# -SEAMAP Reef Fish Survey of Offshore Banks: <br> Yearly Indices of Abundance for Vermilion Snapper, Greater Amberjack, and Gray Triggerfish 

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## INTRODUCTION

The objective of the annual Southeast Area Monitoring and Assessment Program (SEAMAP) offshore reef fish survey is to provide an index of the relative abundances of fish species associated with topographic features (banks, ledges) located on the continental shelf of the Gulf of Mexico (Gulf) in the area from Brownsville, TX to the Dry Tortugas, FL (Figure 1). The total reef area surveyed is approximately $1771 \mathrm{~km}^{2} ; 1244 \mathrm{~km}^{2}$ in the eastern and $527 \mathrm{~km}^{2}$ in the western Gulf. The offshore reef fish survey was initiated in 1992, with sampling conducted during the months of May to August from 1992-1997, and in 2001-2002. No surveys were conduced from 1998 to 2000 and in 2003. A survey was conducted in 2004, however data edits were not completed for SEDAR 9, and will be made available prior to the stock assessments. The 2001 survey was abbreviated due to ship scheduling.


Figure 1. Gulf of Mexico shelf-edge banks sampled during SEAMAP offshore reef fish survey with sample blocks.

## SAMPLE DESIGN

The survey area is large. Therefore, a two-stage sampling design is used to minimize travel times between sample stations. The first-stage or primary sampling units (PSUs) are blocks 10 minutes of latitude by 10 minutes of longitude (Figures 2 and 3). The first-stage units are selected by stratified random sampling. The blocks were stratified, with strata defined by geographic region (4 regions: South Florida, Northeast Gulf, Louisians-Texas Shelf, and

SouthTexas), and by reef habitat area (Blocks $\leq 20 \mathrm{~km}^{2}$ reef, Block $>20 \mathrm{~km}^{2}$ reef). There are a total of 7 strata. The ultimate sample sites (second stage units) within a block are selected randomly. However, stratum 1 (South Florida, small blocks) and stratum 7 (S. Texas, small blocks) were not consistently sampled. So, these were dropped from annual indices.

## GEAR

The SEAMAP reef fish survey currently employs four Sony VX2000 DCR digital camcorders mounted in Gates PD150M underwater housings. The housings are rated to a maximum depth of 150 meters. The four Sony VX2000 camcorders are mounted orthogonally and a height of 30 cm above the bottom of the pod. A chevron (or arrow) fish trap with 1.5 -inch vinyl-clad mesh is used to capture fish for biological samples. In its greatest dimensions, the trap is 1.76 m in length, 1.52 m in width and 0.61 m in depth. A 0.4 m by 0.29 m blow out panel is placed on one side and kept closed using 7-day magnesium releases. The magnesium releases are examined after each soak and replaced as needed. The trap is deployed at a randomly selected subset of video stations. Both the camera pod and fish trap are baited with squid.

## VIDEO TAPE VIEWING PROCEDURES

One video tape from each station is selected out of the four for viewing. If all four video cameras face reef fish habitat and are in focus, the viewed tape is selected randomly. Tape viewers examine 20 minutes of the selected video tape, identify, and enumerate all species for the duration of the tape. Identifications are made to the lowest taxonomic level and the time when each fish enters and leaved the field of view is recorded. This is referred as a time in - time out procedure (TITO).

Tapes are viewed from the time when the view clears from any silt plume raised by the gear when it landed. Less than 20 minutes may be viewed if the duration when water is not clear enough to count fish is less than 20 minutes, or if the camera array is dragged. If a tape contains a large amount of fish, it is sub-sampled. There are four cases for sub-sampling: 1) when there is generally a large number of fish of a given species present throughout the tape so that following individual fish is difficult; 2) large number of fish occur in pulses periodically during the tape; 3) a single school of fish; and, 4) multiple schools of fish. Three estimators of relative abundance are available from the video data: 1) presence and absence; 2) maximum count (each fish of each taxon is counted each time it appears on the screen); and, 3 ) a minimum count (i.e., mincount: the greatest number of a taxon that appears on screen at one time). Presence and absence (frequency of occurrence) and mincount estimators are advantageous because they avoid the potential of multiple counting of fish, and are reported here.


