# Chapter 2: Assessment of the Pacific Cod Stock in the Gulf of Alaska 

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## EXECUTIVE SUMMARY

## Summary of Major Changes

Relative to the November edition of last year's GOA SAFE report, the following substantive changes have been made in the Pacific cod stock assessment.

## Changes in the Input Data

1) Catch data for 1964-1977 were incorporated, catch data for 2004 were updated, and preliminary catch data for 2005 were incorporated.
2) Size composition data from the 1977 commercial fisheries were incorporated, size composition data from the 2004 commercial fisheries were updated, and preliminary size composition data from the 2005 commercial fisheries were incorporated.
3) Size composition data from the GOA bottom trawl survey were incorporated.
4) The biomass estimate from the GOA bottom trawl survey was incorporated (the 2005 estimate of $308,102 \mathrm{t}$ was up about $4 \%$ from the 2003 estimate).
5) Age composition data from the 2003 GOA bottom trawl survey were incorporated.
6) Length-at-age data from the 2003 GOA bottom trawl survey were incorporated.
7) A new maturity-at-length schedule was incorporated.

## Changes in the Assessment Model

Three alternative models are presented. Model 1 is identical to last year's model, which was developed using the Stock Synthesis 1 assessment software that has formed the basis of the GOA Pacific cod model since 1994. Models 2 and 3 were developed under the new Stock Synthesis 2 assessment software, which uses automatic differentiation (via the ADMB programming language) to minimize the objective function rather than the finite-difference algorithm used in Stock Synthesis 1. In addition, Stock Synthesis 1 and Stock Synthesis 2 differ with respect to several technical details which are described in the main text of this chapter. The primary difference between Model 2 and Model 3 is that Model 2 fixes the natural mortality rate $M$ and the EBS bottom trawl survey catchability coefficient $Q$ at values of 0.37 and 1.00 , respectively (identical to the values assumed in Model 1), whereas Model 3 allows the values of these two parameters to be estimated internally.

## Changes in Assessment Results

1) Based on Model 3, the estimated 2006 female spawning biomass for the GOA stock is $165,000 \mathrm{t}$, up about $80 \%$ from last year's estimate for 2005 and up about $89 \%$ from last year's $F_{A B C}$ projection for 2006. These changes are due largely to use of the new maturity schedule.
2) Based on Model 3, the estimated 2006 total age $3+$ biomass for the GOA stock is $453,000 \mathrm{t}$, down about $4 \%$ from last year's estimate for 2005 and up about $8 \%$ from last year's $F_{40 \%}$ projection for 2006.
3) Based on Model 3, the recommended 2006 ABC for the GOA stock is $58,700 \mathrm{t}$, up about $1 \%$ from last year's estimate for 2005 and up about 15\% from last year's $F_{A B C}$ projection for 2006.
4) Based on Model 3, the estimated 2006 OFL for the GOA stock is $130,000 \mathrm{t}$, up about $51 \%$ from last year's estimate for 2005.

## Responses to Comments of the Scientific and Statistical Committee (SSC)

## SSC Comments Specific to the Pacific Cod Assessments

From the December, 2004 minutes: "The SSC was intrigued by the stock-recruit fits for the periods 1977-1988 and 1989-2002 and we thank the authors for including this analysis. For the 2006 assessment, the SSC asks the authors to explore whether these findings can be used to elevate the GOA cod stock to tier 1 or 2. If it is deemed that MSY is too variable between periods to apply any MSY estimates to this stock, then next year's assessment should consider potential implications of this variability in stock productivity on estimation of the F35\% and F40\% reference points." The Ricker stock-recruitment curves shown in last year's assessment were intended to be illustrative only, because the statistical technique used to compute the parameters of those curves has significant drawbacks, as described in last year's assessment. Therefore, it does not seem advisable to use those parameters as the basis for elevating the GOA Pacific cod stock to Tier 1 or 2 . While statistically valid estimates of stock-recruitment parameters and the associated uncertainties may soon be available for GOA Pacific cod, it was not possible to develop them in time for this year's assessment. Most assessment effort this year went toward understanding and applying the new Stock Synthesis 2 modeling software. Unfortunately, Stock Synthesis 2 currently supports only the Beverton-Holt stock-recruitment function, although support for the Ricker function will undoubtedly be forthcoming. Regarding the issue of decadal-scale variability in stock-recruitment parameters, no determination has been made as to whether such variability would likely detract from the applicability of any future estimates of MSY or related quantities in the case of GOA Pacific cod. The subject of nonstationary stock-recruitment relationships is an active area of research at the Alaska Fisheries Science Center, and results of this research may be applicable to future assessments of the GOA Pacific cod stock. As a first, albeit small, step toward incorporating a stock-recruitment relationship into the provision of fishery management advice, the standard program used to make future projections in all Tier 1-3 BSAI and GOA groundfish assessments now includes an option to fit a Ricker stock-recruitment relationship by assuming that $\mathrm{F}_{35 \%}$ and $\mathrm{B}_{35 \%}$ correspond to $\mathrm{F}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$, respectively. The projections provided in the present assessment make use of this option.
From the December, 2004 minutes: "The authors are asked to examine interannual variability in cod weight-at-length estimates (index of condition) and potential relationships with cod density, stock-recruit, or environmental conditions. Condition indices have been useful metrics in analyses of the health of Atlantic cod stocks." Interannual variability of the weight-at-length relationship and condition factor is explored in the "Weight at Length" subsection of the "Data" section and in Figures 2.4 and 2.5.
From the December, 2004 minutes: "The SSC also requests that the authors provide justification for their assumption that there are no gender-based differences in length-at-age or weight-at-length for Pacific cod. If there is sexual dimorphism in growth, then size-based selection in the fisheries will generate time variations in sex ratios that can have important consequences to the stock's productivity." Sex-specific length at age and weight at length is explored in the "Length at Age" and "Weight at Length" subsections of the "Data" section and in Figures 2.2 and 2.3.
From the October, 2005 minutes: "Given the amount of time required to update the Bering Sea model, the feasibility of implementing a Gulf area model in SS2 this year is unclear. Nevertheless, the SSC
encourages development of the Gulf model in SS2 for comparability with the Bering Sea assessment." Two SS2-based models have been developed for the GOA Pacific cod stock. These are described and evaluated in the "Analytic Approach" and "Model Evaluation" sections.

## SSC Comments on Assessments in General

From the December, 2004 minutes: "In its review of the SAFE chapters, the SSC noted that there is variation in the information presented. Several years ago, the SSC developed a list of items that should be included in the document. The SSC requests that stock assessment authors exert more effort to address each item contained in the list." Every reasonable effort has been made to respond to all SSC requests and to ensure that the GOA Pacific cod assessment complies with the "Guide to the Preparation of Alaska Groundfish SAFE Report Chapters" produced by the Alaska Fisheries Science Center (last revised in June, 2003).

## INTRODUCTION

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $63^{\circ} \mathrm{N}$ latitude. Pacific cod is distributed widely over Gulf of Alaska (GOA), as well as the eastern Bering Sea (EBS) and the Aleutian Islands (AI) area. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and GOA. Although at least one previous genetic study (Grant et al. 1987) failed to show significant evidence of stock structure within these areas, current genetic research underway at the Alaska Fisheries Science Center may soon shed additional light on the issue of stock structure of Pacific cod within the BSAI (M. Canino, AFSC, pers. commun.). Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the GOA.

## Fishery

During the two decades prior to passage of the Magnuson Fishery Conservation and Management Act (MFCMA) in 1976, the fishery for Pacific cod in the GOA was small, averaging around $3,000 \mathrm{t}$ per year. Most of the catch during this period was taken by the foreign fleet, whose catches of Pacific cod were usually incidental to directed fisheries for other species. By 1976, catches had increased to $6,800 \mathrm{t}$. Catches of Pacific cod since 1978 are shown in Tables 2.1a and 2.1b. In Table 2.1a, catches for 19781990 are broken down by year, fleet sector, and gear type. In Table 2.1b, catches for 1991-2005 are broken down by year, jurisdiction, and gear type. The foreign fishery peaked in 1981 at a catch of nearly $35,000 \mathrm{t}$. A small joint venture fishery existed through 1988, averaging a catch of about $1,400 \mathrm{t}$ per year. The domestic fishery increased steadily through 1986, then increased more than three-fold in 1987 to a catch of nearly $31,000 \mathrm{t}$ as the foreign fishery was eliminated. Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components. Trawl gear has traditionally accounted for the bulk of the catch (approximately $55 \%$ on average during the period 19972004), although pot gear took the largest share of the catch in 2003 and 2004. Figure 2.1 shows areas in which sampled hauls for each of the three main gear types (trawl, longline, and pot) were concentrated during 2004. To create this figure, the EEZ off Alaska was divided into $20 \mathrm{~km} \times 20 \mathrm{~km}$ squares. A square is shaded if more than two hauls containing Pacific cod were sampled in it during 2004.

The history of acceptable biological catch (ABC) and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate commercial catches in Table 2.2. For the first year of management under the MFCMA (1977), the catch limit for GOA Pacific cod was established at slightly less than the 1976 total reported landings. During the period 1978-1981, catch limits varied between 34,800 and $70,000 \mathrm{t}$, settling at $60,000 \mathrm{t}$ in 1982. Prior to 1981 these limits were assigned for "fishing years" rather than calendar years. In 1981 the catch limit was raised temporarily to $70,000 \mathrm{t}$ and the
fishing year was extended until December 31 to allow for a smooth transition to management based on calendar years, after which the catch limit returned to $60,000 \mathrm{t}$ until 1986, when ABC began to be set on an annual basis. From 1986 (the first year in which an ABC was set) through 2005, TAC averaged about $84 \%$ of ABC and catch averaged about $96 \%$ of TAC. In 8 of these 20 years (40\%), TAC equaled ABC exactly. In 9 of these 20 years ( $45 \%$ ), catch exceeded TAC. However, it should be noted that all but two of these apparent overages occurred in the most recent nine years, when a substantial fishery for Pacific cod was conducted inside State of Alaska waters, mostly in the Western and Central Regulatory Areas. To accommodate the State-managed fishery, TAC was set well below ABC in each of those years ( $15 \%$ in 1997 and 1998; 20\% in 1999; 23\% in 2000-2003; and $24 \%$ in 2004-2005). Thus, the apparent overages in 1997-2000 and 2002-2004 are basically an artifact of the bi-jurisdictional nature of the fishery. Catch has exceeded ABC only twice (in 1992 and 1996). Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3) changes in the stock assessment model. For example, from 1986 through 2004, three different assessment models were used (Table 2.2), though the model was largely unchanged from 1997 through last year.
Historically, the majority of the GOA catch has come from the Central regulatory area. To some extent the distribution of effort within the GOA is driven by regulation, as catch limits within this region have been apportioned by area throughout the history of management under the MFCMA. Changes in areaspecific allocation between years have usually been traceable to changes in biomass distributions estimated by Alaska Fisheries Science Center trawl surveys or management responses to local concerns. Currently, the ABC allocation follows the average biomass distribution estimated by the three most recent trawl surveys, and the TAC allocation is within one percent of this distribution on an area-by-area basis. The complete history of allocation (in percentage terms) by regulatory area within the GOA is shown in Table 2.3.
In addition to area allocations, GOA Pacific cod is also allocated on the basis of processor component (inshore/offshore) and season. The inshore component is allocated $90 \%$ of the TAC and the remainder is allocated to the offshore component. Within the Central and Western Regulatory Areas, $60 \%$ of each component's portion of the TAC is allocated to the A season (January 1 through June 10) and the remainder is allocated to the B season (June 11 through December 31, although the B season directed fishery does not open until September 1). The longline and trawl fisheries are also associated with a Pacific halibut mortality limit which sometimes constrains the magnitude and timing of harvests taken by these two gear types.

The catches shown in Tables 2.1a-b and 2.2 include estimated discards for all years since 1980. Discard rates of Pacific cod in the various GOA target fisheries are shown for each year 1991-2002 in Table 2.4a and for the years 2003-2004 in Table 2.4b.

## DATA

This section describes data used in the current assessment. It does not attempt to summarize all available data pertaining to Pacific cod in the GOA.

## Commercial Catch Data

## Catch Biomass

Catches (including estimated discards) taken in the GOA since 1964 are shown in Table 2.5, broken down by the three main gear types and the following within-year time intervals, or "periods": January-May, June-August, and September-December. This particular division, which was suggested by participants in the BSAI fishery, is intended to reflect actual intra-annual differences in fleet operation (e.g., fishing operations during the spawning period may be different than at other times of year). In years for which
estimates of the distribution by gear or period were not available, proxies based on other years' distributions were used.

## Catch Size Composition

Fishery size compositions are presently available, by gear, for the years 1977 through the first part of 2004. For ease of representation and analysis, length frequency data for Pacific cod can usefully be grouped according to the following set of 25 intervals or "bins," with the upper and lower boundaries shown in cm:

Lower $\quad 912151821242730333639424550556065707580859095100105$ Bound:
Upper Bound: 1114172023262932353841444954596469747984899499104115

Total length sample sizes for each year, gear, and period are shown in Table 2.6. The collections of relative length frequencies are shown by year, period, and size bin for the pre-1987, 1987-1999, and post1999 trawl fisheries in Tables 2.7a, 2.7b, and 2.7c, respectively; the pre-1987, 1987-1999, and post-1999 longline fisheries in Tables 2.8a, 2.8b, and 2.8c, respectively; and the 1987-1999 and post-1999 pot fisheries in Tables 2.9a and 2.9b. Fishery length frequencies since 1997 include samples from the Statemanaged fishery.

## Survey Data

## Survey Size Composition

The relative size compositions from trawl surveys of the GOA conducted by the Alaska Fisheries Science Center since 1984 are shown in Table 2.10, using the same length bins defined above for the commercial catch size compositions. Total sample sizes are shown below:

| Year: | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 | 2003 | 2005 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Sample size: | 17413 | 19589 | 11440 | 17152 | 12190 | 8645 | 6772 | 9125 | 6844 |

## Survey Age Composition

Following a decade-long hiatus in production ageing of Pacific cod, the Age and Growth Unit of the Alaska Fisheries Science Center began ageing samples of Pacific cod from the EBS shelf bottom trawl surveys a few years ago (Roberson 2001, Roberson et al. 2005). This year, age data have become available for the 2003 survey in the GOA as well. Age composition estimates for that survey are shown in Table 2.11 (sample size = 711). Age data are not yet available for any other year.

## Abundance Estimates

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl surveys are shown in Table 2.12, together with the standard errors and upper and lower $95 \%$ confidence intervals (CI) for the biomass estimates.

The highest biomass ever observed by the survey was the 1984 estimate of $550,971 \mathrm{t}$, and the low point is the 2001 estimate of 279,332 $t$ (the 2001 estimate was obtained by summing the 2001 estimate for the Western and Central areas with the 1999 estimate for the Eastern area, because the 2001 survey did not cover the Eastern area). In terms of population numbers, the record high was observed in 1984, when the population was estimated to include over 320 million fish.

## Length at Age

Production ageing of Pacific cod at the Alaska Fisheries Science Center was curtailed in the early 1990s and did not resume for approximately ten years. During the intervening period, age data were used only sparingly in the GOA Pacific cod assessment. However, as noted above, length-at-age data from the 2003 survey are now available. These data provide the following relationship between age and length (cm) and the amount of spread around that relationship (data were collected during summer; ages assume a January 1 birthdate):

| Age: | 1.5 | 2.5 | 3.5 | 4.5 | 5.5 | 6.5 | 7.5 | 8.5 | 9.5 | 10.5 | 11.5 | 12.5 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Mean Length: | 21.0 | 31.7 | 42.3 | 48.8 | 58.3 | 66.4 | 74.0 | 78.1 | 88.0 | 85.5 | 77.0 | 80.5 |
| St. Deviation: | 3.5 | 3.6 | 5.6 | 5.8 | 5.2 | 4.5 | 5.7 | 7.6 | 5.9 | 7.8 | 4.2 | 14.8 |
| N: | 54 | 45 | 119 | 141 | 140 | 91 | 62 | 41 | 12 | 2 | 2 | 2 |

The SSC has asked that the potential significance of sex-specific length at age be addressed (SSC minutes, December 2004). Figure 2.2 shows sex-specific schedules of mean length at age based on the 2003 surveys data, together with $95 \%$ confidence intervals (only ages 1.5-8.5 are shown because of small sex-specific sample sizes at ages $9.5-12.5$ ). The sex-specific means appear to be very close throughout most of the age range. Although the female curve is slightly lower than the male curve at ages less than 6.5 and higher than the male curve at older ages, the confidence intervals for the two sexes overlap at all ages except 7.5 and 8.5. More data may be necessary to determine whether or not there is a consistent and biologically important difference between the length-at-age schedules of the two sexes in the GOA. However, in the BSAI Pacific cod assessment (Thompson and Dorn 2005), where eight years of survey ages are available, no major difference between the length-at-age schedules of the two sexes is apparent.

## Weight at Length

Weight measurements taken during summer bottom trawl surveys since 1987 yield the following data regarding average weights (in kg ) at length, grouped according to size composition bin (as defined under "Catch Size Composition" above):
 Ave. weight: $\quad n / a 0.00 .00 .10 .10 .20 .20 .30 .40 .50 .70 .81 .11 .52 .02 .53 .24 .05 .26 .38 .09 .511 .513 .213 .9$

The SSC has asked that the potential significance of sex-specific weight at length be addressed (SSC minutes, December 2004). Figure 2.3 shows sex-specific schedules of mean weight at length based on the 2003 survey data, together with $95 \%$ confidence intervals. The sex-specific means appear to be very close throughout most of the length range. At longer lengths, the means may not overlap, but there is no obvious trend (i.e., males have higher average weight than females at some lengths but not at others) and the means for each sex typically fall within the confidence interval for the other sex.

The SSC has also asked that the potential significance of interannual variability in weight-at-length relationships and condition factor be addressed (SSC minutes, December 2004). For this purpose, a set of six example years was chosen: 1980, 1985, 1990, 1995, 2000, and 2005. For each of these years, all the available weight-at-length data from the commercial fisheries were compiled (commercial fishery data were chosen rather than survey data because Pacific cod weights were not collected in all surveys and because the commercial time series extends farther back than the survey time series). The average month of collection (where January=1) ranged from 2.3 to 5.7, with the data from the three most recent years (1995, 2000, and 2005) tending to be collected somewhat earlier in the respective year than the data from the other years. By gear type, the data for 1980 were collected predominantly from the longline fishery, the data for 1985 were collected predominantly from the trawl fishery, the data from 1990 and 1995 were collected predominantly from motherships, the data for 2000 were collected predominantly from the pot fishery, and the data for 2005 were collected predominantly from the jig fishery.

