model, and modeled with a lognormal distribution

$$
N_{t, 3}=e^{\left(\mu_{R}+v_{t}\right)}
$$

where $v$ is a time-variant deviation. The number of recruits is divided equally between males and females. The numbers at age in the first year are modeled to be in equilibrium with an historical catch of 1500 t , and requires estimation of a historic recruitment parameter $\left(R_{0}\right)$ and a historic fishing mortality rate $\left(f_{\text {hist }}\right)$.

The fishing mortality rate for a specific age and time $\left(F_{t, a}\right)$ is modeled as the product of a fishery age-specific selectivity function (fishasel) and a year-specific fullyselected fishing mortality rate $f$. The fully selected mortality rate is modeled as the product of a mean $\left(\mu_{f}\right)$ and a year-specific deviation $\left(\epsilon_{t}\right)$, thus $F_{t, a}$ is

$$
F_{t, a}=\text { fishasel }_{a} * f_{t} \equiv \text { fishasel }_{a} * e^{\left(\mu_{j}+\varepsilon_{l}\right)}
$$

The fishery selectivity at age is obtained from the selectivity at length and the transition matrix $\mathbf{T R}_{\mathbf{s}}$, where the transition matrix $\mathbf{T R}_{\mathbf{s}}$ indicates the proportion of each age (rows) in each length group (columns) for each sex; the sum across each age is equal to one. Because of growth differences between the sexes, there is a separate transition matrix and age -based selectivity vector for each sex; these matrices were computed as described above. The selectivity at age vector is computed from the fishery selectivity at length vector (fishlsel) as

$$
\text { fishasel }_{\mathrm{s}}=\mathbf{T R}_{\mathrm{s}} * \text { fishlsel }
$$

Finally, the selectivity at length vector, assumed identical for each sex, was modeled as

$$
\text { fishlsel }_{l}=\frac{1}{1+e^{- \text {slope }(l-\text { fift } y)}}
$$

where the parameter slope affects the steepness of the curve and the parameter fifty is the length at which ishlsel $_{l}$ equals 0.5 . There are 24 length bins ranging from 6 to 58 cm , and 19 age groups ranging from 3 to $21+$. The age- and length-based selectivity for the survey is modeled in a similar manner.

The mean numbers at age for each year and sex were computed as

$$
\bar{N}_{s, t, a}=N_{s, t, a} *\left(1-e^{-Z_{s, t, a}}\right) / Z_{s, t, a}
$$

The transition matrix and vector of mean numbers at age were used to compute the vector of mean numbers at length, by sex and year, as

$$
\overline{\mathbf{N L}}_{s, t}=\overline{\mathbf{N A}}_{s, t} * \mathbf{T R}_{s}^{\mathbf{T}}
$$

The vector of mean numbers at length was used to compute the catch as

$$
\begin{aligned}
& C_{l, s, t}=\overline{N L}_{l, s, t} * \text { fishlsel }_{l} * f_{t} \\
& \text { pred_cat }_{t}=\sum_{l, s} C_{l, s, t} * F W_{l, s}
\end{aligned}
$$

where $F W_{l, s}$ is the fishery weights by length and sex, and pred_cat is the predicted catch from the model. Similarly, the predicted survey biomass (pred_biom) is computed as

$$
\text { pred_biom }_{t}=\text { qsurv } \sum_{l, s}\left(\overline{N L}_{l, s, t} * \text { survlsel }_{l} * P W_{l, s}\right)
$$

where $P W_{l, s}$ is the population weight by length and sex, and qsurv is the trawl survey catchability.

Finally, age composition data are assumed to be unbiased, but with some aging error. The distribution of read ages around the "true" age is assumed to be normal with a variance of 0.02 times the true age, resulting in a coefficient of variation of 0.14 . The vector of mean number of fish by age available to the survey is multiplied by the aging error matrix in order to produce the observed survey age compositions.

## Parameters Estimated Independently

The parameters estimated independently include the age error matrix, the transition matrix, individual weight at length, natural mortality, and survey catchability ( $q \_s r v$ ). The age error matix was taken directly from the stock synthesis model used in previous assessments. The individual weights at age were obtained from trawl survey data, whereas qsurv and $M$ were fixed at 1.0 and 0.2 , respectfully, consistent with recent assessments.