Figure 2. SEAMAP offshore reef fish survey sample blocks in the eastern Gulf of Mexico.


Figure 3. SEAMAP offshore reef fish survey sample blocks in western Gulf of Mexico.

## STATISTICS

## Design-based Estimator

The design-based estimators of abundance are those for stratified, two-stage sampling (Cochran, 1977). The number of strata and number of blocks sampled in the eastern and western Gulf of Mexico during SEAMAP reef fish survey are shown in Table 1.

## 1. Block means

$\bar{x}_{h i}=\frac{\sum_{j=1}^{m_{h i}} x_{h i j}}{m_{h i}}$, where $x_{h i j}$ is the number of fish observed at the $j$-th site in the $i$-th block within the $h$-th stratum, and $m_{h i}$ in the number of sites sampled in the $i$-th block and $h$-th stratum.
2. Stratum means
$\bar{x}_{h}=\frac{\sum_{j=1}^{n_{h}} \bar{x}_{h i}}{n_{h}}$, where $\bar{x}_{h i}$ is the $i$-th block mean in the $h$-th stratum and $n_{h}$ is the number of blocks sampled in the $h$-th stratum.

## 3. Stratified mean

$\bar{x}_{s t}=\sum_{h} w_{h} \bar{x}_{h}$, where $w_{h}$ is the stratum weight estimated as the area of the stratum divided by
the total survey area $\left(A_{h} / A\right)$.

## 4. Variance of the stratified mean $\left(V\left(\bar{x}_{s t}\right)\right.$ ), ignoring finite population correction

$V_{\bar{x}_{s t}}=\sum_{h} w_{h}^{2}\left[\frac{s_{1 h}^{2}}{n_{h}}+\frac{s_{2 h}^{2}}{n_{h} m_{h}}\right]$, where $w_{h}$ is the stratum weight, $s^{2}{ }_{1 h}$ and $s^{2}{ }_{2 h}$ are the variances among
the first-stage and second-stage units, $n_{h}$ and $m_{h}$ are the number of first stage and second-stage units sampled.
5. Variance among first-stage units, $s^{2}{ }_{1 h}$
$s_{1 h}^{2}=\frac{\sum_{h}\left(\bar{x}_{h i}-\bar{x}_{h}\right)^{2}}{n_{h}-1}$.

## 6. Variance among second-stage units, $s^{2}{ }_{2 h}$

$s_{2 h}^{2}=\frac{\sum_{i} \sum_{j}\left(x_{h i j}-\bar{x}_{h i}\right)^{2}}{n_{h}\left(m_{h i}-1\right)}$.
The estimates for the frequency of occurrence of each species were calculated using the same equations where $x_{h i j}$ was either 0 or 1 . The final estimate is a stratified mean proportion.

## Model-based estimator

In addition to the calculations of stratified means, a delta-lognormal modeling approach (Lo et al., 1992) was used to develop abundance indices. In order to develop standardized indices of annual average mincount for each species in the U.S. Gulf of Mexico, a delta-lognormal model, as described by Lo et al. (1992), was employed. This index is a mathematical combination of yearly mincount estimates from two distinct generalized linear models: a binomial (logistic) model which describes proportion of positive mincounts (i.e., presence/absence) and lognormal model which describes variability in only the nonzero mincount data. The GLMMIX and MIXED procedures in SAS was employed to provide yearly index values for both the binomial and lognormal sub-models, respectively. The parameters included in each sub-model were year, stratum, and block nested within stratum. Also, the estimates from each model were weighted using the stratum area, and separate covariance structures were developed for each survey year. For the binomial models, a logistic-type mixed model was employed. The fit of each model was evaluated using the fit statistics provided by the GLMMIX macro. Initially, several model types were used to describe the nonzero mincount data. These included lognormal, Poisson and negative binomial. Based on analyses of residual scatter and QQ plots, the lognormal model was more fitting than the others in describing the variability in the nonzero data in most of the models. Those models where a lognormal approach did not fit as well were those where there were very few data points. In those cases, the other two model types did not perform any better. As with the design-based analyses, model-based estimators were developed for each species Gulfwide, East Gulf only and West Gulf only.