The mean weights at each length are shown for each year in Figure 2.4 (to reduce the possibility of outliers resulting from small sample sizes, only those points representing the average of at least 5 data points are shown). For the most part, the mean weights at length appear very close for all years.
"Condition factor," conventionally defined as the ratio of weight to the cube of length, is commonly used to compare the health of individual fish of the same species (Fulton 1911, Ricker 1975). The average condition factor (across lengths) for each of the six example years is plotted together with $95 \%$ confidence intervals in Figure 2.5. Because condition factor is a relative measure, the values in Figure 2.5 have been normalized by expressing each point as the ratio of the year-specific estimate to the estimate for the entire time series. Statistically speaking, the point estimates in Figure 2.5 are significantly different under any reasonable criterion. However, it should be emphasized that such a result would probably be expected, given that only one parameter is being estimated for each year and a total of approximately 20,000 points is used in the analysis. More important questions are, "How different are they?" and, "Would such differences be important to incorporate into the stock assessment?" It may be useful to pursue these questions further in future assessments, but for the time being it may be sufficient to note that there does not appear to be any obvious time trend to the points in Figure 2.5 (although there probably would be if the 2005 point were omitted), and all but one of the points is within $5 \%$ of the longterm average.

## Maturity at Length

For many years, the GOA Pacific cod assessments have relied on maturity-at-length data collected from the Bering Sea. As different Bering Sea data sets have been compiled over the years, the schedule of maturity at length has changed from time to time. Prior to 1995, the Pacific cod assessments used a maturity schedule based on a gonadosomatic index calculated from a sample of 1900 fish collected during the 1981 and 1982 survey seasons and described by Teshima (1985).

More recently (since 1995 in the BSAI and 1996 in the GOA), the Pacific cod assessments have used a maturity schedule based on a sampling program conducted in 1993-1994, using observer data collected from the EBS commercial fisheries. The data consisted of observers' visual determinations regarding the spawning condition of 2312 females. Of these 2312 females, 231 were smaller than 42 cm (the lower boundary of length bin 12). None of these sub- 42 cm fish were mature. The observed proportions of mature fish in the remaining length bins, together with the numbers of fish sampled in those length bins, are shown below (bins are defined under "Catch Size Composition" above):

| Bin number: | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Prop. mature: | 0.03 | 0.05 | 0.14 | 0.19 | 0.28 | 0.53 | 0.69 | 0.82 | 0.89 | 0.94 | 0.94 | 0.91 | 0.89 | 1.00 |
| Sample size: | 39 | 122 | 226 | 313 | 295 | 300 | 320 | 177 | 103 | 70 | 50 | 35 | 19 | 12 |

Recently, Stark (2005) completed an in-depth histological study of Pacific cod maturity in both the BSAI and GOA. In the GOA, 346 female fish were collected in the central GOA during October 1998, January 1999, April 1999, June 1999, and January 2004. Methods were the same as those used by Stark (2004). Oocytes within each ovary were classified into seven histological stages based on the criteria of Hunter et al. (1992) and Stark (2004). Fish with ovaries containing either hydrated oocytes or post-ovulatory follicles were classified as spawners. Specimens collected from the GOA ranged in size from 13-98 cm. The smallest spawning female collected from the GOA was 42 cm . Ovary weights were found to represent up to $30 \%$ of total body weight.

## ANALYTIC APPROACH

## Model Structure

Beginning with the 1994 SAFE report (Thompson and Zenger 1994), a model using the SS1 assessment program (Methot 1986, 1990, 1998, 2000) and based largely on length-structured data has formed the primary analytical tool used to assess the GOA Pacific cod stock. SS1 is a program that uses the parameters of a set of equations governing the assumed dynamics of the stock (the "model parameters") as surrogates for the parameters of statistical distributions from which the data are assumed to be drawn (the "distribution parameters"), and varies the model parameters systematically in the direction of increasing likelihood until a maximum is reached. The overall likelihood is the product of the likelihoods for each of the model components. Each likelihood component is associated with a set of data assumed to be drawn from statistical distributions of the same general form (e.g., multinomial, lognormal, etc.). Typically, likelihood components are associated with data sets such as catch size (or age) composition, survey size (or age) composition, and survey biomass.

SS1 permits each data time series to be divided into multiple segments, resulting in a separate set of parameter estimates for each segment. In the base model for the GOA Pacific cod assessment, for example, possible differences in selectivity between the mostly foreign (also joint venture) and mostly domestic fisheries have traditionally been accommodated by splitting the fishery size composition time series into pre-1987 and post-1986 segments.

The base model for GOA Pacific cod remained completely unchanged from 1997 to 2001. A minor modification of the base model was suggested by the SSC in 2001, namely, that consideration be given to dividing the domestic era into pre-2000 and post-1999 segments. This modification was tested in the 2002 assessment (Thompson et al. 2002), where it was found to result in a statistically significant improvement in the model's ability to fit the data.

However, after so many years of application, the SS1 architecture has by this time become somewhat dated. Three features can be identified as no longer state-of-the-art: First, SS1 uses a finite difference algorithm to minimize the objective function, whereas most state-of-the-art assessments use automatic differentiation (e.g., Greiwank and Corliss 1991), for example, as found in the ADMB modeling package (Fournier 2005). Benchmark tests have tended to indicate that automatic differentiation is a superior algorithm. Second, SS1 attempts to estimate all parameters simultaneously, whereas models programmed in ADMB can include "phased" estimation, where attention is focused initially on only a subset of parameters, and additional parameters are added to the "active" list with each subsequent phase, until finally all parameters are active in the final phase. By attempting to estimate all parameters simultaneously, SS1 is more likely to get trapped in a local minimum. Third, SS1 does not include utilities for estimating confidence intervals or posterior distributions of derived quantities (e.g, spawning biomass), whereas models programmed in ADMB can easily be tailored to estimate such confidence intervals or distributions so long as the estimated Hessian matrix is positive definite.

Therefore, SS1 is being replaced by a new program, SS2, which, for the most part, is simply SS1 rewritten in ADMB. A full description of SS2, including the equations used to model population dynamics and the various observation processes, is given by Methot (2005a). This year's Pacific cod assessment includes three alternative models. Model 1 was configured under SS1, while Models 2 and 3 were configured under SS2 (see "Model Evaluation" below). The structure of Model 1 configured under SS1 is identical to that described in last year's assessment (Thompson et al. 2004).

Although the main difference between SS1 and SS2 is the use of ADMB by SS2, there are a number of other technical differences. The most important of these, and how they were addressed in the present assessment, are described in the following paragraphs.

## Minimum and Maximum Age

SS1 allowed the user to specify the minimum age in the model, whereas SS2 automatically sets the minimum age equal to zero. This does not mean, however, that the data have to include age 0 fish; it simply means that SS2 always begins calculating the age structure of the population at age 0 . Moreover, the SS2 user can still specify a "summary age range" for use in reporting output, where the minimum age is completely flexible. Another difference between SS1 and SS2 is that users of SS1 were encouraged to set a fairly low age for the boundary of the "plus" group, with the age structure of the plus group governed by a user-specified "old age discount" parameter, whereas users of SS2 are encouraged to set a fairly high age for the boundary of the plus group, so that the age structure of the plus group essentially does not matter (again, setting a high maximum age does not mean that the data must include all ages up to that maximum, it simply means that SS2 will calculate the age structure of the population up to that maximum). In SS1, maximum age for the Pacific cod model has always been set equal to 12, and this assumption is retained for the SS1 model included in the present assessment. For the models developed under SS2 in this year's assessment, maximum age is set equal to 20.

## Initial Numbers at Age

SS1 provided users with the choice of setting the numbers-at-age vector in the initial year equal to the equilibrium numbers-at-age vector associated with user-specified levels of catch and recruitment, or estimating each element of the numbers-at-age vector in the initial year as a free parameter. Previous GOA Pacific cod assessment models have always used the second option, where the initial year was set equal to 1978. However, SS2 requires use of the first option. Use of an equilibrium initial numbers-atage vector necessitates a number of modifications to the GOA Pacific cod assessment model. This is because previous assessments of this stock, as well as conventional wisdom, have consistently indicated that one or more exceptionally large year classes spawned in or around 1977 were present in the population in 1978, but most other age groups were at very low levels of abundance in that year, meaning that the assumption of initial equilibrium would likely be very misleading (i.e, it would either cause the large year classes to be drastically under-estimated, or the other year classes to be drastically overestimated).

It is clear that the assumption of initial equilibrium requires an earlier initial year for the assessment model. Annual catch data are available as far back as 1964. Setting the initial year equal to 1964 would give the model plenty of time to generate a reasonable age structure by the time the large year classes of the mid-to-late 1970s were spawned. However, setting the initial year any earlier than 1977 requires estimating one or more year classes prior to the well-documented 1977 environmental regime shift (e.g., Hare and Mantua 2000), which should have a lower median value than year classes spawned after the 1977 regime shift. Establishing different pre-1977 and post-1976 medians is easily accomplished in SS2 by creating a regime shift "dummy variable" for each year in the time series and estimating a link between median recruitment and the dummy variable. However, this creates another problem, because the parameter governing the amount of stochastic variability in recruitment ( $\sigma_{R}$ ) cannot be linked to the dummy variable. This means that the mean recruitment deviation for each portion of the time series (pre1977 and post-1976) will not necessarily equal zero, even though SS2 forces the mean recruitment deviation for the overall time series to equal zero. This, in turn, means that the estimates of the pre- and post-regime shift medians will be confounded with the estimate of $\sigma_{R}$.
To resolve the problem of confounding between the estimates of the pre-1977 and post-1976 recruitment medians with the estimate of $\sigma_{R}$, the following iterative algorithm was used to implement an environmental regime shift in SS2.

1. Candidate values for the pre-1977 log-scale mean and $\sigma_{R}$ were chosen.
2. SS2 was allowed to estimate the post-1976 log-scale mean and the recruitment deviations for the entire time series (deviations are expressed as the difference between the logarithm of
annual recruitment at age 0 and the log-scale mean for the respective environmental regime), conditional on the candidate values for the pre-1977 log-scale mean and $\sigma_{R}$.
3. The mean of the estimated pre-1977 recruitment deviations and the standard deviation of the entire time series of recruitment deviations were computed.
4. If the absolute value of the mean computed in Step 3 was less than 0.005 and the standard deviation computed in Step 3 was equal to $\sigma_{R}$ with three significant digits, the candidate values were determined to be the final estimates. If either of these conditions did not hold, the candidate value for the pre-1977 log-scale mean was set equal to the old value plus the mean computed in Step 3, the candidate value for $\sigma_{R}$ was set equal to the standard deviation computed in Step 3, and the process returned to Step 2.
The above algorithm was tested many times under different initial candidate values and consistently returned the same final estimates.

## Selectivity

As alluded to above, a total of eleven selectivity curves are specified by the GOA Pacific cod model. Three curves apiece are specified for the June-December trawl fishery and the longline fishery, corresponding to the time periods 1964-1986, 1987-1999, and 2000-2005. Two curves are specified for the January-May trawl fishery, corresponding to the time periods 1964-1999 and 2000-2005 (although a single selectivity curve is specified for the years 1964-1999 in the January-May trawl fishery, the parameters for this curve are estimated entirely from data collected during the 1987-1999 time period, because almost no size composition data were collected from the January-May trawl fishery during the 1964-1986 time period). Two curves are also specified for the pot fishery, corresponding to the time periods 1987-1999 and 2000-2005 (there was no significant pot fishery for Pacific cod prior to 1987). A single curve is specified for the GOA bottom trawl survey.

Although SS2 includes several options for specifying the functional form of the selectivity curve, the most flexible and commonly used option involves a pair of scaled logistic curves joined by a horizontal linear segment. The first (ascending) logistic curve begins at the minimum length specified in the data file ( 9 cm in the case of the GOA Pacific cod model), where the selectivity is less than 1.0 , and ends at some intermediate length, where selectivity is exactly 1.0 . A horizontal linear segment extends from the right-hand end of the first logistic to the left-hand end of the second logistic. Selectivity equals 1.0 throughout this linear segment. The second (descending) logistic curve begins at the end of the horizontal linear segment, where selectivity is still exactly 1.0 , and ends at the maximum length specified in the data file ( 110 cm in the case of the GOA Pacific cod model), where the selectivity is less than 1.0. This selectivity function is similar to the primary selectivity function used in SS1, except that the function used in SS1 omits the horizontal linear segment that joins the two logistic curves in the SS2 version of the function (i.e., selectivity in the SS1 version equals 1.0 at a single point only, whereas the SS2 version allows selectivity to equal 1.0 throughout a range of values).
Eight parameters are used to define the SS2 selectivity function: the size at which selectivity first reaches a value of 1.0 (peak location), the selectivity at the minimum length represented in the data (S(Lmin)), the logit transform of the size corresponding to the inflection of the ascending logistic curve (logit(infl1)), the relative slope of the ascending logistic curve (slope1), the logit transform of the size corresponding to the inflection of the descending logistic curve (logit(infl2)), the relative slope of the descending logistic curve (slope2), the logit transform of the selectivity at the maximum length represented in the data ( $\operatorname{logit(S(Lmax))),~and~the~width~of~the~length~range~at~which~selectivity~equals~} 1$ (peak width). The parameters are similar in the SS1 version of the selectivity function, except that peak width is implicitly set equal to zero.

## Prior Distributions

A potentially major difference between SS1 and SS2 is that SS2 is explicitly cast in a Bayesian framework, with specification of a prior distribution required for each parameter. Of course, a noninformative prior can be chosen for any or all parameters if so desired. However, use of informative priors is probably appropriate for many of the parameters in the GOA Pacific cod model, because one or both Plan Teams and the SSC have indicated in the past that certain values, or ranges of values, of various parameters are either relatively likely or unlikely. For example, the SSC has indicated that a natural mortality rate of 0.37 is likely close to the true value (SSC minutes, December 1994). As another example, the BSAI Plan Team has expressed concern that previous assessments' estimates of large-fish selectivity in the EBS shelf bottom trawl survey may be too low (BSAI Plan Team minutes, November 2004), and the GOA Pacific cod assessment has typically produced survey selectivity patterns similar to those obtained in the BSAI assessment. By utilizing a Bayesian framework, SS2 provides a logical means of integrating perspectives such as these into the stock assessment model. The specific priors used in this assessment are described under "Parameters Estimated Conditionally" below.

## Parameters Estimated Independently

## Natural Mortality

In the 1993 BSAI Pacific cod assessment (Thompson and Methot 1993), the natural mortality rate $M$ was estimated using SS1 at a value of 0.37 . All subsequent assessments of the BSAI and GOA Pacific cod stocks (except the 1995 GOA assessment) have used this value for $M$. Other published estimates of $M$ for Pacific cod are shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

As the above table indicates, the natural mortality rate for Pacific cod is either highly variable by time or area or it is very hard to estimate. In Models 1 and 2, $M$ is fixed at the traditional value of 0.37 . In Model 3 , $M$ is estimated internally.

## Trawl Survey Catchability

The base model used in all previous GOA Pacific cod assessments has fixed the catchability coefficient $(Q)$ for the GOA bottom trawl survey independently of other parameters at a value of 1.0. Somerton (2004) has shown that $Q$ for Pacific cod in the GOA is very unlikely to be greater than 1.0. In Models 1 and 2, the survey catchability is set equal to 1.0. In Model 3, the values of this parameter is estimated internally.

## Weight at Length

Parameters governing the allometric relationship between weight and length were estimated in a previous assessment by log-log regression from the available data (see "Data" above, with weights given in kg and lengths in cm ), giving a multiplicative constant of $5.80 \times 10^{-6}$ and an exponent of 3.159.

## Variability in Estimated Age

In keeping with the assumptions used in last year's assessment, Model 1 does not use age data. Models 2 and 3, however, do use age data. To use age data, SS2 requires an estimate of the standard deviation of estimated age at each age. Weighted least squares regression was used to estimate a linear relationship between standard deviation and age. The resulting estimates of the intercept and slope were 0.161 and 0.038 , respectively. This relationship was used in Models 2 and 3.

## Maturity at Length

As in previous assessments of GOA Pacific cod, the present assessment uses a single (i.e., time-invariant), length-based maturity schedule. Although the maturity schedule is constant within a given assessment, the values of the parameters describing the maturity schedule have changed over time. As noted in the "Data" section above, the GOA Pacific cod assessments have always relied on maturity-at-length data collected from the Bering Sea. As different Bering Sea data sets have been compiled over the years, the schedule of maturity at length has changed from time to time. The history of maturity schedules used previously or now available for use in the GOA Pacific cod assessment may be summarized as follows, where the length at $50 \%$ maturity ( $L 50$ ) and slope of the linearized logistic equation (A) are used to characterize each schedule:

1) From 1984 through 1994, the maturity schedule was based on gonado-somatic index values from the 1981-1982 surveys, with $L 50$ and $A$ values of 61 cm and -0.248 , respectively (Teshima 1985).
2) From 1995 through 2004, the maturity schedule was based on macroscopic observations ("scans") from the 1994 commercial fishery, with $L 50$ and $A$ values of 67 cm and -0.142 , respectively (Thompson 1995).
3) For this year's assessment, another possible candidate is Stark's (2005) maturity schedule, based on histological samples collected in the central GOA at various times of year during the period 1998-2004, with $L 50$ and $A$ values of 50 cm and -0.222 , respectively.

To provide some context for the above schedules, it may be helpful to consider alternative estimates. Two categories of alternative estimates are those derived from "rules of thumb" based on life history parameters and those derived from biological samples. The method suggested by Roff (1984), based on the Brody growth coefficient $K$ and the natural mortality rate, falls within the "rules of thumb" category. The available length-at-age data (see "Length at Age" under "Data" above) suggest a $K$ value of about 0.094 (based on maximum likelihood). Using this estimate of $K$ and the conventional Pacific cod $M$ value of 0.37, Roff's method implies an age at maturity of about 6 years, corresponding to a length of about 62 cm (based on linear interpolation). However, an age at maturity of 6 years seems very high given that the maximum age observed in the 2003 survey was only 12 years. Royce (1972) suggested another rule of thumb, namely, that the age at maturity should typically be less than one-third of the maximum age. For a maximum age of 12 years, Royce's method would imply an age at maturity around 4 years or lower, corresponding to a length of about 46 cm (based on linear interpolation). In the category of estimates derived from biological samples, Rovnina et al. (1997) estimated $L 50$ at $55-60 \mathrm{~cm}$ for Pacific cod in the Sea of Okhotsk, Welch and Foucher (1988) estimated $L 50$ at $45-55 \mathrm{~cm}$ for Pacific cod in British Columbia, and Hattori et al. (1992) estimated that $50 \%$ of Pacific cod in the Sea of Japan were mature by age 4 which, for GOA Pacific cod, corresponds to a length of about 46 cm (based on linear interpolation). All of these alternative estimates are closer to Stark's (2005) estimate of $L 50$ than the estimate used in recent assessments.