## Parameters Estimated Conditionally

Parameter estimation is facilitated by comparing the model output to several observed quantities, such as the age compositions of the survey, length composition of the fishery and survey catches, the survey biomass, and the catch biomass. The general approach is to assume that deviations between model estimates and observed quantities are attributable to observation error and can be described with statistical distributions. Each data component provides a contribution to a total log-likelihood function, and parameter values that minimize the log-likelihood are selected.

The log-likelihood of the initial recruitments were modeled with a lognormal distribution

$$
\lambda_{1} \sum_{t} \frac{\left(v_{t}+\frac{\sigma^{2}}{2}\right)^{2}}{2 \sigma^{2}}+n \ln (\sigma)
$$

where $\sigma$ is a parameter representing the standard deviation of recruitment, respectively, on a $\log$ scale. The adjustment of adding $\sigma^{2} / 2$ to the deviation was made to correct for bias and produce deviations from the mean, rather than the median, recruitment.

The log-likelihoods of the fishery and survey age and length compositions were modeled with a multinomial distribution. The log of the multinomial function (excluding constant terms) for the fishery length composition data, with the addition of a term that scales the likelihood, is

$$
n_{f, s, t, l} \sum_{s, t, l} p_{f, s, t, l} \ln \left(\hat{p}_{f, s, t, l}\right)-p_{f, s, t, l} \ln \left(p_{f, s, l, l}\right)
$$

where $n$ is the number of fish aged, and $p_{f, s, t, l .}$ and $\hat{p}_{f, s, t, l}$ are the observed and estimated proportion at length in the fishery by sex, year and length. The likelihood for the age and length proportions in the survey, $p_{\text {surv,s,t,a}}$ and $p_{s u r v, s, t, l}$, respectively, follow similar equations.

The log-likelihood of the survey biomass was modeled with a lognormal distribution:

$$
\lambda_{2} \sum_{t}\left(\ln \left(\text { obs_biom }_{t}\right)-\ln \left(\text { pred_biom }_{t}\right)\right)^{2} / 2 c v_{t}^{2}
$$

where obs_biom $_{t}$ is the observed survey biomass at time $t, c v_{t}$ is the coefficient of variation of the survey biomass in year $t$, and $\lambda_{2}$ is a weighting factor.

The log-likelihood of the catch biomass was modeled with a lognormal distribution:

$$
\lambda_{3} \sum_{t}\left(\ln \left(o b s_{-} c a t_{t}\right)-\ln \left(\text { pred_cat } t_{t}\right)\right)^{2}
$$

where obs_cat ${ }_{t}$ and pred_cat $t_{t}$ are the observed and predicted catch. Because the catch biomass is generally thought to be observed with higher precision that other variables, $\lambda_{3}$ was given a very high weight so as to fit the catch biomass nearly exactly. This can be accomplished by varying the $F$ levels, and the deviations in $F$ are not included in the overall likelihood function. The overall negative log-likelihood function (excluding the catch component) is

$$
\begin{aligned}
& \lambda_{1}\left(\sum_{t}\left(\frac{\left(v_{t}+\sigma^{2} / 2\right)^{2}}{2 \sigma^{2}}\right)+n \ln (\sigma)\right)+ \\
& \lambda_{2} \sum_{t}\left(\ln \left(o b s_{-} \text {biom }_{t}\right)-\ln \left(\text { pred_biom }_{t}\right)\right)^{2} / 2 * c v_{t}^{2}+ \\
& n_{f, s, t, l} \sum_{s, t, l} p_{f, s, t, l} \ln \left(\hat{p}_{f, s, t, l}\right)-p_{f, s, t, l} \ln \left(p_{f, s, t, l}\right)+ \\
& n_{s u r v, s, t, a} \sum_{s, t, a} p_{s u r r_{, s, t, a}} \ln \left(\hat{p}_{s u r v, s, t, a}\right)-p_{s u r v, s, t, a} \ln \left(p_{s u r v, s, t, a}\right)+ \\
& n_{s u r v, s, t, l} \sum_{s, t, a} p_{s u r v, s, t, l} \ln \left(\hat{p}_{s u r v, s, t, l}\right)-p_{s u r v, s, t, l} \ln \left(p_{s u r v, s, t, l}\right)+ \\
& \lambda_{3} \sum_{t}\left(\ln \left(o b s_{-} c a t_{t}\right)-\ln \left(\text { pred_c }_{-} c a t_{t}\right)\right)^{2}
\end{aligned}
$$