## Fish Sizes

The size of the fish observed during the survey come from two sources, fish captured in traps and fish measured on video tape with lasers. Lasers were first introduced in 1995.
However, since both the capture of fish in traps, and the instances where fish are hit by lasers is infrequent, size distributions were not estimated. We report only the average size and size range of fish.

## RESULTS

Abundance data from five strata were included for analysis during all years except 2001 (design-based estimates: Tables 2-7; model-based estimates: Tables 8-16). Stratum 1 was sampled only in 1994, 1996 and 1997. This stratum was $62.847 \mathrm{~km}^{2}$ in area. Stratum 7 was sampled only in 1996, 1997, and 2002, and was $13.030 \mathrm{~km}^{2}$ in area. Since these strata were not sampled during all years of the survey, they were excluded from design-based and model-based estimates of annual mean mincount. However, when included for those years, the stratified means and variances changed very little since their stratum weights were small. Figure 4 illustrates similarities between designed-based and model-based indices for each species.

The 2001 survey was abbreviated. Only one stratum was sampled in the eastern Gulf of Mexico. We recommend that the 2001 estimates of abundance for the entire Gulf of Mexico, and for the eastern Gulf of Mexico not be used for estimating trends in fish abundance. However, the two strata in the western Gulf of Mexico were sampled, and provide useful estimates of abundance.

Tables 17 and 18 provide size information on each species captured in traps and on video, respectively.

## LITERATURE CITED

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Lo, N. C. H., L.D. Jacobson, and J.L. Squire. 1992. Indices of relative abundance from fish spotter data based on delta-lognormal models. Can. J. Fish. Aquat. Sci. 49: 2515-1526.

Table 1. The number of strata and number of blocks sampled in the eastern and western Gulf of Mexico during SEAMAP reef fish survey.

|  | Eastern Gulf of Mexico |  | Western Gulf of Mexico |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Number of Strata | Number of <br> Blocks | Number of <br> Strata | Number of <br> Blocks |
| 1992 | 3 | 13 | 2 | 11 |
| 1993 | 3 | 18 | 2 | 9 |
| 1994 | 3 | 14 | 2 | 9 |
| 1995 | 3 | 12 | 2 | 10 |
| 1996 | 3 | 21 | 2 | 11 |
| 1997 | 3 | 20 | 2 | 17 |
| 2001 | 1 | 5 | 2 | 9 |
| 2002 | 3 | 19 | 2 | 14 |

Table 2. Stratified mean abundance of vermilion snapper observed during SEAMAP reef fish video survey.

|  | Gulf of Mexico Stratified Means |  |  | Eastern Gulf Region Stratifed Means |  |  | Western Gulf Region Stratified Means |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n | mean | se | n | mean | se | n | mean | se |
| 1992 | 23 | 5.02859 | 6.69132 | 13 | 2.14732 | 2.74426 | 11 | 11.83490 | 16.01540 |
| 1993 | 27 | 3.53715 | 5.43834 | 18 | 2.09246 | 3.57096 | 9 | 6.94989 | 9.84960 |
| 1994 | 23 | 4.19894 | 4.58391 | 14 | 4.22666 | 3.64164 | 9 | 4.13345 | 6.80980 |
| 1995 | 22 | 3.40568 | 3.96039 | 12 | 1.53844 | 1.62635 | 10 | 7.81663 | 9.47400 |
| 1996 | 32 | 1.36499 | 1.89149 | 21 | 0.7576 | 1.21043 | 11 | 2.79982 | 3.50032 |
| 1997 | 37 | 3.98796 | 3.98332 | 20 | 2.23192 | 2.18983 | 17 | 8.13620 | 8.22010 |
| 2001 | 14 | 1.01887 | 1.49052 | 5 | 0 | 0 | 9 | 3.42572 | 5.01155 |
| 2002 | 33 | 2.78000 | 3.4748 | 19 | 3.12016 | 4.03667 | 14 | 1.97630 | 2.14770 |
| Table 3 Stratified mean proportion of sites where vermilion snapper were observed during SEAMAP reef fish video survey. |  |  |  |  |  |  |  |  |  |
|  | Gulf of Mexico Stratified Means |  |  | Eastern Gulf Region Stratifed Means |  |  | Western Gulf Region Stratified Means |  |  |
| Year | n | mean | se | n | mean | se | n | mean | se |
| 1992 | 23 | 0.21714 | 0.17407 | 13 | 0.18460 | 0.15881 | 11 | 0.29401 | 0.21014 |
| 1993 | 27 | 0.16605 | 0.18906 | 18 | 0.15061 | 0.17225 | 9 | 0.20253 | 0.22876 |
| 1994 | 23 | 0.20655 | 0.17306 | 14 | 0.20378 | 0.16186 | 9 | 0.21311 | 0.19953 |
| 1995 | 22 | 0.22400 | 0.15726 | 12 | 0.16856 | 0.13951 | 10 | 0.35194 | 0.19917 |
| 1996 | 32 | 0.13976 | 0.12090 | 21 | 0.13216 | 0.11884 | 11 | 0.15773 | 0.12576 |
| 1997 | 37 | 0.20361 | 0.14324 | 20 | 0.14350 | 0.11970 | 17 | 0.34560 | 0.19887 |
| 2001 | 14 | 0.03806 | 0.04651 | 5 | 0.00000 | 0.00000 | 9 | 0.12796 | 0.15637 |
| 2002 | 33 | 0.18661 | 0.16645 | 19 | 0.15109 | 0.14425 | 14 | 0.27051 | 0.21888 |