In addition to the above, the following reasons support use of Stark's (2005) maturity-at-length schedule (the "new schedule"):

1) The new schedule is based on a published methodology (Stark 2004) that is the source of the maturity schedules used in several other BSAI and GOA groundfish assessments (BSAI flathead sole, GOA flathead sole, GOA northern rock sole, GOA southern rock sole).
2) The author of the new maturity schedule has extensive experience in both macroscopic and histological estimation of Pacific cod maturity and is convinced that the histological methods are more accurate.
3) The method used to determine the maturity schedule used in recent Pacific cod assessments is subject to factors that might cause the resulting $L 50$ value to be biased high, whereas the method used to determine the new schedule is not subject to these factors, as described below.
Ova that contain yolk (mature ova) appear transparent, in contrast to the opaque appearance of ova that do not contain yolk. The success of macroscopic maturity classification systems depends in part on the ability of observers to distinguish transparent ova from opaque ova. This distinction can be difficult to make, because the ova are not observed directly, but through the ovary wall. The difficulty is greater for smaller fish, and can bias classifications of smaller fish in favor of immaturity. The reason for this is that, as fish grow, the number of ova contained in each ovary increases more than proportionally, which in turn causes greater distention of the ovary wall when yolk accumulates in the ova. Greater distention results in greater transparency of the ovary wall, which in turn increases observers' ability to identify transparent ova through the ovary wall. Conversely, it is harder for observers to detect mature ova in smaller fish, because the ovary wall is typically less distended than in large fish, even when the ova are mature. For the same reason (disproportionately less stretching of the ovary wall in small fish), macroscopic observation of small fish that have already spawned during the year may result in an incorrect classification of "developing" or "immature" because it is difficult to detect the presence of disintegrating ova (a criterion used to distinguish "spent" ovaries) through the ovary wall in such fish.

In contrast, histological maturity classifications are not subject to these biases because the maturity classifications are based on a comprehensive microscopic assessment of each ovum and associated structures, such as post ovulatory follicles, contained within each ovary section. The examinations are conducted under controlled laboratory conditions. The probability of detecting yolk within an ovary is very high because all ovary slide sections are stained with eosin dye which attaches to any yolk protein present, giving it a distinctive pink coloration.

## Parameters Estimated Conditionally

With a few exceptions, Models 1, 2, and 3 estimate similar parameters, although the number of parameters of a given type estimated by the three models may differ in some cases due to the fact that Model 1 sets the initial year at 1978 while Models 2 and 3 set the initial year at 1964. The parameters that all three models attempt to estimate internally consist of the following:

1) mean length at age 1.5 , mean length at age 12 , Brody growth coefficient $K$
2) log-scale mean recruitment for the post-1976 environmental regime
3) annual recruitments (Model 1) or annual recruitment deviations (Models 2 and 3)
4) selectivity parameters ( 7 for Model 1,8 for Models 2 and 3 ) for each of 11 selectivity curves
5) initial fishing mortality (initial year $=1978$ for Model 1, 1964 for Models 2 and 3)
6) year-, gear-, and season-specific fishing mortality rates

It should be noted that the fishing mortality rates in (6) are somewhat different from the other parameters in that their values are determined exactly given the values of the other parameters and the input catch data, which are assumed to be true values rather than estimates.

In addition to the parameters estimated internally by all three models, the following parameters are estimated by some subset of the three models:

1) Model 1 estimates each element of the initial numbers-at-age vector.
2) Models 2 and 3 estimate the log-scale mean recruitment for the pre-1977 environmental regime and the standard deviation of the recruitment deviations (though not quite internally, but rather through an iterative process described under "Model Structure" above).
3) Model 3 estimates the natural mortality rate $M$ and the logarithm of the bottom trawl survey catchability coefficient $Q$.

In the case of Model 1, the estimator used is the peak of the logarithm of the likelihood function (see below). In the cases of Models 2 and 3, the estimator used is the mode of the logarithm of the joint posterior distribution, which is in turn calculated as the sum of the logarithms of the parameter-specific prior distributions (see below) and the logarithm of the likelihood function.

## Prior Distributions

For the two models developed under SS2 in this year's assessment, the informative prior distributions described in the following paragraphs were specified (all distributions are normal):

Parameters with priors based on a specified coefficient of variation (CV)
Initial fishing mortality: The mean was set at 0.1 , reflecting the conventional wisdom that the stock was lightly exploited during the 1960s. The standard deviation was set at 0.03 , corresponding to a CV of 30\%.

Selectivity parameter $S($ Lmin): For the commercial fisheries, this was not an estimated parameter, but was set at a fixed value of 0.001 . This choice was based on the fact that almost no fish in the sub- 30 cm range are taken in the commercial fisheries and because preliminary model runs invariably resulted in this parameter being bound at whatever minimum value was specified. For the surveys, the prior distribution was assigned a mean of 0.1 and a standard deviation of 0.03 , corresponding to a $30 \% \mathrm{CV}$. In contrast to the commercial fisheries, $10 \%$ of the average bottom trawl survey size composition has consisted of fish smaller than 30 cm .

Selectivity parameters slope1 and slope2: These two parameters had identical priors, with the mean set at 0.2 and the standard deviation set at 0.06 , corresponding to a $30 \%$ CV. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of these parameters.

Selectivity parameter peak width: The mean was set at 10 and the standard deviation was set at 3 , corresponding to a $30 \% \mathrm{CV}$. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of this parameter, in addition to results from preliminary model runs which indicated that values much higher than 10 tended to cause the model to get "stuck."
Log survey catchability $\ln (Q)$ : Model 3 treats $\ln (Q)$ as a free parameter with a prior distribution. This prior distribution was assigned a mean of -0.29 and a standard deviation of 0.05 , corresponding to a lognormal prior for $Q$ with a median of 0.75 and CV of $5 \%$. The choices of mean and standard deviation for this prior distribution were difficult ones. In previous assessments, $Q$ has always been fixed at a value of 1.0 , which equates to a log value of 0 . One natural way to convert a fixed constant into a free parameter with a normal prior is to treat the former fixed value as the mean of the new prior distribution and set a reasonable value for the standard deviation. However, this is not the only logical option. In the case of $\ln (Q)$ for GOA Pacific cod, for example, a value of 0 was used in last year's assessment not only because it was consistent with the results of Somerton (2004), but because it was the upper limit implied by those results (i.e., the results showed that it is very unlikely for the true value of $\ln (Q)$ to be positive). If the former fixed value of 0 is viewed as an upper limit, it does not make sense to treat it as the mean of the new prior. Rather, the mean and standard deviation of the new prior distribution should be set so that exceeding the upper limit is highly unlikely. The choice of -0.29 as the mean for the prior distribution was largely a subjective one, although it has some support in that the implied median value of 0.75 for $Q$ is close to, and midway between, the estimates of 0.7 and 0.8 obtained for this parameter in the 2001 and 2004 GOA walleye pollock assessments (Dorn et al. 2001 and 2004). The choice of 0.05 for the standard
deviation was pragmatic. A standard deviation of 0.05 probably underestimates the true uncertainty that ought to be associated with this prior distribution. However, preliminary model runs with higher values for the standard deviation inevitably resulted in point estimates for $Q$ that were much higher than 1.0 (often in the neighborhood of 2.0), which cannot presently be reconciled with the results of Somerton (2004).

Natural mortality M: As with $\ln (Q)$, Model 2 fixes $M$ at the value used in last year's assessment (0.37), while Model 3 treats it as a free parameter with a prior distribution. The prior distribution was assigned a mean of 0.37 and a standard deviation of 0.019 , corresponding to a CV of $5 \%$. Similar to the situation with $\ln (Q)$, the choice of 0.019 for the standard deviation was a pragmatic one. Although it probably underestimates the true uncertainty that ought to be associated with this prior distribution, preliminary model runs in this year's BSAI Pacific cod assessment (Thompson and Dorn 2005) with higher values for the standard deviation inevitably resulted in point estimates for $M$ that were much lower than 0.37 (often in the neighborhood of 0.20 ), which are so far from the traditionally accepted value that it does not seem wise to accept them without further investigation. Furthermore, higher values of the standard deviation for $M$ tended to push the point estimates of $Q$ in the BSAI assessment to very high values that cannot be presently be reconciled with the results of Somerton (2004). Because it would be hard to justify different standard deviations for the prior distribution on $M$ in the EBS and GOA given current knowledge, a CV of $5 \%$ was used for this prior in both the EBS and GOA assessments.

Parameters with priors based on one or both endpoints of the $98 \%$ confidence interval Selectivity parameters logit(infl1) and logit(infl2): These two parameters had identical priors, with the mean set at 0 and the standard deviation set at 0.944 . The mean corresponds to an inflection point located midway between Lmin and peak location, in the case of infl1, or between peak location and Lmax, in the case of infl2. The mean and standard deviation together imply a $98 \%$ confidence interval extending from $10 \%$ to $90 \%$ of the difference between Lmin and peak location, in the case of infl1, or between peak location and Lmax, in the case of infl2. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of these parameters.
Selectivity parameter logit(S(Lmax)): The mean was set at 2.197 and the standard deviation was set at 0.944 . The mean corresponds to a selectivity of 0.9 at Lmax. The mean and standard deviation together imply a $1 \%$ chance of selectivity at $\operatorname{Lmax}$ being less than 0.5 . These parameter values were chosen in part to reflect the Plan Team's belief that selectivity of large fish in the bottom trawl survey should be fairly high.

Parameters with priors based on the data
Length at age parameters: Mean values for length at age 1.5, length at age 12, and the Brody growth coefficient $K$ were set at 20.7, 95.5 , and 0.094 , respectively, corresponding to the maximum likelihood estimates obtained from the data collected during the 2003 GOA bottom trawl survey. The standard deviations for these parameters were set at $0.232,1.616$, and 0.008 , respectively, corresponding to the values associated with the inverted Hessian matrix obtained in the process of estimating the means.

Selectivity parameter peak location: The mean and standard deviation were set individually for each selectivity curve by identifying the length associated with the maximum frequency in each length frequency record, then computing the mean and standard deviation for each respective gear type and portion of the time series. This was done in order to give the model a reasonable starting value and place reasonable constraints on peak location, a parameter which is typically very difficult to estimate. The SS2 User Manual suggests that this parameter "should be an integer and should be at bin boundary and not estimated," but it also suggests that recent improvements to the code "may allow estimation" (Methot 2005b). Extensive testing during preliminary runs of the EBS Pacific cod model (Thompson and Dorn 2005) revealed that the value of this parameter can be quite important in determining model results and that free estimation (with a reasonably strong prior) was much more likely to find an optimal value than
profiling manually over the range of possible integer values, especially considering the practical difficulty of manually tuning 11 parameters (one peak location for each selectivity curve) at the same time. The resulting means (cm) and standard deviations (cm) for peak location in each of the 11 selectivity curves were as follow:

| Fishery/Survey | Years | Mean | Std. Dev. |
| :--- | :---: | :---: | :---: |
|  | $1964-$ |  |  |
| Jan-May Trawl | 1999 | 63.8 | 2.12 |
|  | $2000-$ |  |  |
| Jan-May Trawl | 2005 | 61.1 | 4.32 |
|  | $1964-$ |  |  |
| Jul-Dec Trawl | 1986 | 54.1 | 5.71 |
|  | $1987-$ |  |  |
| Jul-Dec Trawl | 1999 | 58.7 | 5.68 |
|  | $2000-$ |  |  |
| Jul-Dec Trawl | 2005 | 61.4 | 5.22 |
|  | $1964-$ |  |  |
| Longline | 1986 | 58.3 | 6.96 |
|  | $1987-$ |  |  |
| Longline | 1999 | 63.1 | 3.10 |
| Longline | $2000-$ |  |  |
|  | 2005 | 63.1 | 2.42 |
| Pot | $1987-$ |  |  |
|  | 1999 | 64.5 | 2.53 |
| Pot | $2000-$ |  |  |
| Bottom Trawl Survey | 2005 | 61.9 | 2.99 |
|  | 2005 | 54.5 | 3.60 |

## Likelihood Components

Likelihood components included in all three models are of four types: size composition, age composition, survey biomass, and mean size at age. There are five size composition components in the likelihood: one each for the January-May trawl fishery, the June-December trawl fishery, the longline fishery, the pot fishery, and the bottom trawl survey. There is only one age composition component and one size-at-age component in the likelihood, because all age data currently come from the trawl survey. There is one survey biomass component in the likelihood, corresponding to the bottom trawl survey. In addition to the above, Models 2 and 3 include a recruitment deviations component.

Both SS1 and SS2 allow the user to specify "emphasis" factors that determine which components receive the greatest attention during the parameter estimation process. As in previous assessments, each component in each model was given an emphasis of 1.0 in the present assessment.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery, and time period within the year. In the parameter estimation process, SS1 and SS2 weight a given size composition observation (i.e., the size frequency distribution observed in a given year, gear/fishery, and period) according to the emphasis associated with the respective likelihood component and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which SS1 was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to compensate for
contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. As in previous assessments, the present assessment uses a multinomial sample size equal to the square root of the true length sample size, rather than the true length sample size itself. Given the true length sample sizes observed in the present assessment, this procedure tends to give values somewhat below 400 while still providing the SS1 and SS2 programs with usable information regarding the appropriate effort to devote to fitting individual length samples. Multinomial length sample sizes derived by this procedure for the commercial fishery size compositions are shown in Table 2.13. In the case of GOA bottom trawl survey size composition data, the square root assumption was also used. The square roots (sqrt) of the true survey length sample sizes are shown below:

| Year | sqrt(N) |
| :--- | :--- |
| 1984 | 132 |
| 1987 | 140 |
| 1990 | 107 |
| 1993 | 131 |
| 1996 | 110 |
| 1999 | 93 |
| 2001 | 82 |
| 2003 | 96 |
| 2005 | 83 |

## Use of Age Composition Data in Parameter Estimation

Like the size composition data, the age composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery (in this case, the GOA bottom trawl survey), and time period within the year (in this case, the June-August period). However, selection of an appropriate input sample size is more complicated for age composition data than for length composition data, because age composition data are generated not only from the set of otolith readings but from the estimated size composition as well. Therefore, even if a square root transformation is appropriate for size composition data, taking the square root of the number of otoliths read may underestimate the weight that should be given to the age composition data. Last year's assessment BSAI Pacific cod assessment (Thompson and Dorn 2004) introduced a method for setting an input sample size appropriate to age composition, a method which is retained in the present assessment. The steps are as follow:

1) The proportions of age at length are assumed to be approximately multivariate normally distributed, with a variance-covariance matrix determined by the matrix of proportions and the number of otoliths actually read at each length. A set of 10,000 random age-length keys was then simulated.
2) Survey numbers at each length are assumed to be approximately lognormally distributed with a mean equal to the point estimate and for that length and a constant (across lengths) coefficient of variation (CV) equal to the amount that sets the sum of the variances in numbers at length equal to the variance of the survey estimate of population size. A set 10,000 of random numbers-atlength distributions was then simulated.
3) For each combination of randomly simulated age-key and numbers-at-length distribution, an effective sample size was computed.
4) The "true" input sample size was set equal to the harmonic mean of the distribution of randomly simulated effective sample sizes, based on the asymptotic equivalence of these two quantities. The following table was thereby obtained for the age composition data (the last row shows the value used as "true" input sample sizes):

| Year | 2003 |
| :--- | ---: |
| Number of fish aged: | 711 |
| Square root of number of fish aged: | 27 |
| CV of numbers at length: | 0.83 |
| Harmonic mean effective sample size: | 80 |

Note that this procedure gives an input sample size larger than would be achieved simply by taking the square root of the number of fish aged (third row in the above table). This reflects the added precision achieved by use of both age-at-length and numbers-at-length data in constructing a numbers-at-age estimate. To avoid double counting of the same data, Models 2 and 3 ignore length composition data from the 2003 GOA bottom trawl survey.

It may be noted that the harmonic mean effective sample size computed above (80) is smaller than the sample size (96) obtained for the corresponding length composition using the "square root method" in the preceding subsection, suggesting that the two methods of computing sample sizes are not entirely consistent. This is not surprising, given that the square root method was adopted only as a simple approximation in the first place, but it does suggest a need for further work in this area.

## Use of Size-at-Age Data in Parameter Estimation

Each size at age datum is assumed to be drawn from a normal distribution specific for that age and year. The model's estimate of mean size at age serves as the mean for that year's distribution, and the standard deviation is inversely proportional to the sample size (Methot 2000, Methot 2005a).

## Use of Survey Biomass Data in Parameter Estimation

Each year's survey biomass datum is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey biomass in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey biomass datum's standard error to the survey biomass datum itself serves as the distribution's coefficient of variation.

## Use of Recruitment Deviation "Data" in Parameter Estimation

The recruitment deviations likelihood component is different from traditional likelihoods because it does not involve "data" in the same sense that traditional likelihoods do. Instead, the log-scale recruitment deviation plays the role of the datum and the log-scale recruitment mean and $\sigma_{R}$ play the role of the parameters in a normal distribution, but, of course, all of these are treated as parameters by SS2.

## MODEL EVALUATION

As described in the preceding section, three alternative models are evaluated in the present assessment. Model 1 is identical to the SS1 model used in the 2004 assessment, where the natural mortality rate $M$ and the trawl survey catchability coefficient $Q$ were fixed at values of 0.37 and 1.0 , respectively. Model 2 is developed under SS2 and differs from Model 1 in several respects, such as use of an earlier initial year and use of prior distributions for many model parameters, but retains Model 1's assumptions regarding the values of $M$ and $Q$. Model 3 is also developed under SS2 and is identical to Model 2, except that the values of $M$ and $Q$ are estimated rather than fixed.

## Evaluation Criteria

In previous GOA Pacific cod assessments, evaluation criteria have typically focused on effective sample sizes of the size composition data (and, more recently, the age composition data), the root mean squared error (RMSE) of the fit to the survey biomass data, and the overall reasonableness of the parameter values. These criteria are retained in the present assessment, not so much to determine which one of the three models is "best," but as a check to see whether any of the three can reasonably be rejected. Given that a model passes these tests, two additional evaluation criteria are as follow:

1) Do the model's estimates of total biomass achieve a reasonable relationship with the trawl survey's estimates of biomass? (This is different from the question of how well the model's estimates of survey biomass fit the survey's own estimates, which is addressed by the RMSE.)
2) Does the model appropriately reflect the uncertainty associated with key assessment outputs?