For the model run in this analysis, $\lambda_{1}, \lambda_{2}$, and $\lambda_{3}$ were assigned weights of 1,2 , and 500 , respectively, and $n$ was set to 200. The likelihood function was minimized by varying the following parameters:

| Parameter type | Number |
| :--- | ---: |
| 1) fishing mortality mean $\left(\mu_{f}\right)$ | 1 |
| 2) fishing mortality deviations $\left(\epsilon_{t}\right)$ | 25 |
| 3) recruitment mean $\left(\mu_{r}\right)$ | 1 |
| 4) recruitment standard deviation $(\sigma)$ | 1 |
| 4) recruitment deviations $\left(v_{t}\right)$ | 25 |
| 5) historic recruitment $\left(R_{0}\right)$ | 1 |
| 6) historic fishing mortality $\left(f_{\text {hist }}\right)$ | 1 |
| 7) fishery selectivity parameters | 2 |
| 8) survey selectivity parameters | 2 |
| Total parameters | 59 |

## Model Results

The model results show that estimated total biomass (ages $3+$ ) increased from a low of $227,194 \mathrm{t}$ in 1977 to a peak of $897,867 \mathrm{t}$ in 1991 (Figure 2, Table 10). Since 1991, estimated total biomass has declined to an estimated value of $612,337 \mathrm{t}$ for 2001. Female spawning biomass shows a similar trend, although the peak value ( $388,974 \mathrm{t}$ ) occurred in 1995 (Figure 3). The estimated survey biomass shows an increase from 1982 to the peak level of $567,695 \mathrm{t}$ in 1993, and a subsequent decline to $418,152 \mathrm{t}$ in 2001 (Figure 4). The model fits the survey biomass time-series well during the period of increasing biomass, but provides a poor fit to the 1994, 1997 and 1998 estimates, when it indicates a population decline while survey biomass estimates remain high and relatively stable. The
continued trend of declining estimated biomass since the early 1990s results in the estimated 1999 and 2000 survey biomass matching the observed biomass more closely than the estimated 2001 biomass matches the observed biomass (Figure 4). The model provided a good fit to the survey size compositions for the past 10 years for males and females as shown Figures 5 and 6. Reasonable fits also resulted for fishery size composition observations (Figures 7 and 8 ) and the survey age composition (Figure 9).

The changes in stock biomass are primarily a function of recruitment, as fishing pressure has been relatively light. The fully selected fishing mortality estimates remain small, and have averaged 0.05 from 1990 to 2001 (Figure 10), and the fishery shows little selectivity for flathead sole less that 30 cm (Figure 11). Estimated recruitment at age 3 has generally been higher during the early portion of the data series, averaging $8.8 \times 10^{8}$ for the 1975-1988 year classes, and $4.1 \times 10^{8}$ for the 1989-98 year classes (Figure 12). The scatterplot of stock and recruitment data reveals a decreasing trend in recruitment with an increasing trend in spawner biomass (Figure 13). The survey size composition from 1994-2000 indicates that the proportion of fish at smaller sizes is reduced from the high recruitment years of the 1980s, leading to the decline in estimated biomass.

The extent to which the density-dependence observed in the scatterplot of spawerrecruit data (Figure 13) is affected by environmental conditions is unresolved. For example, a series of high spawner stock biomasses and low recruitments were observed for the post-1988 year classes, coinciding with changes in the environmental indices such as the Aleutian low pressure index (Hare and Mantua 2000). Stock-recruitment analyses that consider this environmental variability are a priority for future flathead sole research.