Table 4 Stratified mean abundance of greater amberjack observed during SEAMAP reef fish video survey.

|  | Gulf of Mexico Stratified Means |  |  | Eastern Gulf Region Stratifed Means |  |  | Western Gulf Region Stratified Means |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | n | mean | se | n | mean | se | n | mean | se |
| 1992 | 23 | 1.94588 | 3.27287 | 13 | 2.37573 | 4.32501 | 11 | 0.93047 | 0.78744 |
| 1993 | 27 | 0.15007 | 0.35306 | 18 | 0.18103 | 0.46528 | 9 | 0.07694 | 0.08798 |
| 1994 | 23 | 0.67244 | 0.83264 | 14 | 0.54303 | 0.86382 | 9 | 0.97815 | 0.75898 |
| 1995 | 22 | 0.37868 | 0.32167 | 12 | 0.2218 | 0.25732 | 10 | 0.74927 | 0.47369 |
| 1996 | 32 | 0.37322 | 0.47432 | 21 | 0.34044 | 0 | 11 | 0.45064 | 0.63833 |
| 1997 | 37 | 0.43260 | 0.67295 | 20 | 0.42112 | 0.77474 | 17 | 0.45972 | 0.43249 |
| 2001 | 14 | 0.59547 | 0.63732 | 5 | 0.65982 | 0.73351 | 9 | 0.44345 | 0.41007 |
| 2002 | 33 | 1.22084 | 1.06756 | 19 | 1.43846 | 1.30273 | 14 | 0.70670 | 0.51200 |

Table 5 Stratified mean proportion of sites where greater amberjack observed during SEAMAP reef fish video survey

|  | Gulf of Mexico Stratified Means |  | Eastern Gulf Region Stratifed Means |  | Western Gulf Region Stratified Means |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | $\mathbf{n}$ | mean | se | $\mathbf{n}$ | mean | se | mean |  |  |
| 1992 | 23 | 0.30128 | 0.21692 | 13 | 0.23295 | 0.21621 | 11 | 0.46270 |  |
| 1993 | 27 | 0.04695 | 0.07234 | 18 | 0.03425 | 0.06572 | 0.21861 |  |  |
| 1994 | 23 | 0.26442 | 0.20502 | 14 | 0.21533 | 0.19127 | 0.07694 |  |  |
| 1995 | 22 | 0.20754 | 0.14771 | 12 | 0.11744 | 0.11564 | 0.08798 |  |  |
| 1996 | 32 | 0.16126 | 0.14295 | 21 | 0.15309 | 0.14736 | 0.38040 |  |  |
| 1997 | 37 | 0.12597 | 0.12254 | 20 | 0.07573 | 0.09449 | 0.23750 |  |  |
| 2001 | 14 | 0.15096 | 0.11972 | 5 | 0.11183 | 0.08713 | 11 | 17 | 0.22349 |
| 2002 | 33 | 0.33869 | 0.19412 | 19 | 0.30335 | 0.16532 | 0.18058 |  |  |

Table 6. Stratified mean abundance of gray triggerfish observed during SEAMAP reef fish video survey.