## Effective Sample Size

Once maximum likelihood estimates of the model parameters have been obtained, SS1 and SS2 compute an "effective" sample size for the size or age composition data specific to a particular year, gear/fishery, and time period within the year. Roughly, the effective sample size can be interpreted as the multinomial sample size that would typically be required in order to produce the given fit. More precisely, it is the sample size that sets the sum of the marginal variances of the proportions implied by the multinomial distribution equal to the sum of the squared differences between the sample proportions and the estimated proportions (McAllister and Ianelli 1997). As a function of a multinomial random variable, the effective sample size has its own distribution. The harmonic mean of the distribution is asymptotically equal to the true sample size in the multinomial distribution. Thus, if the effective sample size is less than the true sample size in the multinomial distribution, it is reasonable to conclude that the fit is not as good as expected. The following table shows the average of the input sample sizes (Ninp), the average of the effective sample sizes (Neff), and the ratio of Neff to Ninp for each of the size composition components and the trawl survey age composition component in each of the three models:

|  | Model 1 |  |  | Model 2 |  |  | Model 3 |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Likelihood component | Ninp | Neff | Ratio | Ninp | Neff | Ratio | Ninp | Neff | Ratio |
| Jan-May trawl fishery length | 120 | 279 | 2.33 | 120 | 332 | $\mathbf{2 . 7 8}$ | 120 | 322 | 2.69 |
| Jun-Dec trawl fishery length | 38 | 80 | 2.09 | 37 | 88 | $\mathbf{2 . 3 4}$ | 37 | 87 | 2.32 |
| Longline fishery length | 82 | 367 | 4.50 | 82 | 439 | $\mathbf{5 . 3 7}$ | 82 | 423 | 5.18 |
| Pot fishery length | 90 | 363 | $\mathbf{4 . 0 3}$ | 90 | 264 | 2.93 | 90 | 266 | 2.95 |
| Non-2003 shelf survey length | 110 | 151 | $\mathbf{1 . 3 8}$ | 110 | 124 | 1.13 | 110 | 129 | 1.17 |
| 2003 shelf survey length | 96 | 231 | 2.41 | -96 | -150 | -1.56 | -96 | -157 | -1.63 |
| 2003 shelf survey age | n/a | n/a | n/a | 80 | 39 | $\mathbf{0 . 4 9}$ | 80 | 39 | 0.48 |

Notes:

1) For each row, the average values of Ninp and Neff are computed with respect to all years and periods present in the respective time series.
2) The average input sample size for the Jun-Dec trawl fishery lengths in Model 1 is slightly different from those in Models 2 and 3 because more years of data are included in Models 2 and 3.
3) The negative values in the row for 2003 trawl survey lengths indicate that those data are "turned off" in the Models 2 and 3 to avoid double-counting of length data in the year with age data.
4) Bold font indicates the maximum ratio for the respective row.

All three models produce average effective sample sizes larger than the average input values for all length components. Neither Model 2 nor Model 3 produces an average effective sample size greater than the average input value for the trawl survey age component (the age component is not applicable to Model 1), which is somewhat disappointing. Of the six components (not counting the 2003 trawl survey length component), Model 1 had the highest ratio in two cases, Model 2 had the highest ratio in four cases, and Model 3 never had the highest ratio. However, nearly all of the differences between models are extremely small. It should also be noted that the use of prior distributions by Models 2 and 3 might be expected to cause those models to perform less well than Model 1 with respect to likelihood components such as these, but this does not appear to be the case, as both Models 2 and 3 outperformed Model 1 (though only slightly) in three of the five likelihood components shared by all three models. In summary, the main conclusion to be drawn from the above table is that all three models are performing reasonably well with respect to most or all of the size composition components.

## Fit to Survey Biomass Data

The average value of the lognormal "sigma" parameter in the trawl survey biomass data is 0.169 . The log-scale root-mean-squared-errors (log-scale RMSEs) from Models 1, 2, and 3 are 0.162, 0.195, and 0.190 , respectively. Although Model 1 performs slightly better than the other two models, all three logscale RMSEs are close to the value of the average sigma. With respect to the trawl survey point estimates of biomass, all three models had the greatest difficulty fitting the 1984 and 1996 values (all three models indicate that the survey point estimates in those two years were likely overestimates).

## Reasonableness of Parameter Values

Although hundreds of parameters are estimated by all three models, three items of special interest are the natural mortality rate $M$, the survey catchability coefficient $Q$, and the trawl survey's selectivity at Lmax. The values of these parameters (to two significant digits) for each of the three models are shown below:

| Parameter | Model 1 | Model 2 | Model 3 |
| :--- | ---: | ---: | ---: |
| $M$ | 0.37 | 0.37 | 0.38 |
| $Q$ | 1.00 | 1.00 | 0.78 |
| $S($ Lmax $)$ | 0.29 | 0.60 | 0.50 |

Of course, the values of $M$ and $Q$ in Models 1 and 2 are fixed rather than estimated, so presumably those values are reasonable, although the $Q$ value of 1.0 from Models 1 and 2 should probably be considered to be at the upper limit of the reasonable range. The values of $M$ and $Q$ in Model 3 are estimated, but with very tight prior distributions, and their final estimates are extremely close to their respective prior medians. The values of $S(\operatorname{Lmax})$ produced by the three models provide more contrast than the values of $M$ and $Q$. Model 1's estimate of 0.29 is similar to the value of 0.25 estimated in last year's assessment. The estimates given by Models 2 and 3 are higher and seem easier to reconcile with the design of the survey.

## Relationship of Total Biomass to Survey Biomass

The time series of age 3+ biomass, spawning biomass, and survey biomass estimated by the three models, along with the observed survey biomass time series, in Table 2.14. The past several assessments have tended to result in estimates of age 3+ biomass that were much greater than the survey biomass. All three models in the present assessment behave likewise, although the biomass estimates produced by Model 2 tend not to be as high as those produced by Models 1 and 3. On average, the estimates of age 3+ biomass exceed the observed survey biomass by about $80 \%$, $31 \%$, and $78 \%$ for Models 1 , 2 , and 3 , respectively. While it is possible to imagine mechanisms that could cause the bottom trawl survey to underestimate the total biomass of Pacific cod (e.g., a large portion of the population occurring in the water column above the headrope), the existence of any such mechanism has yet to be verified experimentally. Until such
verification takes place, age 3+ biomass estimates in the neighborhood of those produced by Model 2 should probably be viewed as more realistic than estimates in the neighborhood of those produced by Models 1 and 3 , all else being equal.

## Characterization of Uncertainty

One of the main drawbacks of SS1 is that it does not include utilities for estimating the statistical uncertainty surrounding derived quantities such as spawning biomass. Because the SS1-based Model 1 provides only point estimates, it can represent uncertainty adequately only if the true uncertainty is very small or if the most important uncertainties consist of natural random variability rather than statistical imprecision. However, because SS2 is coded in ADMB, it provides for straightforward estimation of the statistical uncertainty surrounding any quantity of interest, which gives some hope that the SS2-based Models 2 and 3 can do an adequate job of describing uncertainty. As an example, the three models' estimates of spawning biomass for the years 1978-2005 (the years that all three models have in common) are shown in Figure 2.6, together with $95 \%$ confidence intervals for Models 2 and 3. The relative trend of the point estimates is qualitatively similar across models although the magnitudes differ, with Model 1 giving the highest estimates during the years 1978-1988, Model 3 giving just slightly higher estimates than Model 1 during the years 1989-2005, and Model 2 consistently giving the lowest estimates. From the point of view of uncertainty, however, the key feature of Figure 2.6 is that the confidence intervals from Model 3 are noticeably broader than the confidence intervals from Model 2 (Model 1, of course, cannot generate confidence intervals). Averaged across the entire time series (1964-2005), the confidence intervals from Model 3 are 52\% wider than those from Model 2. The confidence intervals from Model 3 encompass the point estimates from Model 1 for every year from 1984 to the present and they substantially overlap the confidence intervals from Model 2 in all years. The fact that Model 3 produces wider confidence intervals than Model 2 is likely due to the fact that natural mortality and survey catchability are estimated in Model 3 but not Model 2.

A related issue has to do with whether statistical estimates of key parameters such as $M$ and $Q$ are preferable to assumed values. It should be remembered that the traditional $M$ value of 0.37 was produced by a model very similar to Model 3, but with fewer data and a less sophisticated estimation algorithm. Although the prior distributions used by Model 3 to help estimate $M$ and $Q$ were very tight, at least they were broader than the prior distributions implicitly assumed by Models 1 and 2, each of which had a CV of zero.

## Selection of Final Model

Evaluation of the three models using the above criteria may be summarized as follows: 1) For the length composition likelihood components, all three models performed reasonably well in all categories and performed extremely well in at least some categories. 2) For the age composition likelihood component, neither Model 2 nor Model 3 performed very well (Model 1 did not use age data). 3) For the fit to the survey biomass time series, all three models performed approximately the same, with RMSEs close to the value that would be predicted from the sampling variation in the surveys. 4) For the overall reasonableness of the parameter values, all three models are associated with reasonable values of $M$ and $Q$, although the $Q$ value of 1.0 from Models 1 and 2 should probably be considered to be at the upper limit of the reasonable range. Also, Model 3 has the advantage of being associated with values of $M$ and $Q$ that are not only reasonable but estimated (albeit with very tight priors) rather than assumed. Model 2 probably gives the most reasonable estimates of large-fish selectivity in the trawl bottom trawl survey, followed fairly closely by Model 3. 5) Relative to the survey biomass time series, the estimated age 3+ biomass time series obtained under Model 2 is considerably closer than the time series obtained under Models 1 or 3. 6) Regarding characterization of uncertainty, Models 2 and 3 obviously perform better than Model 1, because Model 1 was not designed to produce estimates of uncertainty. Model 3's confidence intervals around spawning biomass are much wider than those for Model 2. Given that Model

2's confidence intervals are predicated on the assumption that $M$ and $Q$ are known with certainty whereas Model 3's confidence intervals do not make this assumption, Model 3's representation of uncertainty is probably more realistic.

Whether Model 2 or Model 3 is the best choice depends on how the above evaluation criteria are weighted. Assuming that a high priority is placed on achieving a realistic representation of uncertainty (both in estimates of parameters such as $M$ and $Q$ and in derived quantities such as spawning biomass), Model 3 appears to be the best choice overall.

## Final Parameter Estimates and Associated Schedules

Final estimates of some key scalar parameters are shown below:

| Parameter | Value |
| :--- | ---: |
| Length at age 1.5 | 20.6 |
| Length at age 12 | 88.3 |
| Brody growth coefficient K | 0.117 |
| Natural mortality rate M | 0.38 |
| Trawl survey catchability | 0.78 |
| Q |  |
| Recruitment variability $\sigma$ R | 0.469 |

Estimates of fishing mortality rates are shown in Table 2.15, estimates of regime-specific median recruitments and annual recruitment deviations are shown in Table 2.16, and estimates of selectivity parameters are shown in Table 2.17.

Schedules of selectivity at length are shown for the commercial fisheries and bottom trawl survey in Table 2.18. The schedules in Table 2.18 are plotted in Figure 2.7. As examples of how the schedules of selectivity at length translate into size compositions, Figures 2.8 a, 2.8 b, 2.8 c, and 2.9 show observed and estimated size compositions from the 2003 January-May fisheries, the 2004 January-May fisheries, the 2005 January-May fisheries, and the 2001-2005 trawl surveys, respectively.

Schedules of selectivity at age for the most recent portion of the time series are shown in Table 2.19 and Figure 2.10. To demonstrate how the schedules of selectivity at length translate into age compositions, Figure 2.11 shows observed and estimated age compositions from the 2003 trawl survey. Schedules of length at age, proportion mature at age, and weight at age are shown in Table 2.20.

## RESULTS

## Definitions

The biomass estimates presented here will be defined in three ways: 1 ) age 3+ biomass, consisting of the biomass of all fish aged three years or greater in January of a given year; 2) spawning biomass, consisting of the biomass of all spawning females in a given year; and 3) survey biomass, consisting of the biomass of all fish that the model estimates should have been observed by the survey in July of a given year. The recruitment estimates presented here will be defined as numbers of age 0 fish in a given year. The fishing mortality rates presented here will be defined as full-selection, instantaneous fishing mortality rates expressed on a per annum scale.

## Biomass

Model 3's estimated time series (1977-2005) of GOA Pacific cod age 3+ biomass and spawning biomass are shown in Table 2.21, together with estimates provided in last year's SAFE report (Thompson et al. 2004) and $95 \%$ confidence intervals for the spawning biomass estimates from Model 3 . The biomass
trends (age 3+, spawning, and survey) estimated in the present assessment are also shown in Figure 2.12, with $95 \%$ confidence intervals for the spawning biomass estimates. The model's estimated age $3+$ biomass shows a near-continual decline from about 1990 through the present. The model's estimated spawning biomass shows a near-continual decline from about 1995 through the present. It should be noted that the new maturity schedule used this year led to a significant increase in the overall magnitude of estimated spawning biomass.

## Recruitment

Model 3's estimated time series (1977-2004) of age 0 recruitment is shown in Table 2.22, together with estimates inferred from last year's SAFE report (Thompson et al. 2004) and $95 \%$ confidence intervals for this year's estimates. Because last year's assessment used 1 as the initial age in the model, age 0 recruitments for last year's assessment were inferred here by multiplying last year's estimates of age 1 recruits by $\exp (0.37)$, where 0.37 is the value of the natural mortality rate used in last year's assessment. Values for this year's assessment that exceed Model 3's estimate of the 1977-2004 average recruitment of 401,982,571 fish are shown in bold in Table 2.22.
This year's recruitment estimates for the entire time series (1964-2004) are shown in Figure 2.13, along with their respective $95 \%$ confidence intervals and regime-specific averages. For the time series as a whole, the largest year class was the 1977 cohort. Other exceptional year classes include those spawned in 1980, 1984, 1987, 1989, and 1995. Of the 15 year classes that have followed the strong 1989 year class, only four (1990, 1995, 1999, and 2000) have point estimates higher than the 1977-2004 average, and only one (1995) has a confidence interval that falls entirely above the 1977-2004 average. Although the upper ends of the confidence intervals for all but four of these post-1989 year classes exceed the 1977-2004 average recruitment, three of the four most recent year classes (2001, 2002, and 2003) happen to be among those that do not. As these dramatically sub-par cohorts work their way through the age structure in the coming years, continued decreases in stock biomass are likely.

To date, it has not been possible to estimate a reliable stock-recruitment relationship for this stock. With the move to SS2, prospects for future estimation of such a relationship should improve. In the interim, Figure 2.14 is provided to give some indication of the relationship between stock and recruitment. The Ricker (1954) curve shown in this figure was not fit statistically, but rather by assuming that $F_{35 \%}$ and $B_{35 \%}$ correspond to $F_{M S Y}$ and $B_{M S Y}$, respectively. This curve is intended to be illustrative only, and is not recommended for management purposes.

## Exploitation

The model's estimated time series of the ratio between EBS catch and age 3+ biomass is shown in Table 2.23 , together with the estimates provided in last year's SAFE report (Thompson et al. 2004). The average value of this ratio over the entire time series is about 0.08 , slightly higher than the average value of 0.07 obtained in last year's assessment. The estimated values exceed the average for every year after 1989 except 1993 and 1994, whereas none of the estimated values exceed the average in any year prior to 1990 except for 1980 and 1981. This finding is basically similar to that obtained in last year’s assessment.

Figure 2.15 plots the trajectory of relative fishing mortality and relative female spawning biomass from 1977 through 2005 based on Model 3, overlaid with the current harvest control rules (fishing mortality rates in the figure are standardized relative to $F_{35 \%}$ and biomasses are standardized relative to $B_{100 \%}$ ). The entire trajectory lies underneath both control rules. This figure indicates that, in spawning-per-recruit terms, the stock has been fished more lightly than previously thought (Thompson et al. 2004), a result due largely to the new maturity curve used in the present assessment.

## PROJECTIONS AND HARVEST ALTERNATIVES

## Amendment 56 Reference Points

Amendment 56 to the GOA Groundfish Fishery Management Plan (FMP) defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL ( $F_{O F L}$ ), the maximum permissible ABC, and the fishing mortality rate used to set the maximum permissible $A B C$. The fishing mortality rate used to set ABC ( $F_{A B C}$ ) may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the GOA are managed under Tier 3 of Amendment 56. Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

3a) Stock status: $B / B_{40 \%}>1$
$F_{\text {OFL }}=F_{35 \%}$
$F_{A B C} \leq F_{40 \%}$
3b) Stock status: $1 / 20<B / B_{40 \%} \leq 1$
$F_{O F L}=F_{35 \%} \times\left(B / B_{40 \%}-1 / 20\right) \times 20 / 19$
$F_{A B C} \leq F_{40 \%} \times\left(B / B_{40 \%}-1 / 20\right) \times 20 / 19$
3c) Stock status: $B / B_{40 \%} \leq 1 / 20$
$F_{\text {OFL }}=0$
$F_{A B C}=0$
Estimation of the $B_{40 \%}$ reference point used in the above formulae requires an assumption regarding the equilibrium level of recruitment. In this assessment, it is assumed that the equilibrium level of recruitment is equal to the post-1976 average (i.e., the arithmetic mean of all estimated recruitments from year classes spawned in 1977 or later). Other useful biomass reference points which can be calculated using this assumption are $B_{100 \%}$ and $B_{35 \%}$, defined analogously to $B_{40 \%}$. These reference points are estimated as follows:

| Reference point: | $B_{35 \%}$ | $B_{40 \%}$ | $B_{100 \%}$ |
| ---: | ---: | ---: | ---: |
| Value: | $115,000 \mathrm{t}$ | $132,000 \mathrm{t}$ | $329,000 \mathrm{t}$ |

For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. For this assessment, the apportionment was based on Model 3's estimates of fishing mortality by gear for the three most recent complete years of data (2002-2004). The average fishing mortality rates for those years implied that total fishing mortality was divided among the three main gear types according to the following percentages: trawl $30.2 \%$, longline $22.6 \%$, and pot $47.2 \%$. This apportionment results in estimates of $F_{35 \%}$ and $F_{40 \%}$ equal to 0.69 and 0.56 , respectively. The differences between these values and last year's estimates of 0.36 and 0.31 are due largely to the use of the new maturity schedule in this year's assessment.

## Specification of OFL and Maximum Permissible ABC

GOA Pacific cod spawning biomass for 2006 is estimated at a value of $165,000 \mathrm{t}$. This is about $26 \%$ above the $B_{40 \%}$ value of $132,000 t$, thereby placing Pacific cod in sub-tier "a" of Tier 3. Given this, the model estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2006 as follows:

| Quantity | Overfishing Level | Maximum Permissible ABC |
| :--- | ---: | ---: |
| Catch: | $130,000 \mathrm{t}$ | $109,000 \mathrm{t}$ |
| Fishing mortality rate: | 0.69 | 0.56 |

The age $3+$ biomass estimate for 2006 is $453,000 \mathrm{t}$.