## Projections and Harvest Alternatives

The reference fishing mortality rate for flathead sole is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Estimates of $F_{0.40}$, $F_{0.35}$, and $S P R_{0.40}$ were obtained from a spawner-per-recruit analysis. Assuming that the average recruitment from the 1977-1998 year classes estimated in this assessment represents a reliable estimate of equilibrium recruitment, then an estimate of $B_{0.40}$ is calculated as the product of $S P R_{0.40}$ * equilibrium recruits, and this quantity is $141,930 \mathrm{t}$. The year 2002 spawning stock biomass is estimated as $262,402 \mathrm{t}$. Since reliable estimates of the 2002 spawning biomass $(B), B_{0.40}, F_{0.40}$, and $F_{0.35}$ exist and $B>B_{0.40}$ ( $262,402 \mathrm{t}>141,930 \mathrm{t}$ ), flathead sole reference fishing mortality is defined in tier 3a. For this tier, $\mathrm{F}_{\mathrm{ABC}}$ is constrained to be $\leq F_{0.40}$, and $F_{O F L}$ is defined to be $F_{0.35}$. The values of these quantities are

$$
\begin{aligned}
& 2002 \text { SSB estimate }(\mathrm{B})=262,402 \mathrm{t} \\
& B_{0.40}= \\
& F_{0.40}=141,930 \mathrm{t} \\
& F_{A B C} \leq \\
& F_{0.35}= \\
& F_{O F L}=0.300 \\
& 0.375 \\
&
\end{aligned}
$$

The estimated catch level for year 2002 associated with the overfishing level of $F$ $=0.375$ is $100,770 \mathrm{t}$. Because the flathead sole stock has not been overfished in recent years and the stock biomass is relatively high, it is not recommended to adjust $F_{A B C}$ downward from it upper bound; thus, the year 2002 recommended ABC associated with $F_{A B C}$ of 0.300 is $82,572 \mathrm{t}$.

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2001 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2002 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2001. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2002, are as follow (" $\max F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2002 recommended in the assessment to the $\max F_{A B C}$ for 2002. (Rationale: When $F_{A B C}$ is set at a value below max $F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 1996-2000 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

The recommended $F_{A B C}$ and the maximum $F_{A B C}$ are equivalent in this assessment, and five-year projections of the mean harvest and spawning stock biomass for the remaining four scenarios are shown in Tables 11.

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether the flathead sole stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follows (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2002, then the stock is not overfished.)

Scenario 7: In 2002 and 2003, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2004 under this scenario, then the stock is not approaching an overfished condition.)

The results of these two scenarios indicate that the flathead sole are neither overfished or approaching an overfished condition. With regard to assessing the current stock level, the expected stock size in the year 2002 of scenario 6 is 2.09 times its $B_{35 \%}$ value of $124,189 \mathrm{t}$. With regard to whether the stock is likely to be in an overfished condition in the near future, the expected stock size in the year 2004 of scenario 7 is 1.29 times its $B_{35 \%}$ value.

## Other considerations

Trophic studies indicate that flathead sole feed mainly on ophiuroids, tanner crab, osmerids, bivalves and polychaetes. Groundfish predators include Pacific cod, Pacific halibut, arrowtooth flounder and also cannibalism by large flathead sole, mostly on fish less than 20 cm standard length.

## Summary

In summary, several quantities pertinent to the management of the flathead sole are listed below.

| Quantity | Value |
| :--- | :--- |
| $M$ | 0.20 |
| Year 2001 Spawning stock biomass | $262,402 \mathrm{t}$ |
| $F_{\text {OFL }}$ | 0.375 |
| Maximum $F_{A B C}$ | 0.300 |
| Recommended $F_{A B C}$ | 0.300 |
| OFL | $100,770 \mathrm{t}$ |
| Recommended ABC | $82,572 \mathrm{t}$ |

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Table 1. Harvest ( t ) of flathead sole from 1977-2001

| Year | Catch <br> Biomass |
| :--- | :---: |
| 1977 | 7909 |
| 1978 | 6957 |
| 1979 | 4351 |
| 1980 | 5247 |
| 1981 | 5218 |
| 1982 | 4509 |
| 1983 | 5240 |
| 1984 | 4458 |
| 1985 | 5636 |
| 1986 | 5208 |
| 1987 | 3595 |
| 1988 | 6783 |
| 1989 | 3604 |
| 1990 | 20245 |
| 1991 | 15602 |
| 1992 | 14239 |
| 1993 | 13664 |
| 1994 | 18455 |
| 1995 | 14707 |
| 1996 | 17344 |
| 1997 | 20704 |
| 1998 | 24397 |
| 1999 | 17842 |
| 2000 | 19983 |
| 2001 | $16131^{*}$ |
| *NFS Regional | Office Report through October 6, 2001 |