## Gulf of Mexico Stratified Means <br> Eastern Gulf Region Stratifed Means

Western Gulf Region Stratified Means

| Year | n | mean | se | n | mean | se | n | mean | se |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 23 | 0.68549 | 0.53380 | 13 | 0.76107 | 0.47108 | 11 | 0.50695 | 0.68196 |
| 1993 | 27 | 0.37395 | 0.30779 | 18 | 0.49471 | 0.38822 | 9 | 0.08869 | 0.11778 |
| 1994 | 23 | 0.33632 | 0.26596 | 14 | 0.35244 | 0.22379 | 9 | 0.29825 | 0.36556 |
| 1995 | 22 | 0.31823 | 0.27911 | 12 | 0.42140 | 0.32738 | 10 | 0.07451 | 0.16509 |
| 1996 | 32 | 0.29654 | 0.23129 | 21 | 0.33761 | 0.20766 | 11 | 0.19954 | 0.28711 |
| 1997 | 37 | 0.62533 | 0.59674 | 20 | 0.55431 | 0.41697 | 17 | 0.79310 | 1.02143 |
| 2001 | 14 | 0.05343 | 0.06709 | 5 | 0.07605 | 0.09549 | 9 | 0.00000 | 0.00000 |
| 2002 | 33 | 0.29957 | 0.28094 | 19 | 0.36289 | 0.30140 | 14 | 0.14997 | 0.23260 |

Table 7. Stratified mean proportion of sites where gray triggerfish observed during SEAMAP reef fish video survey.

| Year | Gulf of Mexico Stratified Means |  |  | Eastern Gulf Region Stratifed Means |  |  | Western Gulf Region Stratified Means |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | n | mean | se | n | mean | se | n | mean | se |
| 1992 | 23 | 0.31484 | 0.17499 | 13 | 0.37885 | 0.18336 | 11 | 0.16363 | 0.15524 |
| 1993 | 27 | 0.20569 | 0.13648 | 18 | 0.25522 | 0.14440 | 9 | 0.08869 | 0.11778 |
| 1994 | 23 | 0.23263 | 0.17613 | 14 | 0.26254 | 0.16794 | 9 | 0.18976 | 0.21651 |
| 1995 | 22 | 0.16870 | 0.13775 | 12 | 0.21559 | 0.14534 | 10 | 0.05792 | 0.11982 |
| 1996 | 32 | 0.19306 | 0.12238 | 21 | 0.23834 | 0.12799 | 11 | 0.08610 | 0.10912 |
| 1997 | 37 | 0.26684 | 0.16111 | 20 | 0.30158 | 0.16252 | 17 | 0.18478 | 0.15777 |
| 2001 | 14 | 0.04243 | 0.04934 | 5 | 0.06039 | 0.07023 | 9 | 0.00000 | 0.00000 |
| 2002 | 33 | 0.15436 | 0.11654 | 19 | 0.18839 | 0.12272 | 14 | 0.07399 | 0.10193 |

Table 8. Gulfwide model-based index for vermilion snapper.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | $\mathbf{L C L}$ | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.20301 | 133 | 1.35170 | 1.10362 | 0.42437 | 0.48907 | 2.49040 |
| 1993 | 0.14970 | 167 | 1.23911 | 1.01169 | 0.45195 | 0.42715 | 2.39615 |
| 1994 | 0.20661 | 121 | 2.89631 | 2.36474 | 0.37234 | 1.15029 | 4.86137 |
| 1995 | 0.13242 | 219 | 0.81285 | 0.66366 | 0.37390 | 0.32192 | 1.36820 |
| 1996 | 0.13495 | 289 | 0.32937 | 0.26892 | 0.42117 | 0.11985 | 0.60342 |
| 1997 | 0.23643 | 258 | 1.63294 | 1.33324 | 0.30594 | 0.73298 | 2.42506 |
| 2001 | 0.09091 | 77 | 0.45596 | 0.37228 | 0.91059 | 0.07869 | 1.76131 |
| 2002 | 0.17625 | 261 | 1.08008 | 0.88185 | 0.36054 | 0.43828 | 1.77435 |