## ABC Recommendation

Review of Past Approaches
For the years 1997-1999, the GOA Pacific cod assessments advocated a harvest strategy that attempted to address some of the statistical uncertainty in the assessment model, namely the uncertainty surrounding parameters the natural mortality rate $M$ and survey catchability $Q$ (Thompson et al. 1997, 1998, 1999). For the 2000-2003 assessments, the strategy was simplified by assuming that the ratio between the recommended $F_{A B C}$ and $F_{40 \%}$ estimate given in the 1999 assessment ( 0.87 ) was an appropriate factor by which to multiply the current maximum permissible $F_{A B C}$ to obtain a recommended $F_{A B C}$ (Thompson et al. 2003). By the time of the 2004 assessment, however, concern arose that the $87 \%$ adjustment factor might have outlived its usefulness, given that the survey time series had changed appreciably since the adjustment factor was last estimated, most notably with the addition of two more survey biomass estimates in 2001 and 2003 and the recalibration of the entire time series in 2003. It was also noted, by way of comparison, that the $87 \%$ adjustment factor had not been used to set the ABC for BSAI Pacific cod since the 2002 fishery (Thompson and Dorn 2003). Therefore, the 2004 assessment based its recommendation for the 2005 ABC on a new method. This method, which focused on the mean-variance tradeoff associated with future catches predicted by the standard projection model, resulted in a 2005 ABC of 58,100 t.

## Recommendation for 2006

Based on Model 3, the maximum permissible ABC (Tier 3a) for 2006 is 109,000 t. An ABC of this magnitude would represent an increase of $50,900 \mathrm{t}$, or $88 \%$, relative to the 2005 ABC. However, it should be remembered that the 2001-2003 year classes are almost certainly below average, meaning that biomass is very likely to decrease in coming years as these cohorts work their way through the age structure. For example, projections show that continued harvesting at the maximum permissible rate would be expected to result in ABCs for 2007 and 2008 of $71,400 t$ and $48,600 t$, respectively. In other words, the $50,900 \mathrm{t}$ increase in ABC between 2005 and 2006 would likely be followed by an even larger decrease $(60,400 \mathrm{t}$ ) over the next two years.

In contrast, harvesting at $50 \%$ of the maximum permissible rate would result in a 2006 ABC of $58,700 \mathrm{t}$, an increase of 600 t , or $1 \%$, relative to the 2005 ABC. Continued harvesting at $50 \%$ of the maximum permissible rate would still be expected to result in decreased ABCs during 2007 and 2008, but these decreases would likely be much smaller (both in absolute and relative terms) than they would be under the maximum permissible rate.

In conclusion, given the likelihood of decreases in future biomass and the volatility in future catches that would be expected if the stock were fished at the maximum permissible rate, it does not seem prudent to recommend harvesting at the maximum permissible rate in 2006. Harvesting at $50 \%$ of the maximum permissible rate is recommended instead, resulting in a 2006 ABC of 58,700 $t$.

## Area Allocation of Harvests

For the past several years, ABC has been allocated among regulatory areas on the basis of the three most recent surveys. The recent time series of area-specific biomass estimates are shown below, together with the proportions corresponding to a three-year weighted average (in keeping with past calculations, the 1999 estimate of biomass in the Eastern regulatory area is used as a proxy for the 2001 value, because the 2001 survey did not include the Eastern regulatory area):

| Year | Western | Central | Eastern | Total |
| :---: | ---: | ---: | ---: | ---: |
| 2001 | 133214 | 124400 | 21718 | 279332 |
| 2003 | 75632 | 207080 | 14689 | 297402 |
| 2005 | 134029 | 160118 | 13954 | 308102 |
| Average | 114292 | 163866 | 16787 | 294945 |
| Proportion | 0.39 | 0.55 | 0.06 | 1.00 |

Thus, if the previous approach for allocating ABC is retained for the 2006 fishery, the proportions would be $39 \%$ Western, $55 \%$ Central, and $7 \%$ Eastern. Relative to the 2005 values, these proportions represent an increase of $3 \%$ in the Western regulatory area, a decrease of $2 \%$ in the Central regulatory area, and a decrease of $1 \%$ in the Eastern regulatory area (Table 2.3).

## Standard Harvest and Recruitment Scenarios and Projection Methodology

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2005 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2006 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2005. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2006, are as follow (" $\max F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2006 recommended in the assessment to the max $F_{A B C}$ for 2006. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2001-2005 average $F$, which was 0.23 . (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be 1) above its MSY level in 2006 or 2) above $1 / 2$ of its MSY level in 2006 and above its MSY level in 2016 under this scenario, then the stock is not overfished.)

Scenario 7: In 2006 and 2007, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2018 under this scenario, then the stock is not approaching an overfished condition.)

## Projections and Status Determination

## Scenario Projections and Two-Year Ahead Overfishing Level

Table 2.24 defines symbols used to describe projections of spawning biomass, fishing mortality rate, and catch corresponding to the seven standard harvest scenarios. These projections are shown for Model 3 in Tables 2.25-2.30.

In addition to the seven standard harvest scenarios, Amendments 48/48 to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for $2006(130,000 \mathrm{t})$, it does not provide the best estimate of OFL for 2007, because the mean 2007 catch under Scenario 6 is predicated on the 2006 catch being equal to the 2006 OFL, whereas the actual 2006 catch will likely be less than the 2006 OFL. Therefore, the projection model was re-run with the 2006 catch fixed at the recommended 2006 ABC value of $58,700 \mathrm{t}$ and the 2007 fishing mortality rate fixed at $F_{\text {OFL }}$. The resulting estimate of the 2007 OFL was $106,000 \mathrm{t}$.

## Status Determination

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock overfished? This depends on the stock's estimated spawning biomass in 2006:
a. If spawning biomass for 2006 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2006 is estimated to be above $B_{35 \%}$ the stock is above its MSST.
c. If spawning biomass for 2006 is estimated to be above $1 / 2 B_{35 \%}$ but below B35\%, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 2.29). If the mean spawning biomass for 2016 is below $B_{35 \%}$, the stock is below its MSST. Otherwise, the stock is above its MSST.

Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7 (Table 2.30):
a. If the mean spawning biomass for 2008 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2008 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c. If the mean spawning biomass for 2008 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2018. If the mean spawning biomass for 2018 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.
In the case of BSAI Pacific cod, spawning biomass for 2006 is estimated to be above $B_{35 \%}$. Therefore, the stock is above its MSST and is not overfished. Mean spawning biomass for 2008 in Table 2.30 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, and mean spawning biomass for 2018 is above $B_{35 \%}$. Therefore, the stock is not approaching an overfished condition.

## ECOSYSTEM CONSIDERATIONS

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Boldt (ed.), 2005). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g., Hare and Mantua 2000). In the present assessment, an attempt was made to estimate the change in median recruitment of EBS Pacific cod associated with the 1977 regime shift. According to Model 3, pre-1977 median recruitment was only about $30 \%$ of post-1976 median recruitment. Establishing a link between environment and recruitment within a particular regime is more difficult. In last year’s assessment (Thompson et al. 2004), for example, the correlations between age 1 recruits spawned since 1977 and monthly values of the Pacific Decadal Oscillation (Mantua et al. 1997) were computed and found to be very weak.
The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), Lang et al. (2003), Westrheim (1996), and Yang (2004). The composition of Pacific cod prey varies to some extent by time and area. In terms of percent occurrence, some of the most important items in the diet of Pacific cod in the BSAI and GOA have been polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, some of the most important dietary items have been euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, some of the most important dietary items have been walleye pollock, fishery offal, yellowfin sole, and crustaceans. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include Pacific cod, halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

## Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Bycatch of Nontarget and "Other" Species

Bycatch of nontarget species and members of the "other species" group are shown in the following set of tables (for the 2003-2005 tables, the "hook and line" gear type includes both longline and jig gear): Tables 2.31a and 2.31b show bycatch for the GOA Pacific cod trawl fishery in 1997-2002 and 20032005, respectively. Tables 2.32 a and 2.32 b show bycatch for the GOA Pacific cod longline fishery in 1997-2002 and the GOA Pacific cod hook and line fishery in 2003-2005, respectively. Tables 2.33a and 2.33b show bycatch for the GOA Pacific cod pot fishery in 1997-2002 and 2003-2005, respectively.

It is not clear how much bycatch of a particular species constitutes "too much" in the context of ecosystem concerns. As a first step toward possible prioritization of future investigation into this question, it might be reasonable to focus on those species groups for which a Pacific cod fishery had a bycatch in excess of 100 t and accounted for more than $10 \%$ of the total bycatch in at least two of the three most recent years. This criterion results in the following list of impacted species groups (an " X " indicates that the criterion was met for that area/species/gear combination).

| Species group | Hook and Line | Pot |
| :--- | :---: | :---: |
| Large sculpins |  | X |
| Sea star | X | X |
| Skate | X |  |

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery operates to some extent in the same geographic areas used by Steller sea lion as foraging grounds (Livingston (ed.), 2002).
The Fisheries Interaction Team of the Alaska Fisheries Science Center has been engaged in research to determine the effectiveness of recent management measures designed to mitigate the impacts of the Pacific cod fisheries (among others) on Steller sea lions. Results from studies conducted in 2002-2003 were summarized by Conners et al. (2004). These studies included a tagging feasibility study, which may evolve into an ongoing research effort capable of providing information on the extent and rate to which Pacific cod move in and out of various portions of Steller sea lion critical habitat. Nearly 6,000 spaghetti tags were released, of which approximately 1,000 had been returned as of September, 2003.

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the hook and line fishery for Pacific cod (Tables 2.36b and 2.39b). Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross (Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft . LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed sets was as follows:

| Gear | BS | AI | GOA |
| :--- | ---: | ---: | ---: |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513, 517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort was dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.
Impacts of the Pacific cod fisheries on essential fish habitat were further analyzed in an environmental impact statement by NMFS (2005).

## Data Gaps and Research Priorities

Understanding of the above ecosystem considerations would be improved if future research were directed toward closing certain data gaps. Such research would have several foci, including the following: 1) ecology of the Pacific cod stock, including spatial dynamics, trophic and other interspecific relationships, and the relationship between climate and recruitment; 2) behavior of the Pacific cod fishery, including spatial dynamics; 3) determinants of trawl survey selectivity; 4) ecology of species taken as bycatch in the Pacific cod fisheries, including estimation of biomass, carrying capacity, and resilience; and 5) ecology of species that interact with Pacific cod, including estimation of biomass, carrying capacity, and resilience.

## SUMMARY

The major results of the Pacific cod stock assessment are summarized in Table 2.34.

## ACKNOWLEDGMENTS

Jim Stark provided maturity data. Eric Brown and Mark Wilkins provided survey data. Andy Smoker, Josh Keaton, Tom Pearson, Kally Spalinger, and Kim Phillips provided fishery data. Angie Greig provided fishery data and produced Figure 2.1. Nathan Mantua provided values of the Pacific Decadal Oscillation. Dave Ackley, Sarah Gaichas, and Terry Hiatt provided bycatch data. Jim Ianelli answered numerous questions about ADMB and wrote the projection model. Rick Methot and David Sampson provided advice on SS2. Anne Hollowed and the GOA Groundfish Plan Team provided reviews of this assessment.

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## Tables

Table 2.1a—Summary of catches (t) of Pacific cod by fleet sector and gear type, 1964-1990. All catches since 1980 include discards. Jt. Vent. = joint venture.

|  | Fleet Sector |  |  |  | Gear Type |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Year | Foreign | Jt. Vent. | Domestic | Trawl | Longline | Pot | Other | Total |  |  |
| 1964 | 196 | 0 | 0 | 56 | 140 | 0 | 0 | 196 |  |  |
| 1965 | 599 | 0 | 0 | 172 | 427 | 0 | 0 | 599 |  |  |
| 1966 | 1,376 | 0 | 0 | 396 | 980 | 0 | 0 | 1,376 |  |  |
| 1967 | 2,225 | 0 | 0 | 640 | 1,585 | 0 | 0 | 2,225 |  |  |
| 1968 | 1,046 | 0 | 0 | 301 | 745 | 0 | 0 | 1,046 |  |  |
| 1969 | 1,335 | 0 | 0 | 384 | 951 | 0 | 0 | 1,335 |  |  |
| 1970 | 1,805 | 0 | 0 | 519 | 1,286 | 0 | 0 | 1,805 |  |  |
| 1971 | 523 | 0 | 0 | 150 | 373 | 0 | 0 | 523 |  |  |
| 1972 | 3,513 | 0 | 0 | 1,010 | 2,503 | 0 | 0 | 3,513 |  |  |
| 1973 | 5,963 | 0 | 0 | 1,715 | 4,248 | 0 | 0 | 5,963 |  |  |
| 1974 | 5,182 | 0 | 0 | 1,491 | 3,691 | 0 | 0 | 5,182 |  |  |
| 1975 | 6,745 | 0 | 0 | 1,940 | 4,805 | 0 | 0 | 6,745 |  |  |
| 1976 | 6,764 | 0 | 0 | 1,946 | 4,818 | 0 | 0 | 6,764 |  |  |
| 1977 | 2,267 | 0 | 0 | 652 | 1,615 | 0 | 0 | 2,267 |  |  |
| 1978 | 11,370 | 7 | 813 | 4,547 | 6,800 | 0 | 843 | 12,190 |  |  |
| 1979 | 13,173 | 711 | 1,020 | 3,629 | 9,545 | 0 | 1,730 | 14,904 |  |  |
| 1980 | 34,245 | 466 | 634 | 6,464 | 27,780 | 0 | 1,101 | 35,345 |  |  |
| 1981 | 34,969 | 58 | 1,104 | 10,484 | 25,472 | 0 | 175 | 36,131 |  |  |
| 1982 | 26,937 | 193 | 2,335 | 6,679 | 22,667 | 0 | 119 | 29,465 |  |  |
| 1983 | 29,777 | 2,426 | 4,337 | 9,512 | 26,756 | 0 | 272 | 36,540 |  |  |
| 1984 | 15,896 | 4,649 | 3,353 | 8,805 | 14,844 | 0 | 249 | 23,898 |  |  |
| 1985 | 9,086 | 2,266 | 3,076 | 4,876 | 9,411 | 2 | 139 | 14,428 |  |  |
| 1986 | 15,211 | 1,357 | 8,444 | 6,850 | 17,619 | 141 | 402 | 25,012 |  |  |
| 1987 | 0 | 1,978 | 30,961 | 22,486 | 8,261 | 642 | 1,550 | 32,939 |  |  |
| 1988 | 0 | 1,661 | 32,141 | 27,145 | 3,933 | 1,422 | 1,302 | 33,802 |  |  |
| 1989 | 0 | 0 | 43,293 | 37,637 | 3,662 | 376 | 1,618 | 43,293 |  |  |
| 1990 | 0 | 0 | 72,517 | 59,188 | 5,919 | 5,661 | 1,749 | 72,517 |  |  |

Table 2.1b—Summary of catches ( t ) of Pacific cod since 1991 by management jurisdiction and gear type. Longl. = longline, Subt. = subtotal. All entries include discards. Catches for 2005 are complete through early October.

|  | Federal |  |  |  | State |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Trawl | Longl. | Pot | Other | Subt. | Pot | Other | Subt. | Total |
| 1991 | 58,093 | 7,656 | 10,464 | 115 | 76,328 | 0 | 0 | 0 | 76,328 |
| 1992 | 54,593 | 15,675 | 10,154 | 325 | 80,746 | 0 | 0 | 0 | 80,746 |
| 1993 | 37,806 | 8,962 | 9,708 | 11 | 56,487 | 0 | 0 | 0 | 56,487 |
| 1994 | 31,446 | 6,778 | 9,160 | 100 | 47,484 | 0 | 0 | 0 | 47,484 |
| 1995 | 41,875 | 10,978 | 16,055 | 77 | 68,985 | 0 | 0 | 0 | 68,985 |
| 1996 | 45,991 | 10,196 | 12,040 | 53 | 68,280 | 0 | 0 | 0 | 68,280 |
| 1997 | 48,405 | 10,977 | 9,065 | 26 | 68,474 | 7,224 | 1,319 | 8,542 | 77,017 |
| 1998 | 41,569 | 10,011 | 10,510 | 29 | 62,120 | 9,088 | 1,316 | 10,404 | 72,524 |
| 1999 | 37,167 | 12,362 | 19,015 | 70 | 68,613 | 12,075 | 1,096 | 13,171 | 81,784 |
| 2000 | 25,457 | 11,667 | 17,351 | 54 | 54,528 | 10,388 | 1,643 | 12,031 | 66,559 |
| 2001 | 24,382 | 9,913 | 7,171 | 155 | 41,621 | 7,836 | 2,084 | 9,920 | 51,541 |
| 2002 | 19,809 | 14,666 | 7,694 | 176 | 42,345 | 10,423 | 1,714 | 12,137 | 54,483 |
| 2003 | 18,799 | 9,475 | 12,675 | 88 | 41,037 | 8,031 | 3,429 | 11,461 | 52,498 |
| 2004 | 17,351 | 10,337 | 13,671 | 310 | 17,351 | 10,117 | 2,804 | 12,922 | 54,591 |
| 2005 | 14,252 | 5,089 | 10,844 | 188 | 30,372 | 9,660 | 2,818 | 12,478 | 42,850 |

Table 2.2—History of Pacific cod ABC, TAC, total catch, and type of stock assessment model used to recommend ABC. ABC was not used in management of GOA groundfish prior to 1986. Catch for 2005 is current through early October. The values in the column labeled "TAC" correspond to "optimum yield" for the years 1980-1986, "target quota" for the year 1987, and true TAC for the years 1988-2005. "SS1" refers to Stock Synthesis 1. Each cell in the "Stock Assessment Model" column lists the type of model used to recommend the ABC in the corresponding row, meaning that the model was produced in the year previous to the one listed in the corresponding row.