Table 2. Research catches ( t ) of flathead sole in the BSAI area from 1979 to 2001.

| Year | Research Catch $(\mathrm{t})$ |
| :--- | :---: |
| 1979 | 11.85 |
| 1980 | 6.19 |
| 1981 | 11.23 |
| 1982 | 20.36 |
| 1983 | 13.86 |
| 1984 | 13.51 |
| 1985 | 44.83 |
| 1986 | 13.79 |
| 1987 | 12.97 |
| 1988 | 29.86 |
| 1989 | 24.60 |
| 1990 | 26.76 |
| 1991 | 35.92 |
| 1992 | 18.92 |
| 1993 | 21.86 |
| 1994 | 30.23 |
| 1995 | 26.52 |
| 1996 | 20.87 |
| 1997 | 30.31 |
| 1998 | 23.02 |
| 1999 | 16.82 |
| 2000 | 19.09 |
| 001 | 18.50 |

Table 3. Restrictions on the flathead sole fishery from 1994 to 2001 in the Bering Sea - Aleutian Islands management area. Unless otherwise indicated, the closures were applied to the entire BSAI management area. Zone 1 consists of areas $508,509,512$, and 516 , whereas zone 2 consists of areas 513, 517, and 521.

| Year | Dates | Bycatch Closure |
| :---: | :---: | :---: |
| 1994 | 2/28-12/31 | Red King crab cap (Zone 1 closed) |
|  | $5 / 7-12 / 31$ | Bairdi Tannner crab (Zone 2 closed) |
|  | 7/5-12/31 | Annual halibut allowance |
| 1995 | 2/21-3/30 | First Seasonal halibut cap |
|  | 4/17-7/1 | Second seasonal halibut cap |
|  | 8/1-12/31 | Annual halibut allowance |
| 1996 | 2/26-4/1 | First Seasonal halibut cap |
|  | 4/13-7/1 | Second seasonal halibut cap |
|  | 7/31-12/31 | Annual halibut allowance |
| 1997 | 2/20-4/1 | First Seasonal halibut cap |
|  | 4/12-7/1 | Second seasonal halibut cap |
|  | 7/25-12/31 | Annual halibut allowance |
| 1998 | 3/5-3/30 | First Seasonal halibut cap |
|  | 4/21-7/1 | Second seasonal halibut cap |
|  | 8/16-12/31 | Annual halibut allowance |
| 1999 | 2/26-3/30 | First Seasonal halibut cap |
|  | 4/27-7/04 | Second seasonal halibut cap |
|  | 8/31-12/31 | Annual halibut allowance |
| 2000 | 3/4-3/31 | First Seasonal halibut cap |
|  | 4/30-7/03 | Second seasonal halibut cap |
|  | 8/25-12/31 | Annual halibut allowance |
| 2001 | 3/20-3/31 | First Seasonal halibut cap |
|  | 4/27-7/01 | Second seasonal halibut cap |
|  | 8/24-12/31 | Annual halibut allowance |

Table 4. Total retained and discarded flathead sole, 1995-2001.

| Year | Total Catch | Retained | Discarded | Percent Retained |
| :--- | :--- | :--- | :--- | :--- |
| 1995 | 14707 | 7521 | 7186 | 51 |
| 1996 | 17344 | 8964 | 8380 | 52 |
| 1997 | 20704 | 10871 | 9833 | 53 |
| 1998 | 24397 | 17208 | 7189 | 70 |
| 1999 | 17892 | 13282 | 4610 | 74 |
| 2000 | 19983 | 14730 | 5253 | 74 |
| $2001^{*}$ | 16131 | 13204 | 2927 | 82 |
| * |  |  |  |  |

*NMFS regional office report through October 6, 2000


























Table 7. Estimated biomass of flathead sole from the EBS and Aleutian Islands Trawl survey.