Table 9. Eastern Gulf Region model-based index for vermilion snapper.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | $\mathbf{L C L}$ | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.16418 | 67 | 0.83699 | 0.81208 | 0.48851 | 0.32194 | 2.04845 |
| 1993 | 0.13333 | 120 | 0.83654 | 0.81164 | 0.49436 | 0.31858 | 2.06783 |
| 1994 | 0.19737 | 76 | 2.50334 | 2.42882 | 0.40401 | 1.11597 | 5.28615 |
| 1995 | 0.08088 | 136 | 0.47823 | 0.46400 | 0.52873 | 0.17189 | 1.25249 |
| 1996 | 0.13869 | 137 | 0.29689 | 0.28805 | 0.54549 | 0.10379 | 0.79942 |
| 1997 | 0.11486 | 148 | 0.88687 | 0.86047 | 0.46290 | 0.35646 | 2.07711 |
| 2001 | 0.00000 | 32 | 0.00000 | 0.00000 | . | . | . |
| 2002 | 0.14706 | 170 | 1.37589 | 1.33494 | 0.40489 | 0.61240 | 2.90994 |

Table 10. Western Gulf Region model-based index for vermilion snapper.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.24242 | 66 | 3.48810 | 1.35878 | 0.47752 | 0.54889 | 3.36365 |
| 1993 | 0.19149 | 47 | 4.26413 | 1.66107 | 0.58712 | 0.55935 | 4.93285 |
| 1994 | 0.22222 | 45 | 1.84805 | 0.71990 | 0.56082 | 0.25296 | 2.04878 |
| 1995 | 0.21687 | 83 | 1.92420 | 0.74957 | 0.43166 | 0.32792 | 1.71336 |
| 1996 | 0.13158 | 152 | 1.04334 | 0.40643 | 0.45666 | 0.17020 | 0.97052 |
| 1997 | 0.40000 | 110 | 4.43396 | 1.72723 | 0.28213 | 0.99306 | 3.00418 |
| 2001 | 0.15556 | 45 | 2.33142 | 0.90820 | 0.76046 | 0.23519 | 3.50706 |
| 2002 | 0.23077 | 91 | 1.20351 | 0.46882 | 0.38520 | 0.22281 | 0.98648 |

Table 11. Gulfwide model-based index for greater amberjack.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | CV | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.33835 | 133 | 0.45414 | 1.25672 | 0.23452 | 0.79115 | 1.99627 |
| 1993 | 0.04790 | 167 | 0.06868 | 0.19006 | 0.50069 | 0.07380 | 0.48944 |
| 1994 | 0.27273 | 121 | 0.50239 | 1.39023 | 0.24971 | 0.85009 | 2.27358 |
| 1995 | 0.20091 | 219 | 0.29836 | 0.82563 | 0.21890 | 0.53565 | 1.27262 |
| 1996 | 0.14533 | 289 | 0.19732 | 0.54602 | 0.24230 | 0.33865 | 0.88040 |
| 1997 | 0.14341 | 258 | 0.19946 | 0.55196 | 0.24567 | 0.34013 | 0.89574 |
| 2001 | 0.24675 | 77 | 0.55437 | 1.53406 | 0.31823 | 0.82427 | 2.85506 |
| 2002 | 0.33716 | 261 | 0.61625 | 1.70531 | 0.18287 | 1.18649 | 2.45098 |

Table 12. Eastern Gulf Region model-based index for greater amberjack.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.25373 | 67 | 0.38473 | 1.20711 | 0.42095 | 0.53817 | 2.70754 |
| 1993 | 0.02500 | 120 | 0.03806 | 0.11941 | 0.89581 | 0.02572 | 0.55431 |
| 1994 | 0.19737 | 76 | 0.38884 | 1.22001 | 0.40108 | 0.56350 | 2.64139 |
| 1995 | 0.08824 | 136 | 0.09344 | 0.29317 | 0.39318 | 0.13734 | 0.62579 |
| 1996 | 0.16788 | 137 | 0.19401 | 0.60873 | 0.34989 | 0.30850 | 1.20117 |
| 1997 | 0.08784 | 148 | 0.11964 | 0.37539 | 0.46268 | 0.15557 | 0.90581 |
| 2001 | 0.25000 | 32 | 0.74329 | 2.33214 | 0.53581 | 0.85388 | 6.36964 |
| 2002 | 0.29412 | 170 | 0.58772 | 1.84403 | 0.28844 | 1.04767 | 3.24575 |