| Year | ABC | TAC | Catch | Stock Assessment Model (from previous year) |
| :--- | :--- | :--- | :--- | :--- |
| 1980 | n/a | 60,000 | 35,345 | n/a |
| 1981 | n/a | 70,000 | 36,131 | n/a |
| 1982 | $\mathrm{n} / \mathrm{a}$ | 60,000 | 29,465 | $\mathrm{n} / \mathrm{a}$ |
| 1983 | $\mathrm{n} / \mathrm{a}$ | 60,000 | 36,540 | $\mathrm{n} / \mathrm{a}$ |
| 1984 | $\mathrm{n} / \mathrm{a}$ | 60,000 | 23,898 | $\mathrm{n} / \mathrm{a}$ |
| 1985 | $\mathrm{n} / \mathrm{a}$ | 60,000 | 14,428 | $\mathrm{n} / \mathrm{a}$ |
| 1986 | 136,000 | 75,000 | 25,012 | survey biomass |
| 1987 | 125,000 | 50,000 | 32,939 | survey biomass |
| 1988 | 99,000 | 80,000 | 33,802 | survey biomass |
| 1989 | 71,200 | 71,200 | 43,293 | stock reduction analysis |
| 1990 | 90,000 | 90,000 | 72,517 | stock reduction analysis |
| 1991 | 77,900 | 77,900 | 76,328 | stock reduction analysis |
| 1992 | 63,500 | 63,500 | 80,746 | stock reduction analysis |
| 1993 | 56,700 | 56,700 | 56,487 | stock reduction analysis |
| 1994 | 50,400 | 50,400 | 47,484 | stock reduction analysis |
| 1995 | 69,200 | 69,200 | 68,985 | SS1 model (length-based data) |
| 1996 | 65,000 | 65,000 | 68,280 | SS1 model (length-based data) |
| 1997 | 81,500 | 69,115 | 77,017 | SS1 model (length-based data) |
| 1998 | 77,900 | 66,060 | 72,524 | SS1 model (length-based data) |
| 1999 | 84,400 | 67,835 | 81,784 | SS1 model (length-based data) |
| 2000 | 76,400 | 58,715 | 66,559 | SS1 model (length-based data) |
| 2001 | 67,800 | 52,110 | 51,541 | SS1 model (length-based data) |
| 2002 | 57,600 | 44,230 | 54,483 | SS1 model (length-based data) |
| 2003 | 52,800 | 40,540 | 52,498 | SS1 model (length-based data) |
| 2004 | 62,810 | 48,033 | 54,591 | SS1 model (length-based data) |
| 2005 | 58,100 | 44,433 | 42,850 | SS1 model (length-based data) |
|  |  |  |  |  |

Table 2.3-History of GOA Pacific cod allocations by regulatory area.

| Year(s) | Regulatory Area |  |  |
| :--- | :---: | :---: | :---: |
|  | Western | Central | Eastern |
| $1977-1985$ | 28 | 56 | 16 |
| 1986 | 40 | 44 | 16 |
| 1987 | 27 | 56 | 17 |
| $1988-1989$ | 19 | 73 | 8 |
| 1990 | 33 | 66 | 1 |
| 1991 | 33 | 62 | 5 |
| 1992 | 37 | 61 | 2 |
| $1993-1994$ | 33 | 62 | 5 |
| $1995-1996$ | 29 | 66 | 5 |
| $1997-1999$ | 35 | 63 | 2 |
| 2000-2001 | 36 | 57 | 7 |
| 2002 (ABC) | 39 | 55 | 6 |
| 2002 (TAC) | 38 | 56 | 6 |
| 2003 (ABC) | 39 | 55 | 6 |
| 2003 (TAC) | 38 | 56 | 6 |
| 2004 (ABC) | 36 | 57 | 7 |
| 2004 (TAC) | 35.3 | 56.5 | 8.2 |
| 2005 (ABC) | 36 | 57 | 7 |
| 2005 (TAC) | 35.3 | 56.5 | 8.2 |

Table 2.4a—Pacific cod discard rates by area, target species/group, and year for the period 1991-2002
(see Table 2.4b for the period 2003-2004). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

| Target species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Arrowtooth flounder |  | 0.98 | 0.59 | 0.00 | 0.10 | 0.09 | 0.00 | 1.00 | 0.63 | 0.06 |  | 0.00 |
| Atka mackerel |  |  |  | 0.81 | 1.00 | 0.00 |  |  |  |  |  |  |
| Deepwater Flat | 1.00 |  |  | 0.43 | 0.00 | 0.68 | 0.53 | 0.00 | 0.36 | 0.00 | 0.75 |  |
| Flathead sole |  |  |  | 1.00 |  | 0.07 | 0.99 | 0.00 |  | 0.29 | 0.75 | 0.00 |
| Other species | 1.00 | 0.15 | 0.63 |  | 0.10 | 0.91 | 0.00 | 0.00 | 0.96 | 0.01 | 0.00 | 0.00 |
| Pacific cod | 0.05 | 0.03 | 0.03 | 0.02 | 0.03 | 0.02 | 0.02 | 0.01 | 0.01 | 0.00 | 0.02 | 0.02 |
| Pollock | 0.82 | 0.59 | 0.15 | 0.15 | 0.95 | 0.17 | 0.98 | 0.75 | 0.89 | 0.44 | 0.00 | 1.00 |
| Rex sole |  |  |  |  | 0.16 | 0.25 | 0.61 | 0.57 |  |  |  | 1.00 |
| Rockfish | 0.15 | 0.11 | 0.13 | 0.16 | 0.11 | 0.13 | 0.14 | 0.17 | 0.17 | 0.17 | 0.00 | 0.04 |
| Sablefish | 0.84 | 0.72 | 0.72 | 0.77 | 0.55 | 0.78 | 0.54 | 0.66 | 0.52 | 0.25 | 0.27 | 0.22 |
| Shallow-water flatfish | 0.43 | 0.00 | 0.00 | 0.87 | 0.00 | 0.97 | 0.00 | 1.00 | 0.74 | 0.28 |  | 1.00 |
| Unknown | 0.01 |  |  |  |  | 1.00 | 1.00 | 1.00 |  | 1.00 |  |  |
| All targets | 0.03 | 0.03 | 0.04 | 0.02 | 0.03 | 0.02 | 0.03 | 0.01 | 0.02 | 0.00 | 0.02 | 0.02 |

Table 2.4b—Pacific cod discard rates by area, target species/group, and year for the period 2003-2004 (see Table 2.4a for the period 1991-2002; note that the IFQ halibut target does not exist in Table 2.4a). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

| Target species/group | 2003 | 2004 |
| :--- | ---: | ---: |
| Arrowtooth flounder | 0.40 | 0.27 |
| Atka mackerel |  |  |
| Deepwater flatfish | 0.01 | 0.25 |
| Flathead sole | 0.25 | 0.33 |
| IFQ halibut | 0.61 | 0.59 |
| Other species | 0.16 | 0.07 |
| Pacific cod | 0.01 | 0.01 |
| Pollock | 0.05 | 0.26 |
| Rex sole | 0.22 | 0.15 |
| Rockfish | 0.14 | 0.04 |
| Sablefish | 0.64 | 0.23 |
| Shallowwater flatfish | 0.61 | 0.53 |
| Unknown |  |  |
| All targets | 0.05 | 0.02 |

Table 2.5-Catch of Pacific cod by year, gear, and period as used in the stock assessment model. Jig catches have been merged with other gear types. Catches for period 3 in 2005 are based on 2002-2004 averages.

|  | Trawl |  |  | Longline |  |  | Pot |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Period | Period | Period | Period | Period | Period | Period | Period | Period |
| 1964 | 16 | 14 | 27 | 82 | 16 | 42 | 0 | 0 | 0 |
| 1965 | 48 | 42 | 82 | 249 | 50 | 127 | 0 | 0 | 0 |
| 1966 | 111 | 96 | 189 | 572 | 115 | 293 | 0 | 0 | 0 |
| 1967 | 180 | 155 | 305 | 926 | 186 | 473 | 0 | 0 | 0 |
| 1968 | 84 | 73 | 144 | 435 | 88 | 222 | 0 | 0 | 0 |
| 1969 | 108 | 93 | 183 | 555 | 112 | 284 | 0 | 0 | 0 |
| 1970 | 146 | 126 | 248 | 751 | 151 | 384 | 0 | 0 | 0 |
| 1971 | 42 | 36 | 72 | 218 | 44 | 111 | 0 | 0 | 0 |
| 1972 | 284 | 245 | 482 | 1461 | 294 | 747 | 0 | 0 | 0 |
| 1973 | 481 | 415 | 819 | 2480 | 499 | 1268 | 0 | 0 | 0 |
| 1974 | 418 | 361 | 711 | 2156 | 434 | 1102 | 0 | 0 | 0 |
| 1975 | 544 | 470 | 926 | 2806 | 565 | 1434 | 0 | 0 | 0 |
| 1976 | 546 | 471 | 929 | 2814 | 566 | 1438 | 0 | 0 | 0 |
| 1977 | 183 | 158 | 311 | 943 | 190 | 482 | 0 | 0 | 0 |
| 1978 | 916 | 790 | 1558 | 4720 | 950 | 2413 | 0 | 0 | 0 |
| 1979 | 1063 | 917 | 1809 | 5480 | 1103 | 2801 | 0 | 0 | 0 |
| 1980 | 2764 | 2384 | 4702 | 14245 | 2868 | 7282 | 0 | 0 | 0 |
| 1981 | 387 | 3532 | 6565 | 10504 | 5312 | 9656 | 0 | 0 | 0 |
| 1982 | 1143 | 2041 | 3495 | 9912 | 2890 | 9865 | 0 | 0 | 0 |
| 1983 | 2861 | 2844 | 3807 | 10960 | 4651 | 11145 | 0 | 0 | 0 |
| 1984 | 3429 | 2008 | 3368 | 11840 | 425 | 2579 | 0 | 0 | 0 |
| 1985 | 2427 | 571 | 1878 | 9127 | 6 | 280 | 0 | 0 | 0 |
| 1986 | 2999 | 431 | 3420 | 15927 | 460 | 1373 | 0 | 0 | 0 |
| 1987 | 5377 | 7928 | 9181 | 5343 | 983 | 1935 | 219 | 141 | 282 |
| 1988 | 16021 | 6569 | 4555 | 2979 | 507 | 447 | 1081 | 23 | 318 |
| 1989 | 24614 | 12857 | 166 | 2378 | 356 | 928 | 241 | 103 | 32 |
| 1990 | 43279 | 7514 | 8395 | 5557 | 109 | 253 | 2577 | 1008 | 2076 |
| 1991 | 55977 | 631 | 1484 | 7296 | 332 | 142 | 9591 | 0 | 873 |
| 1992 | 51911 | 1189 | 1494 | 12946 | 802 | 2251 | 9672 | 14 | 468 |
| 1993 | 33632 | 2624 | 1550 | 8485 | 307 | 181 | 9689 | 18 | 0 |
| 1994 | 29152 | 1421 | 873 | 6696 | 48 | 133 | 8742 | 0 | 418 |
| 1995 | 38476 | 802 | 2597 | 10662 | 166 | 227 | 15419 | 43 | 592 |
| 1996 | 41450 | 3048 | 1493 | 9991 | 152 | 106 | 12014 | 27 | 0 |
| 1997 | 40727 | 1638 | 6040 | 10931 | 967 | 424 | 14007 | 475 | 1807 |
| 1998 | 34690 | 3679 | 3200 | 10566 | 510 | 280 | 18479 | 0 | 1119 |
| 1999 | 30124 | 1501 | 5542 | 12782 | 555 | 191 | 25167 | 3374 | 2548 |
| 2000 | 22133 | 2574 | 750 | 12758 | 436 | 169 | 26947 | 154 | 638 |
| 2001 | 15234 | 2035 | 7113 | 11199 | 662 | 291 | 13047 | 37 | 1923 |
| 2002 | 15829 | 2705 | 1276 | 12963 | 259 | 3334 | 13602 | 83 | 4431 |
| 2003 | 10996 | 2565 | 5239 | 8416 | 407 | 768 | 20997 | 24 | 3087 |
| 2004 | 9137 | 2091 | 6339 | 8236 | 109 | 2027 | 24250 | 4 | 4461 |
| 2005 | 9561 | 1791 | 4285 | 3757 | 116 | 2043 | 22084 | 4 | 3993 |

Table 2.6-Pacific cod length sample sizes from the commercial fisheries.

|  | Trawl Fishery |  |  | Longline Fishery |  | Pot Fishery |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year. | Per. 1 | Per. | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1977 | 0 | 210 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 634 | 0 | 0 | 18670 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 14460 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 783 | 0 | 0 | 18671 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 461 | 0 | 0 | 19308 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 1390 | 0 | 0 | 22856 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 2896 | 0 | 0 | 127992 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 1039 | 0 | 0 | 47485 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 10141 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 87304 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 387 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 2432 | 0 | 0 | 0 |
| 1989 | 660 | 0 | 312 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 25396 | 10892 | 12025 | 9925 | 0 | 0 | 2783 | 2920 | 10711 |
| 1991 | 38514 | 0 | 131 | 12551 | 143 | 0 | 49453 | 139 | 0 |
| 1992 | 39683 | 0 | 2255 | 28817 | 577 | 3603 | 37177 | 664 | 5013 |
| 1993 | 26844 | 0 | 0 | 11748 | 0 | 0 | 20866 | 0 | 0 |
| 1994 | 12579 | 0 | 0 | 5201 | 0 | 0 | 16342 | 0 | 217 |
| 1995 | 26039 | 120 | 2402 | 24635 | 0 | 0 | 46625 | 0 | 1233 |
| 1996 | 17858 | 0 | 0 | 14706 | 0 | 0 | 35256 | 432 | 0 |
| 1997 | 22822 | 225 | 3746 | 7239 | 119 | 154 | 26880 | 252 | 1537 |
| 1998 | 52448 | 3465 | 6763 | 7981 | 410 | 148 | 31569 | 291 | 2902 |
| 1999 | 11550 | 232 | 1101 | 9013 | 86 | 396 | 33876 | 3719 | 3656 |
| 2000 | 6951 | 425 | 69 | 11426 | 47 | 20 | 28991 | 902 | 277 |
| 2001 | 6115 | 665 | 4560 | 12642 | 145 | 141 | 23290 | 0 | 3925 |
| 2002 | 6285 | 808 | 309 | 9583 | 134 | 3009 | 17235 | 0 | 4674 |
| 2003 | 4129 | 1187 | 1761 | 7941 | 375 | 2301 | 12019 | 9343 | 6168 |
| 2004 | 2598 | 471 | 2545 | 6647 | 337 | 2906 | 16676 | 0 | 4817 |
| 2005 | 1760 | 496 | 1065 | 5978 | 417 | 0 | 17124 | 49 | 72 |


| Yr. Per. |  | Length Bin |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| 1977 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 16 | 21 | 38 | 49 | 38 | 8 | 6 | 16 | 10 | 3 | 1 | 0 | 0 | 0 | 0 |
| 1978 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 5 | 9 | 5 | 4 | 14 | 40 | 93 | 125 | 106 | 106 | 59 | 39 | 23 | 3 | 1 | 0 | 0 | 0 |
| 1980 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 6 | 60 | 162 | 96 | 71 | 91 | 134 | 93 | 48 | 17 | 3 | 0 | 0 | 0 |
| 1981 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 29 | 85 | 148 | 145 | 47 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1982 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 26 | 39 | 118 | 255 | 280 | 294 | 174 | 111 | 52 | 14 | 15 | 5 | 2 | 1 | 0 |
| 1983 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 11 | 24 | 106 | 332 | 388 | 403 | 439 | 375 | 310 | 252 | 143 | 76 | 23 | 7 | 3 | 0 |
| 1984 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 49 | 135 | 265 | 127 | 140 | 122 | 70 | 47 | 23 | 19 | 13 | 10 | 6 | 4 | 1 |

Table 2.7b—Length frequencies of Pacific cod in the 1987-1999 trawl fishery by year, period, and length bin.

Table 2.7c—Length frequencies of Pacific cod in the post-1999 trawl fishery by year, period, and length bin.

Table 2.8a-Length frequencies of Pacific cod in the pre-1987 longline fishery by year, period, and length bin.

| Yr. Per. |  | Length Bin |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| 1978 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 38 | 91 | 276 | 1160 | 2235 | 3077 | 4051 | 3359 | 2139 | 1261 | 696 | 224 | 49 | 6 | 1 | 0 |
| 1979 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 35 | 113 | 285 | 475 | 1124 | 1327 | 1744 | 2148 | 2534 | 2258 | 1401 | 651 | 271 | 75 | 12 | 0 | 0 |
| 1980 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 43 | 256 | 1184 | 3776 | 3199 | 1989 | 1555 | 1854 | 1998 | 1630 | 787 | 276 | 99 | 19 | 2 | 1 |
| 1981 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 29 | 83 | 263 | 1558 | 4685 | 5824 | 3243 | 1485 | 844 | 570 | 379 | 199 | 101 | 28 | 8 | 0 |
| 1982 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 40 | 106 | 280 | 498 | 1945 | 3992 | 5101 | 4586 | 3115 | 1729 | 815 | 351 | 181 | 80 | 26 | 6 | 0 |
| 1983 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 24 | 164 | 728 | 2661 | 11515 | 21037 | 24663 | 22224 | 17602 | 13130 | 7842 | 3868 | 1638 | 588 | 234 | 63 | 8 |
| 1984 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 5 | 40 | 135 | 341 | 885 | 4389 | 9372 | 10579 | 7666 | 4722 | 3612 | 2572 | 1666 | 958 | 380 | 134 | 23 | 4 |
| 1985 | 3 | 0 | 0 | 0 | 0 | 1 | 0 | 8 | 45 | 114 | 206 | 316 | 440 | 1036 | 990 | 1847 | 2170 | 1294 | 626 | 462 | 294 | 186 | 89 | 14 | 3 | 0 |
| 1986 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 133 | 387 | 487 | 681 | 2963 | 6979 | 11599 | 12075 | 10988 | 13158 | 12084 | 7943 | 4112 | 2254 | 1025 | 346 | 80 |

Table 2.8b—Length frequencies of Pacific cod in the 1987-1999 longline fishery by year, period, and length bin.

Table 2.8c-Length frequencies of Pacific cod in the post-1999 longline fishery by year, period, and length bin.

Table 2.9a-Length frequencies of Pacific cod in the 1987-1999 pot fishery by year, period, and length bin.

Table 2.9b-Length frequencies of Pacific cod in the post-1999 pot fishery by year, period, and length bin.