| Year | Area | Biomass Estimate |
| :---: | :---: | :---: |
| 1975 | EBS | 100,700 |
| 1979 | EBS | 104,900 |
| 1980 | EBS | 117,500 |
|  | Aleut. | 3,300 |
| 1981 | EBS | 162,900 |
| 1982 | EBS | 191,988 |
| 1983 | EBS | 269,419 |
|  | Aleut. | 1,500 |
| 1984 | EBS | 341,697 |
| 1985 | EBS | 276,350 |
| 1986 | EBS | 357,951 |
|  | Aleut. | 9,000 |
| 1987 | EBS | 394,758 |
| 1988 | EBS | 572,805 |
| 1989 | EBS | 536,433 |
| 1990 | EBS | 628,235 |
| 1991 | EBS | 544,893 |
|  | Aleut. | 6,885 |
| 1992 | EBS | 651,384 |
| 1993 | EBS | 610,259 |
| 1994 | EBS | 726,212 |
|  | Aleut. | 9,917 |
| 1995 | EBS | 593,412 |
| 1996 | EBS | 616,373 |
| 1997 | EBS | 807,825 |
|  | Aleut. | 11,540 |
| 1998 | EBS | 692,234 |
| 1999 | EBS | 394,822 |
| 2000 | EBS | 399,298 |
| 2000 | Aleut | 8,970 |
| $\underline{2001}$ | EBS | 514,023 |

























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Table 10. Estimated total biomass (ages $3+$ ), female spawner biomass, and recruitment (age 3), with comparison to the 2000 SAFE estimates

| Spawning stock <br> biomass $(\mathrm{t})$ | Total biomass $(\mathrm{t})$ |
| :--- | :--- | | Recruitment |
| :--- |
| (thousands) |


| Assessment |  |  |  | Assessment |  |
| ---: | ---: | ---: | ---: | ---: | ---: |
|  | 2001 |  | 2001 | 2000 |  |

Table 11. Projections of spawning biomass, catch, fishing mortality rate, and catch for each of the several scenarios. The values of $\mathrm{B}_{40 \%}$ and $\mathrm{B}_{35 \%}$ are $141,930 \mathrm{t}$ and $124,189 \mathrm{t}$, respectively.