Table 13. Western Gulf Region model-based index for greater amberjack.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.42424 | 66 | 0.55934 | 1.22248 | 0.20760 | 0.81063 | 1.84356 |
| 1993 | 0.10638 | 47 | 0.11890 | 0.25985 | 0.49665 | 0.10160 | 0.66460 |
| 1994 | 0.40000 | 45 | 0.64651 | 1.41298 | 0.24713 | 0.86826 | 2.29944 |
| 1995 | 0.38554 | 83 | 0.68508 | 1.49728 | 0.20069 | 1.00625 | 2.22791 |
| 1996 | 0.12500 | 152 | 0.20486 | 0.44773 | 0.28787 | 0.25465 | 0.78721 |
| 1997 | 0.21818 | 110 | 0.36593 | 0.79975 | 0.23616 | 0.50189 | 1.27439 |
| 2001 | 0.24444 | 45 | 0.49897 | 1.09053 | 0.33466 | 0.56837 | 2.09239 |
| 2002 | 0.41758 | 91 | 0.58081 | 1.26940 | 0.17776 | 0.89207 | 1.80633 |

Table 14. Gulfwide model-based index for gray triggerfish.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | $\mathbf{L C L}$ | $\mathbf{U C L}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.31579 | 133 | 0.18319 | 1.75021 | 0.45295 | 0.73768 | 4.15255 |
| 1993 | 0.25150 | 167 | 0.11955 | 1.14216 | 0.49057 | 0.45121 | 2.89120 |
| 1994 | 0.31405 | 121 | 0.13040 | 1.24584 | 0.48471 | 0.49711 | 3.12228 |
| 1995 | 0.11872 | 219 | 0.06997 | 0.66843 | 0.58091 | 0.22735 | 1.96529 |
| 1996 | 0.17647 | 289 | 0.09409 | 0.89890 | 0.46078 | 0.37376 | 2.16190 |
| 1997 | 0.29070 | 258 | 0.14659 | 1.40046 | 0.45125 | 0.59202 | 3.31288 |
| 2001 | 0.05195 | 77 | 0.01105 | 0.10561 | 1.90184 | 0.00890 | 1.25315 |
| 2002 | 0.15326 | 261 | 0.08252 | 0.78837 | 0.53966 | 0.28681 | 2.16702 |

Table 15. Eastern Gulf Region model-based index for gray triggerfish.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | $\mathbf{L C L}$ | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.43284 | 67 | 0.15471 | 1.82133 | 0.58990 | 0.61059 | 5.43286 |
| 1993 | 0.31667 | 120 | 0.09637 | 1.13449 | 0.59715 | 0.37595 | 3.42355 |
| 1994 | 0.39474 | 76 | 0.10380 | 1.22205 | 0.58687 | 0.41168 | 3.62759 |
| 1995 | 0.16912 | 136 | 0.06139 | 0.72276 | 0.64637 | 0.22171 | 2.35610 |
| 1996 | 0.29927 | 137 | 0.08369 | 0.98528 | 0.57603 | 0.33776 | 2.87414 |
| 1997 | 0.33784 | 148 | 0.09996 | 1.17677 | 0.58935 | 0.39485 | 3.50712 |
| 2001 | 0.12500 | 32 | 0.01018 | 0.11989 | 1.56521 | 0.01295 | 1.11009 |
| 2002 | 0.18824 | 170 | 0.06943 | 0.81743 | 0.62705 | 0.25839 | 2.58597 |

Table 16. Western Gulf Region model-based index for gray triggerfish.