Table 2.10—Length frequencies of Pacific cod in the trawl survey by year (all surveys take place in period 2). Numbers shown are survey estimates of population numbers at length, rescaled so that the sum equals the total size of the actual survey length sample. Note: The values for 1984-2003 have been recalibrated since last year's assessment.

| Yr. Per. |  | Length Bin |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| 1984 | 2 | 718 | 13 | 13 | 39 | 82 | 213 | 417 | 454 | 588 | 549 | 972 | 1124 | 2420 | 3577 | 2827 | 1631 | 772 | 329 | 299 | 233 | 96 | 34 | 11 | 1 | 1 |
| 1987 | 2 | 123 | 22 | 35 | 127 | 130 | 105 | 130 | 536 | 1128 | 1652 | 1823 | 1254 | 1721 | 3310 | 2947 | 2196 | 1300 | 597 | 273 | 96 | 57 | 14 | 10 | 3 | 2 |
| 1990 | 2 | 374 | 1 | 15 | 90 | 195 | 134 | 75 | 114 | 223 | 276 | 414 | 549 | 1645 | 1812 | 1928 | 1632 | 934 | 559 | 207 | 137 | 70 | 35 | 13 | 7 | 1 |
| 1993 | 2 | 46 | 20 | 307 | 372 | 233 | 284 | 191 | 261 | 445 | 513 | 800 | 1275 | 1998 | 2538 | 3079 | 2105 | 1476 | 694 | 294 | 109 | 59 | 22 | 17 | 10 | 1 |
| 1996 | 2 | 4 | 36 | 239 | 883 | 1177 | 891 | 240 | 91 | 193 | 314 | 459 | 458 | 732 | 1037 | 1378 | 1656 | 1388 | 604 | 225 | 86 | 40 | 30 | 15 | 12 | 2 |
| 1999 | 2 | 1 | 16 | 67 | 151 | 164 | 96 | 74 | 140 | 308 | 349 | 398 | 576 | 1085 | 1129 | 1432 | 1194 | 783 | 411 | 166 | 58 | 27 | 7 | 9 | 4 | 0 |
| 2001 | 2 | 6 | 57 | 105 | 192 | 233 | 320 | 229 | 187 | 183 | 310 | 455 | 437 | 749 | 753 | 726 | 764 | 534 | 302 | 133 | 52 | 19 | 15 | 5 | 5 | 0 |
| 2003 | 2 | 1 | 5 | 28 | 76 | 137 | 87 | 66 | 64 | 160 | 296 | 495 | 753 | 1280 | 1319 | 1624 | 1308 | 698 | 372 | 190 | 113 | 36 | 15 | 3 | 0 | 0 |
| 2005 | 2 | 10 | 10 | 106 | 182 | 195 | 113 | 69 | 101 | 188 | 242 | 203 | 216 | 422 | 847 | 1524 | 1257 | 485 | 214 | 103 | 72 | 45 | 88 | 98 | 45 | 7 |



Table 2.12—Biomass, standard error, 95\% confidence interval (CI), and population numbers of Pacific cod estimated by NMFS' bottom trawl survey of the GOA. All figures except population numbers are expressed in metric tons. Population numbers are expressed in terms of individual fish.

| Year | Biomass | Std. Error | Lower 95\% CI | Upper 95\% CI | Numbers |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1984 | 550,971 | 80,385 | 393,417 | 708,525 | $320,524,532$ |
| 1987 | 394,987 | 51,325 | 294,390 | 495,585 | $247,020,039$ |
| 1990 | 416,788 | 63,706 | 291,925 | 541,651 | $212,131,668$ |
| 1993 | 409,848 | 73,431 | 265,924 | 553,772 | $231,963,103$ |
| 1996 | 538,154 | 107,736 | 326,991 | 749,316 | $319,068,011$ |
| 1999 | 306,413 | 38,699 | 230,563 | 382,263 | $166,583,892$ |
| 2001 | 257,614 | 52,457 | 154,799 | 360,429 | $158,424,464$ |
| 2003 | 297,402 | 44,549 | 210,086 | 384,717 | $159,749,380$ |
| 2005 | 308,102 | 80,862 | 149,613 | 466,591 | $139,860,010$ |

Note: The 2001 survey did not cover the eastern GOA. To account for the missing stations, the 1999 survey estimates of biomass, biomass variance, and numbers for the eastern GOA were added to the respective 2001values to produce the figures shown in the above table.

Table 2.13—Pacific cod commercial fishery length sample sizes used in the multinomial distribution. (These values correspond to the square roots of the true sample sizes shown in Table 2.6.)

| Year | Trawl Fishery |  |  | Longline Fishery |  |  | Pot Fishery |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1977 | 0 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 0 | 0 | 25 | 0 | 0 | 137 | 0 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 120 | 0 | 0 | 0 |
| 1980 | 0 | 0 | 28 | 0 | 0 | 137 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 21 | 0 | 0 | 139 | 0 | 0 | 0 |
| 1982 | 0 | 0 | 37 | 0 | 0 | 151 | 0 | 0 | 0 |
| 1983 | 0 | 0 | 54 | 0 | 0 | 358 | 0 | 0 | 0 |
| 1984 | 0 | 0 | 32 | 0 | 0 | 218 | 0 | 0 | 0 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 101 | 0 | 0 | 0 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 295 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 20 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 49 | 0 | 0 | 0 |
| 1989 | 26 | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 159 | 104 | 110 | 100 | 0 | 0 | 53 | 54 | 103 |
| 1991 | 196 | 0 | 11 | 112 | 12 | 0 | 222 | 12 | 0 |
| 1992 | 199 | 0 | 47 | 170 | 24 | 60 | 193 | 26 | 71 |
| 1993 | 164 | 0 | 0 | 108 | 0 | 0 | 144 | 0 | 0 |
| 1994 | 112 | 0 | 0 | 72 | 0 | 0 | 128 | 0 | 15 |
| 1995 | 161 | 11 | 49 | 157 | 0 | 0 | 216 | 0 | 35 |
| 1996 | 134 | 0 | 0 | 121 | 0 | 0 | 188 | 21 | 0 |
| 1997 | 151 | 15 | 61 | 85 | 11 | 12 | 164 | 16 | 39 |
| 1998 | 229 | 59 | 82 | 89 | 20 | 12 | 178 | 17 | 54 |
| 1999 | 107 | 15 | 33 | 95 | 9 | 20 | 184 | 61 | 60 |
| 2000 | 83 | 21 | 8 | 107 | 7 | 4 | 170 | 30 | 17 |
| 2001 | 78 | 26 | 68 | 112 | 12 | 12 | 153 | 0 | 63 |
| 2002 | 79 | 28 | 18 | 98 | 12 | 55 | 131 | 0 | 68 |
| 2003 | 64 | 34 | 42 | 89 | 19 | 48 | 110 | 97 | 79 |
| 2004 | 51 | 22 | 50 | 82 | 18 | 54 | 129 | 0 | 69 |
| 2005 | 42 | 22 | 33 | 77 | 20 | 0 | 131 | 7 | 8 |

Table 2.14-Time series of GOA Pacific cod age 3+ biomass, spawning biomass, and survey biomass as estimated by Models 1, 2, and 3 (M1, M2, and M3). The standard deviation for each estimated spawning biomass ("SB Std. Dev.") is shown for Models 2 and 3. The time series observed by the survey itself is shown on the far right ("Obs.") for comparison to model estimates. All biomass figures are in 1000s of t .

|  | Age 3+ Biomass |  |  | Spawning Biomass |  |  | SB Std. Dev. |  | Survey Biomass |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | M1 | M2 | M3 | M1 | M2 | M3 | M2 | M3 | M1 | M2 | M3 | Obs. |
| 1964 |  | 201 | 253 |  | 82 | 102 | 12.6 | 18.9 |  |  |  |  |
| 1965 |  | 201 | 253 |  | 82 | 102 | 12.6 | 18.9 |  |  |  |  |
| 1966 |  | 201 | 253 |  | 82 | 102 | 12.6 | 18.9 |  |  |  |  |
| 1967 |  | 197 | 248 |  | 81 | 102 | 12.6 | 19.0 |  |  |  |  |
| 1968 |  | 190 | 240 |  | 79 | 100 | 13.2 | 19.6 |  |  |  |  |
| 1969 |  | 184 | 232 |  | 77 | 97 | 14.7 | 21.2 |  |  |  |  |
| 1970 |  | 176 | 222 |  | 74 | 93 | 16.5 | 23.2 |  |  |  |  |
| 1971 |  | 168 | 213 |  | 71 | 89 | 17.8 | 24.6 |  |  |  |  |
| 1972 |  | 162 | 205 |  | 68 | 86 | 18.5 | 25.4 |  |  |  |  |
| 1973 |  | 155 | 197 |  | 64 | 81 | 18.7 | 25.5 |  |  |  |  |
| 1974 |  | 149 | 191 |  | 60 | 77 | 18.6 | 25.2 |  |  |  |  |
| 1975 |  | 153 | 196 |  | 58 | 75 | 18.4 | 24.9 |  |  |  |  |
| 1976 |  | 191 | 247 |  | 60 | 78 | 18.6 | 25.4 |  |  |  |  |
| 1977 |  | 208 | 269 |  | 70 | 91 | 20.1 | 28.1 |  |  |  |  |
| 1978 | 440 | 219 | 282 | 175 | 84 | 109 | 22.3 | 32.0 |  |  |  |  |
| 1979 | 499 | 214 | 276 | 183 | 88 | 114 | 23.3 | 33.6 |  |  |  |  |
| 1980 | 600 | 307 | 401 | 197 | 90 | 118 | 23.4 | 34.0 |  |  |  |  |
| 1981 | 632 | 339 | 452 | 219 | 102 | 137 | 25.8 | 38.6 |  |  |  |  |
| 1982 | 674 | 364 | 492 | 244 | 125 | 169 | 30.0 | 46.2 |  |  |  |  |
| 1983 | 719 | 431 | 588 | 262 | 144 | 197 | 33.3 | 52.0 |  |  |  |  |
| 1984 | 725 | 454 | 624 | 274 | 161 | 223 | 36.3 | 57.4 | 427 | 358 | 372 | 551 |
| 1985 | 726 | 467 | 643 | 285 | 180 | 249 | 39.2 | 62.2 |  |  |  |  |
| 1986 | 733 | 475 | 651 | 288 | 192 | 264 | 40.7 | 64.2 |  |  |  |  |
| 1987 | 771 | 540 | 732 | 287 | 194 | 267 | 40.7 | 63.9 | 408 | 399 | 401 | 395 |
| 1988 | 774 | 564 | 759 | 288 | 202 | 277 | 40.9 | 64.6 |  |  |  |  |
| 1989 | 779 | 561 | 750 | 294 | 217 | 294 | 41.7 | 66.1 |  |  |  |  |
| 1990 | 795 | 591 | 785 | 290 | 221 | 298 | 41.4 | 65.7 | 432 | 423 | 421 | 417 |
| 1991 | 777 | 561 | 750 | 277 | 213 | 289 | 40.7 | 64.8 |  |  |  |  |
| 1992 | 781 | 579 | 776 | 271 | 206 | 281 | 40.0 | 64.1 |  |  |  |  |
| 1993 | 773 | 577 | 777 | 273 | 203 | 280 | 39.9 | 64.3 | 439 | 422 | 426 | 410 |
| 1994 | 770 | 573 | 768 | 283 | 215 | 294 | 40.6 | 65.5 |  |  |  |  |
| 1995 | 764 | 575 | 764 | 285 | 223 | 301 | 40.6 | 65.4 |  |  |  |  |
| 1996 | 731 | 548 | 729 | 273 | 214 | 289 | 39.3 | 63.4 | 397 | 387 | 385 | 538 |
| 1997 | 708 | 511 | 683 | 258 | 201 | 273 | 37.8 | 60.9 |  |  |  |  |
| 1998 | 682 | 509 | 683 | 243 | 184 | 252 | 36.3 | 58.5 |  |  |  |  |
| 1999 | 641 | 476 | 643 | 232 | 176 | 243 | 35.9 | 57.7 | 346 | 328 | 335 | 306 |
| 2000 | 590 | 432 | 591 | 217 | 165 | 231 | 35.6 | 57.0 |  |  |  |  |
| 2001 | 568 | 411 | 564 | 207 | 156 | 219 | 34.4 | 55.1 | 313 | 303 | 310 | 258 |
| 2002 | 575 | 421 | 575 | 201 | 151 | 212 | 33.3 | 53.2 |  |  |  |  |
| 2003 | 574 | 428 | 584 | 201 | 150 | 210 | 33.3 | 53.0 | 315 | 305 | 312 | 297 |
| 2004 | 535 | 406 | 554 | 201 | 153 | 214 | 34.9 | 54.7 |  |  |  |  |
| 2005 | 478 | 368 | 505 | 190 | 148 | 207 | 36.7 | 56.2 | 271 | 270 | 275 | 308 |

Table 2.15-Estimates of Pacific cod fishing mortality rates, expressed on an annual time scale. Empty cells indicate that no catch was recorded.

| Year | Trawl |  |  | Longline |  |  | Pot |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |
| 1964 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |  |  |  |
| 1965 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.001 |  |  |  |
| 1966 | 0.001 | 0.000 | 0.001 | 0.003 | 0.001 | 0.002 |  |  |  |
| 1967 | 0.001 | 0.001 | 0.002 | 0.006 | 0.001 | 0.003 |  |  |  |
| 1968 | 0.001 | 0.000 | 0.001 | 0.003 | 0.001 | 0.001 |  |  |  |
| 1969 | 0.001 | 0.000 | 0.001 | 0.004 | 0.001 | 0.002 |  |  |  |
| 1970 | 0.001 | 0.001 | 0.001 | 0.005 | 0.001 | 0.003 |  |  |  |
| 1971 | 0.000 | 0.000 | 0.000 | 0.002 | 0.000 | 0.001 |  |  |  |
| 1972 | 0.002 | 0.001 | 0.003 | 0.011 | 0.002 | 0.006 |  |  |  |
| 1973 | 0.004 | 0.003 | 0.005 | 0.019 | 0.004 | 0.010 |  |  |  |
| 1974 | 0.004 | 0.002 | 0.005 | 0.018 | 0.004 | 0.009 |  |  |  |
| 1975 | 0.005 | 0.003 | 0.006 | 0.023 | 0.005 | 0.012 |  |  |  |
| 1976 | 0.005 | 0.003 | 0.005 | 0.022 | 0.004 | 0.010 |  |  |  |
| 1977 | 0.001 | 0.001 | 0.001 | 0.006 | 0.001 | 0.003 |  |  |  |
| 1978 | 0.006 | 0.003 | 0.007 | 0.024 | 0.005 | 0.012 |  |  |  |
| 1979 | 0.006 | 0.004 | 0.008 | 0.027 | 0.005 | 0.014 |  |  |  |
| 1980 | 0.016 | 0.010 | 0.017 | 0.069 | 0.014 | 0.033 |  |  |  |
| 1981 | 0.002 | 0.011 | 0.019 | 0.043 | 0.020 | 0.033 |  |  |  |
| 1982 | 0.005 | 0.005 | 0.009 | 0.031 | 0.009 | 0.028 |  |  |  |
| 1983 | 0.010 | 0.007 | 0.008 | 0.030 | 0.012 | 0.028 |  |  |  |
| 1984 | 0.010 | 0.004 | 0.007 | 0.029 | 0.001 | 0.006 |  |  |  |
| 1985 | 0.006 | 0.001 | 0.004 | 0.020 | 0.000 | 0.001 |  |  |  |
| 1986 | 0.007 | 0.001 | 0.006 | 0.033 | 0.001 | 0.003 |  |  |  |
| 1987 | 0.014 | 0.022 | 0.025 | 0.014 | 0.003 | 0.005 | 0.001 | 0.001 | 0.001 |
| 1988 | 0.041 | 0.018 | 0.012 | 0.008 | 0.001 | 0.001 | 0.004 | 0.000 | 0.001 |
| 1989 | 0.059 | 0.033 | 0.000 | 0.006 | 0.001 | 0.002 | 0.001 | 0.000 | 0.000 |
| 1990 | 0.100 | 0.019 | 0.022 | 0.014 | 0.000 | 0.001 | 0.008 | 0.003 | 0.007 |
| 1991 | 0.134 | 0.002 | 0.004 | 0.019 | 0.001 | 0.000 | 0.032 |  | 0.003 |
| 1992 | 0.129 | 0.003 | 0.004 | 0.034 | 0.002 | 0.006 | 0.033 | 0.000 | 0.002 |
| 1993 | 0.084 | 0.007 | 0.004 | 0.023 | 0.001 | 0.000 | 0.034 | 0.000 |  |
| 1994 | 0.069 | 0.004 | 0.002 | 0.017 | 0.000 | 0.000 | 0.028 |  | 0.001 |
| 1995 | 0.087 | 0.002 | 0.007 | 0.025 | 0.000 | 0.001 | 0.047 | 0.000 | 0.002 |
| 1996 | 0.097 | 0.008 | 0.004 | 0.025 | 0.000 | 0.000 | 0.038 | 0.000 |  |
| 1997 | 0.102 | 0.005 | 0.018 | 0.029 | 0.003 | 0.001 | 0.049 | 0.002 | 0.007 |
| 1998 | 0.096 | 0.012 | 0.010 | 0.031 | 0.002 | 0.001 | 0.072 |  | 0.005 |
| 1999 | 0.088 | 0.005 | 0.018 | 0.040 | 0.002 | 0.001 | 0.105 | 0.015 | 0.011 |
| 2000 | 0.065 | 0.009 | 0.002 | 0.044 | 0.002 | 0.001 | 0.099 | 0.001 | 0.002 |
| 2001 | 0.047 | 0.007 | 0.024 | 0.040 | 0.003 | 0.001 | 0.049 | 0.000 | 0.008 |
| 2002 | 0.051 | 0.010 | 0.004 | 0.049 | 0.001 | 0.013 | 0.055 | 0.000 | 0.019 |
| 2003 | 0.036 | 0.009 | 0.018 | 0.033 | 0.002 | 0.003 | 0.088 | 0.000 | 0.013 |
| 2004 | 0.029 | 0.007 | 0.021 | 0.031 | 0.000 | 0.008 | 0.097 | 0.000 | 0.018 |
| 2005 | 0.031 | 0.006 | 0.015 | 0.014 | 0.000 | 0.008 | 0.087 | 0.000 | 0.017 |

Table 2.16—Estimates of Pacific cod regime-specific median recruitments and recruitment deviations. Deviations are expressed as the difference between the logarithm of annual recruitment at age 0 and the logarithm of median recruitment for the respective environmental regime.

| Year | $\ln$ (Median Recruitment) | Annual Deviation |
| :--- | ---: | ---: |
| 1964 | 11.681 | -0.139 |
| 1965 | 11.681 | -0.172 |
| 1966 | 11.681 | -0.208 |
| 1967 | 11.681 | -0.242 |
| 1968 | 11.681 | -0.264 |
| 1969 | 11.681 | -0.255 |
| 1970 | 11.681 | -0.185 |
| 1971 | 11.681 | -0.010 |
| 1972 | 11.681 | 0.333 |
| 1973 | 11.681 | 1.351 |
| 1974 | 11.681 | 0.080 |
| 1975 | 11.681 | -0.174 |
| 1976 | 11.681 | -0.123 |
| 1977 | 12.891 | 0.964 |
| 1978 | 12.891 | -0.177 |
| 1979 | 12.891 | -0.138 |
| 1980 | 12.891 | 0.677 |
| 1981 | 12.891 | -0.152 |
| 1982 | 12.891 | -0.322 |
| 1983 | 12.891 | -0.305 |
| 1984 | 12.891 | 0.870 |
| 1985 | 12.891 | 0.085 |
| 1986 | 12.891 | -0.373 |
| 1987 | 12.891 | 0.623 |
| 1988 | 12.891 | -0.161 |
| 1989 | 12.891 | 0.752 |
| 1990 | 12.891 | 0.307 |
| 1991 | 12.891 | -0.149 |
| 1992 | 12.891 | 0.090 |
| 1993 | 12.891 | -0.002 |
| 1994 | 12.891 | -0.156 |
| 1995 | 12.891 | 0.605 |
| 1996 | 12.891 | -0.453 |
| 1997 | 12.891 | -0.284 |
| 1998 | 12.891 | -0.007 |
| 1999 | 12.891 | 0.310 |
| 2000 | 12.891 | 0.166 |
| 2001 | 12.891 | -0.844 |
| 2002 | 12.891 | -0.833 |
| 2003 | 12.891 | -0.912 |
| 2004 | 12.891 | -0.174 |
|  |  |  |
|  |  |  |