| Sp. Biomass | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 262402 | 262402 | 289671 | 289671 | 289671 | 289671 | 289671 |
| 2002 | 203715 | 203715 | 267432 | 270655 | 272566 | 259925 | 262402 |
| 2003 | 162124 | 162124 | 230638 | 249648 | 261622 | 191595 | 203714 |
| 2004 | 133399 | 133399 | 201317 | 231711 | 252014 | 145954 | 160648 |
| 2005 | 115612 | 115612 | 178338 | 216703 | 243835 | 116850 | 126341 |
| 2006 | 108670 | 108670 | 160787 | 204386 | 236959 | 102141 | 107186 |
| 2007 | 110235 | 110235 | 150147 | 196690 | 233311 | 97460.5 | 100244 |
| 2008 | 115817 | 115817 | 147711 | 195913 | 235735 | 100537 | 102043 |
| 2009 | 122896 | 122896 | 150220 | 198675 | 240566 | 107050 | 107777 |
| 2010 | 129552 | 129552 | 156060 | 204900 | 248731 | 114370 | 114660 |
| 2011 | 134666 | 134666 | 163100 | 212766 | 258558 | 120706 | 120772 |
| 2012 | 138245 | 138245 | 169739 | 220628 | 268396 | 125140 | 125109 |
| 2013 | 140841 | 140841 | 175379 | 227527 | 276974 | 127929 | 127867 |
| 2014 | 138292 | 138292 | 180459 | 234373 | 285804 | 129711 | 129651 |
| F | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 |
| 2001 | 0.0509811 | 0.0509811 | 0.0509809 | 0.0509806 | 0.0509809 | 0.0509812 | 0.0509839 |
| 2002 | 0.300318 | 0.300318 | 0.150159 | 0.05555 | 0 | 0.375421 | 0.300318 |
| 2003 | 0.300318 | 0.300318 | 0.150159 | 0.05555 | 0 | 0.375421 | 0.300318 |
| 2004 | 0.300318 | 0.300318 | 0.150159 | 0.05555 | 0 | 0.375421 | 0.375421 |
| 2005 | 0.281317 | 0.281317 | 0.150159 | 0.05555 | 0 | 0.305591 | 0.332019 |
| 2006 | 0.241699 | 0.241699 | 0.150159 | 0.05555 | 0 | 0.264636 | 0.278682 |
| 2007 | 0.226236 | 0.226236 | 0.149883 | 0.05555 | 0 | 0.251603 | 0.259352 |
| 2008 | 0.229455 | 0.229455 | 0.146941 | 0.05555 | 0 | 0.260112 | 0.264287 |
| 2009 | 0.240518 | 0.240518 | 0.145973 | 0.05555 | 0 | 0.277465 | 0.279452 |
| 2010 | 0.253569 | 0.253569 | 0.146648 | 0.05555 | 0 | 0.296154 | 0.296935 |
| 2011 | 0.264354 | 0.264354 | 0.14761 | 0.05555 | 0 | 0.311533 | 0.311717 |
| 2012 | 0.271759 | 0.271759 | 0.148296 | 0.05555 | 0 | 0.32145 | 0.321396 |
| 2013 | 0.276264 | 0.276264 | 0.148838 | 0.05555 | 0 | 0.327417 | 0.327292 |
| 2014 | 0.279302 | 0.279302 | 0.149298 | 0.05555 | 0 | 0.33109 | 0.330969 |
| Catch | Scenario 1 | Scenario 2 | Scenario 3 | Scenario 4 | Scenario 5 | Scenario 6 | Scenario 7 |
| 2001 | 16131.4 | 16131.4 | 16131.4 | 16131.3 | 16131.4 | 16131.5 | 16132.3 |
| 2002 | 82572 | 82572 | 43360 | 16554.7 | 0 | 100770 | 82571.9 |
| 2003 | 65488.3 | 65488.3 | 37872 | 15388 | 0 | 76238.6 | 65488.2 |
| 2004 | 52910.9 | 52910.9 | 33332.6 | 14344.5 | 0 | 59173.2 | 64635.3 |
| 2005 | 41273.5 | 41273.5 | 29643.9 | 13428.4 | 0 | 39452.5 | 45897.3 |
| 2006 | 31113.1 | 31113.1 | 26814.5 | 12677 | 0 | 30173.4 | 33219.4 |
| 2007 | 27054.2 | 27054.2 | 24753.7 | 12095.1 | 0 | 26943.9 | 28563.4 |
| 2008 | 26860.8 | 26860.8 | 23257.8 | 11783.6 | 0 | 27608 | 28517.6 |
| 2009 | 28470.4 | 28470.4 | 22779.7 | 11646 | 0 | 30110 | 30592.5 |
| 2010 | 31044.1 | 31044.1 | 23224.9 | 11771.9 | 0 | 33466.4 | 33696.3 |
| 2011 | 33658.3 | 33658.3 | 24089.3 | 12066.5 | 0 | 36698.9 | 36785.6 |
| 2012 | 35886.4 | 35886.4 | 25082.1 | 12451.8 | 0 | 39225.1 | 39240.3 |
| 2013 | 37554.8 | 37554.8 | 26030.5 | 12839 | 0 | 41005.3 | 40989.8 |
| 2014 | 38858.1 | 38858.1 | 26950.8 | 13253.4 | 0 | 42269.6 | 42246.4 |



Figure 1. Estimated growth curves and mean length at age for Bering Sea/Aleutian Islands flathead sole.


Figure 2. Estimated survey biomass and $95 \% \mathrm{Cls}$


Figure 3. Estimated total (solid line) and spawner (dashed line) biomass


Figure 4. Estimated survey biomass of flathead sole with observed survey biomass


Figure 5 . Female survey length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 5. Female survey length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 6. Male survey length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 6. Male survey length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 7. Female fishery length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 7. Female fishery length composition by year (solid line = observed, dotted line $=$ predicted)


Figure 7. Female fishery length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 8. Male fishery length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 8. Male fishery length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 8. Male fishery length composition by year (solid line $=$ observed, dotted line $=$ predicted)


Figure 9. Survey age composition by year (solid line $=$ observed,
dotted line = predicted)


Figure 10. Estimated fishing mortality rate of flathead sole


Figure 11. Estimated fishery (solid line) and survey (dashed line) selectivity curve by length


Figure 12. Estimated recruitment (age 3) of flathead sole


Figore 13. Estimated female SSB and recruitment of flathead sole, labeled by year class, with a fitted Ricker corve (solid line).
The replacement line is based on an $\mathrm{F} 40 \%$ value of 0.30

## Appendix

Figures showing the distribution of flathead sole hauls sampled by fishery observers in 2000, by quarters. Flathead sole hauls are defined as having at least $20 \%$ of the total catch consisting of flathead sol