| Survey Year | Frequency | $\mathbf{N}$ | Lo Index | Standardized Index | $\mathbf{C V}$ | LCL | UCL |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 0.19697 | 66 | 0.36808 | 1.59463 | 0.40389 | 0.73285 | 3.46980 |
| 1993 | 0.08511 | 47 | 0.11838 | 0.51285 | 0.86507 | 0.11502 | 2.28671 |
| 1994 | 0.17778 | 45 | 0.26040 | 1.12814 | 0.52773 | 0.41863 | 3.04014 |
| 1995 | 0.03614 | 83 | 0.05886 | 0.25500 | 1.31463 | 0.03438 | 1.89115 |
| 1996 | 0.06579 | 152 | 0.13012 | 0.56373 | 0.57065 | 0.19494 | 1.63020 |
| 1997 | 0.22727 | 110 | 0.55790 | 2.41700 | 0.28635 | 1.37860 | 4.23757 |
| 2001 | 0.00000 | 45 | 0.00000 | 0.00000 | . | . | . |
| 2002 | 0.08791 | 91 | 0.12202 | 0.52863 | 0.68219 | 0.15354 | 1.82004 |

Table 17. Mean fork lengths (mm) of fish captured in fish traps during SEAMAP reef fish survey.

| Species | Year | $\mathbf{n}$ | Mean | Minimum | Maximum | SE |
| :---: | :---: | ---: | :---: | :---: | ---: | ---: |
| Vermilion Snapper | 1992 | 122 | 272.48 | 209 | 421 | 4.0185 |
|  | 1993 | 89 | 224.34 | 165 | 442 | 5.3630 |
|  | 1994 | 130 | 219.78 | 141 | 374 | 4.1069 |
|  | 1995 | 39 | 245.44 | 194 | 372 | 7.1362 |
|  | 1996 | 48 | 311.21 | 195 | 472 | 12.5915 |
|  | 1997 | 6 | 310.17 | 268 | 365 | 14.7680 |
|  | 2002 | 40 | 243.05 | 195 | 409 | 8.8145 |
| Gray Triggerfish | 1992 | 45 | 323.18 | 184 | 535 | 8.8625 |
|  | 1993 | 4 | 299.25 | 268 | 335 | 13.9545 |
|  | 1994 | 14 | 325.43 | 285 | 384 | 8.3808 |
|  | 1995 | 3 | 301.33 | 272 | 322 | 15.0702 |
|  | 1996 | 19 | 316.53 | 255 | 390 | 8.7524 |
|  | 1997 | 19 | 290.70 | 123 | 404 | 34.1406 |
|  | 2002 | 351.84 | 239 | 565 | 20.5172 |  |

Table 18. Mean fork lengths (mm) of fish measured on video tapes with lasers during SEAMAP reef fish survey.

| Species | YEAR | $\mathbf{n}$ | Mean | Minimum | Maximum | SE |
| :---: | :---: | ---: | :---: | :---: | :---: | ---: |
| Vermilion Snapper | 1995 | 47 | 287.62 | 218 | 483 | 51.9557 |
|  | 1996 | 191 | 312.26 | 135 | 475 | 5.0306 |
|  | 1997 | 291 | 260.93 | 152 | 568 | 3.3447 |
|  | 2001 | 8 | 277.50 | 210 | 329 | 16.9284 |
|  | 2002 | 927 | 266.31 | 144 | 586 | 2.0875 |
| Gray triggerfish | 1995 | 4 | 338.75 | 253 | 483 | 51.9557 |
|  | 1996 | 56 | 367.71 | 266 | 623 | 10.2886 |
|  | 1997 | 73 | 369.45 | 210 | 543 | 8.2549 |
|  | 2001 | 1 | 383.00 | 383 | 383 | - |
|  | 2002 | 65 | 361.63 | 272 | 530 | 8.2029 |
| Greater amberjack | 1995 | 72 | 555.50 | 267 | 1381 | 17.8616 |
|  | 1996 | 73 | 665.78 | 255 | 1187 | 22.2849 |
|  | 1997 | 28 | 586.58 | 295 | 1413 | 26.1735 |
|  | 2001 | 748.43 | 345 | 1229 | 53.0216 |  |
|  | 2002 |  | 571.67 | 218 | 1563 | 6.9895 |



Figure 4. Similarities between designed-based and model-based Gulfwide indices for vermilion snapper, greater amberjack, and gray triggerfish.