Table 2.17-Estimates of Pacific cod selectivity parameters. The first column lists the eight parameters of the selectivity function: the size at which selectivity first reaches a value of 1 ("peak location"), selectivity at the minimum length represented in the data ("S(Lmin)"), the logit transform of the size corresponding to the inflection of the ascending logistic curve ("logit(infl1)"), the relative slope of the ascending logistic curve ("slope1"), the logit transform of the size corresponding to the inflection of the descending logistic curve ("logit(infl2)"), the relative slope of the descending logistic curve ("slope2"), the logit transform of selectivity at the maximum length represented in the data ("logit(S(Lmax))"), and the width of the length range at which selectivity equals 1 ("peak width"). The middle portion of the table lists the portion of the time series ("era") to which each parameter value applies (FOR = pre-1987, DOM = 1987-1999, NEW = post-1999), for each of the four fisheries (TWL1 = January-May Trawl, TWL2 = June-December Trawl, LGL = longline, POT = pot). For the January-May trawl fishery, parameters for the pre-1987 selectivity curve were borrowed from the 1987-1999 estimates, because there were very few size composition data from the pre-1988 fishery. For the pot fishery, there was no catch prior to 1987 , so no selectivity parameters are needed for that fishery during the pre-1987 portion of the time series. The right-most column lists the values of the selectivity parameters for the bottom trawl survey. Because the survey selectivity is assumed to be constant over the entire time series, the value of each survey selectivity parameter is shown in triplicate (once for each portion of the time series).

|  |  | Fishery Selectivity |  |  |  |  |
| :--- | :---: | ---: | ---: | ---: | ---: | ---: |
| Parameter | Era | TWL1 | TWL2 | LGL | POT | Survey |
| peak location | FOR | 66.119 | 58.624 | 66.584 |  | 57.622 |
| peak location | DOM | 66.119 | 65.271 | 65.930 | 65.804 | 57.622 |
| peak location | NEW | 64.941 | 65.630 | 65.089 | 64.077 | 57.622 |
| S(Lmin) | FOR | 0.001 | 0.001 | 0.001 |  | 0.009 |
| S(Lmin) | DOM | 0.001 | 0.001 | 0.001 | 0.001 | 0.009 |
| S(Lmin) | NEW | 0.001 | 0.001 | 0.001 | 0.001 | 0.009 |
| logit(infl1) | FOR | 1.920 | 1.486 | 1.078 |  | 0.902 |
| logit(infl1) | DOM | 1.920 | 1.602 | 1.882 | 2.267 | 0.902 |
| logit(infl1) | NEW | 1.586 | 1.501 | 1.978 | 2.079 | 0.902 |
| slope1 | FOR | 0.089 | 0.242 | 0.295 |  | 0.053 |
| slope1 | DOM | 0.089 | 0.116 | 0.180 | 0.185 | 0.053 |
| slope1 | NEW | 0.137 | 0.135 | 0.206 | 0.232 | 0.053 |
| logit(S(Lmax)) | FOR | -0.633 | 1.943 | -0.802 |  | 0.000 |
| logit(S(Lmax)) | DOM | -0.633 | -1.373 | -1.007 | -2.308 | 0.000 |
| logit(S(Lmax)) | NEW | -0.744 | -1.049 | -1.390 | -1.723 | 0.000 |
| logit(infl2) | FOR | -0.301 | 0.036 | -0.104 |  | -1.302 |
| logit(infl2) | DOM | -0.301 | -1.228 | -0.661 | -1.353 | -1.302 |
| logit(infl2) | NEW | -1.062 | -1.239 | -0.996 | -1.370 | -1.302 |
| slope2 | FOR | 0.161 | 0.203 | 0.221 |  | 0.202 |
| slope2 | DOM | 0.161 | 0.218 | 0.184 | 0.195 | 0.202 |
| slope2 | NEW | 0.210 | 0.217 | 0.207 | 0.193 | 0.202 |
| peak width | FOR | 8.842 | 10.054 | 10.916 |  | 9.879 |
| peak width | DOM | 8.842 | 8.343 | 8.863 | 7.924 | 9.879 |
| peak width | NEW | 8.732 | 8.288 | 8.748 | 8.374 | 9.879 |

Table 2.18—Selectivities of Pacific cod selectivities at length in the commercial fisheries and bottom trawl survey as defined by final parameter estimates. Lengths (cm) correspond to mid-points of size bins. Len. = length, FOR = 1964-1986, DOM = 1987-1999, NEW = 2000-2005.

| Len. | Jan-May Trawl |  | Jul-Dec Trawl Fishery |  |  | Longline Fishery |  |  | Pot Fishery |  | Sur. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | DOM | NEW | FOR | DOM | NEW | FOR | DOM | NEW | DOM | NEW |  |
| 10.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 |
| 13.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 |
| 16.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 |
| 19.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.05 |
| 22.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.07 |
| 25.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.10 |
| 28.5 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.15 |
| 31.5 | 0.00 | 0.01 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.20 |
| 34.5 | 0.01 | 0.02 | 0.02 | 0.02 | 0.02 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.26 |
| 37.5 | 0.02 | 0.03 | 0.04 | 0.04 | 0.04 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.33 |
| 40.5 | 0.03 | 0.05 | 0.08 | 0.06 | 0.06 | 0.03 | 0.01 | 0.01 | 0.00 | 0.00 | 0.41 |
| 43.5 | 0.05 | 0.09 | 0.16 | 0.10 | 0.10 | 0.07 | 0.02 | 0.02 | 0.01 | 0.01 | 0.50 |
| 47.5 | 0.10 | 0.18 | 0.37 | 0.18 | 0.19 | 0.21 | 0.06 | 0.05 | 0.02 | 0.04 | 0.63 |
| 52.5 | 0.22 | 0.36 | 0.71 | 0.35 | 0.37 | 0.54 | 0.19 | 0.16 | 0.08 | 0.15 | 0.81 |
| 57.5 | 0.43 | 0.61 | 0.97 | 0.59 | 0.61 | 0.85 | 0.44 | 0.44 | 0.28 | 0.46 | 1.00 |
| 62.5 | 0.73 | 0.88 | 1.00 | 0.86 | 0.86 | 0.97 | 0.79 | 0.83 | 0.68 | 0.89 | 1.00 |
| 67.5 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 72.5 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.99 | 0.85 |
| 77.5 | 0.96 | 0.85 | 0.99 | 0.79 | 0.82 | 1.00 | 0.92 | 0.85 | 0.72 | 0.68 | 0.71 |
| 82.5 | 0.85 | 0.66 | 0.97 | 0.54 | 0.58 | 0.94 | 0.76 | 0.62 | 0.45 | 0.44 | 0.60 |
| 87.5 | 0.71 | 0.49 | 0.94 | 0.36 | 0.41 | 0.82 | 0.58 | 0.41 | 0.26 | 0.29 | 0.54 |
| 92.5 | 0.57 | 0.39 | 0.91 | 0.26 | 0.31 | 0.62 | 0.42 | 0.29 | 0.16 | 0.21 | 0.52 |
| 97.5 | 0.45 | 0.34 | 0.89 | 0.22 | 0.28 | 0.45 | 0.33 | 0.23 | 0.12 | 0.17 | 0.51 |
| 102.5 | 0.38 | 0.33 | 0.88 | 0.21 | 0.26 | 0.35 | 0.29 | 0.21 | 0.10 | 0.16 | 0.50 |
| 107.5 | 0.35 | 0.32 | 0.87 | 0.20 | 0.26 | 0.31 | 0.27 | 0.20 | 0.09 | 0.15 | 0.50 |

Table 2.19—Schedules of Pacific cod selectivities at age for the most recent portion of the time series as implied by final parameter estimates. Per. 1 = January-May, Per. 2 = June-August, Per. 3 = SeptemberDecember. Because selectivity is defined as a function of length, not age, profiles of selectivity at age do not necessarily reach a peak value of 1 .

| Age | Trawl Fishery |  |  | Longline Fishery |  |  | Pot Fishery |  |  | Survey |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 | Per. 1 | Per. 2 | Per. 3 |  |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 |
| 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.06 |
| 2 | 0.01 | 0.02 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.21 |
| 3 | 0.05 | 0.10 | 0.15 | 0.01 | 0.03 | 0.05 | 0.01 | 0.02 | 0.05 | 0.44 |
| 4 | 0.22 | 0.31 | 0.39 | 0.11 | 0.18 | 0.26 | 0.11 | 0.18 | 0.26 | 0.69 |
| 5 | 0.48 | 0.57 | 0.63 | 0.36 | 0.46 | 0.53 | 0.37 | 0.47 | 0.54 | 0.85 |
| 6 | 0.70 | 0.73 | 0.76 | 0.61 | 0.67 | 0.70 | 0.62 | 0.66 | 0.69 | 0.89 |
| 7 | 0.79 | 0.77 | 0.77 | 0.73 | 0.74 | 0.74 | 0.71 | 0.71 | 0.70 | 0.84 |
| 8 | 0.78 | 0.73 | 0.71 | 0.73 | 0.72 | 0.70 | 0.68 | 0.66 | 0.63 | 0.78 |
| 9 | 0.72 | 0.66 | 0.64 | 0.68 | 0.65 | 0.63 | 0.60 | 0.57 | 0.55 | 0.72 |
| 10 | 0.66 | 0.59 | 0.57 | 0.60 | 0.58 | 0.56 | 0.52 | 0.49 | 0.47 | 0.67 |
| 11 | 0.60 | 0.53 | 0.51 | 0.53 | 0.51 | 0.50 | 0.45 | 0.43 | 0.41 | 0.63 |
| 12 | 0.55 | 0.48 | 0.47 | 0.48 | 0.46 | 0.45 | 0.39 | 0.37 | 0.36 | 0.60 |
| 13 | 0.51 | 0.44 | 0.43 | 0.43 | 0.42 | 0.40 | 0.35 | 0.33 | 0.32 | 0.58 |
| 14 | 0.48 | 0.41 | 0.40 | 0.39 | 0.38 | 0.37 | 0.31 | 0.30 | 0.30 | 0.57 |
| 15 | 0.45 | 0.39 | 0.38 | 0.36 | 0.35 | 0.35 | 0.29 | 0.28 | 0.27 | 0.55 |
| 16 | 0.44 | 0.37 | 0.36 | 0.34 | 0.33 | 0.33 | 0.27 | 0.26 | 0.26 | 0.55 |
| 17 | 0.42 | 0.35 | 0.35 | 0.32 | 0.32 | 0.31 | 0.25 | 0.25 | 0.24 | 0.54 |
| 18 | 0.41 | 0.34 | 0.34 | 0.31 | 0.30 | 0.30 | 0.24 | 0.24 | 0.23 | 0.53 |
| 19 | 0.40 | 0.33 | 0.33 | 0.30 | 0.29 | 0.29 | 0.23 | 0.23 | 0.22 | 0.53 |
| 20 | 0.39 | 0.33 | 0.32 | 0.29 | 0.28 | 0.28 | 0.22 | 0.22 | 0.22 | 0.53 |

Table 2.20—Schedules of Pacific cod length (cm), proportion mature, and weight (kg) at period and age as estimated by Model 3. Pop. = population, Per. 1 = Jan-Jun, Per. 2 = Jul-Aug, Per. 3 = Sep-Dec, Beg. $=$ beginning of period, Mid. = middle of period, SDev. = standard deviation, Mat. = proportion mature, Twl. = trawl fishery, Lgl. = longline fishery, pot = pot fishery, survey = bottom trawl survey.

| Per. | Age | Length |  |  | Mat. | Pop. Weight |  | Fishery/Survey Weight |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Beg. | Mid. | S.Dev. |  | Beg. | Mid. | Twl. | Lgl. | Pot | Survey |
| 1 | 0 | 2.224 | 4.975 | 1.897 | 0.000 | 0.010 | 0.010 | 0.010 | 0.010 | 0.010 | 0.010 |
| 1 | 1 | 14.839 | 17.286 | 3.499 | 0.000 | 0.032 | 0.051 | 0.056 | 0.051 | 0.051 | 0.062 |
| 1 | 2 | 26.056 | 28.232 | 4.964 | 0.007 | 0.185 | 0.239 | 0.331 | 0.280 | 0.264 | 0.286 |
| 1 | 3 | 36.032 | 37.966 | 6.335 | 0.091 | 0.517 | 0.611 | 0.919 | 1.089 | 1.140 | 0.728 |
| 1 | 4 | 44.902 | 46.622 | 7.588 | 0.367 | 1.041 | 1.171 | 1.657 | 1.939 | 1.995 | 1.352 |
| 1 | 5 | 52.790 | 54.320 | 8.728 | 0.654 | 1.731 | 1.893 | 2.419 | 2.658 | 2.664 | 2.056 |
| 1 | 6 | 59.804 | 61.165 | 9.762 | 0.818 | 2.566 | 2.755 | 3.149 | 3.314 | 3.257 | 2.782 |
| 1 | 7 | 66.041 | 67.251 | 10.698 | 0.894 | 3.513 | 3.720 | 3.842 | 3.932 | 3.801 | 3.537 |
| 1 | 8 | 71.588 | 72.664 | 11.543 | 0.931 | 4.535 | 4.753 | 4.518 | 4.522 | 4.321 | 4.350 |
| 1 | 9 | 76.520 | 77.477 | 12.305 | 0.950 | 5.595 | 5.816 | 5.206 | 5.101 | 4.844 | 5.237 |
| 1 | 10 | 80.906 | 81.757 | 12.991 | 0.962 | 6.651 | 6.865 | 5.920 | 5.687 | 5.392 | 6.175 |
| 1 | 11 | 84.806 | 85.562 | 13.607 | 0.971 | 7.655 | 7.853 | 6.654 | 6.286 | 5.970 | 7.117 |
| 1 | 12 | 88.274 | 88.947 | 14.160 | 0.977 | 8.568 | 8.745 | 7.384 | 6.891 | 6.568 | 8.011 |
| 1 | 13 | 91.358 | 91.956 | 14.592 | 0.982 | 9.371 | 9.524 | 8.088 | 7.492 | 7.174 | 8.826 |
| 1 | 14 | 94.100 | 94.632 | 14.977 | 0.986 | 10.061 | 10.191 | 8.742 | 8.069 | 7.764 | 9.543 |
| 2 | 0 | 7.660 | 9.240 | 1.897 | n/a | 0.010 | 0.010 | 0.010 | 0.010 | 0.010 | 0.010 |
| 2 | 1 | 19.673 | 21.078 | 3.499 | n/a | 0.032 | 0.051 | 0.116 | 0.096 | 0.096 | 0.115 |
| 2 | 2 | 30.355 | 31.605 | 4.964 | n/a | 0.185 | 0.239 | 0.484 | 0.482 | 0.459 | 0.407 |
| 2 | 3 | 39.855 | 40.965 | 6.335 | n/a | 0.517 | 0.611 | 1.129 | 1.382 | 1.449 | 0.919 |
| 2 | 4 | 48.301 | 49.289 | 7.588 | n/a | 1.041 | 1.171 | 1.885 | 2.185 | 2.226 | 1.581 |
| 2 | 5 | 55.813 | 56.691 | 8.728 | n/a | 1.731 | 1.893 | 2.647 | 2.877 | 2.864 | 2.294 |
| 2 | 6 | 62.492 | 63.273 | 9.762 | n/a | 2.566 | 2.755 | 3.356 | 3.521 | 3.441 | 3.028 |
| 2 | 7 | 68.432 | 69.126 | 10.698 | n/a | 3.513 | 3.720 | 4.010 | 4.130 | 3.975 | 3.799 |
| 2 | 8 | 73.713 | 74.331 | 11.543 | n/a | 4.535 | 4.753 | 4.641 | 4.714 | 4.493 | 4.637 |
| 2 | 9 | 78.410 | 78.959 | 12.305 | n/a | 5.595 | 5.816 | 5.285 | 5.294 | 5.024 | 5.546 |
| 2 | 10 | 82.587 | 83.075 | 12.991 | n/a | 6.651 | 6.865 | 5.958 | 5.886 | 5.582 | 6.492 |
| 2 | 11 | 86.301 | 86.735 | 13.607 | n/a | 7.655 | 7.853 | 6.653 | 6.489 | 6.170 | 7.423 |
| 2 | 12 | 89.603 | 89.989 | 14.160 | n/a | 8.568 | 8.745 | 7.347 | 7.093 | 6.771 | 8.293 |
| 2 | 13 | 92.540 | 92.883 | 14.592 | n/a | 9.371 | 9.524 | 8.020 | 7.688 | 7.373 | 9.076 |
| 2 | 14 | 95.151 | 95.457 | 14.977 | n/a | 10.061 | 10.191 | 8.649 | 8.254 | 7.954 | 9.760 |
| 3 | 0 | 10.797 | 12.838 | 1.897 | n/a | 0.010 | 0.010 | 0.021 | 0.020 | 0.020 | 0.024 |
| 3 | 1 | 22.463 | 24.277 | 3.499 | n/a | 0.032 | 0.051 | 0.192 | 0.155 | 0.151 | 0.179 |
| 3 | 2 | 32.836 | 34.449 | 4.964 | n/a | 0.185 | 0.239 | 0.650 | 0.731 | 0.736 | 0.533 |
| 3 | 3 | 42.060 | 43.495 | 6.335 | n/a | 0.517 | 0.611 | 1.339 | 1.627 | 1.694 | 1.100 |
| 3 | 4 | 50.263 | 51.539 | 7.588 | n/a | 1.041 | 1.171 | 2.107 | 2.391 | 2.417 | 1.784 |
| 3 | 5 | 57.557 | 58.691 | 8.728 | n/a | 1.731 | 1.893 | 2.858 | 3.066 | 3.035 | 2.505 |
| 3 | 6 | 64.043 | 65.052 | 9.762 | n/a | 2.566 | 2.755 | 3.550 | 3.700 | 3.598 | 3.247 |
| 3 | 7 | 69.811 | 70.708 | 10.698 | n/a | 3.513 | 3.720 | 4.194 | 4.301 | 4.125 | 4.035 |
| 3 | 8 | 74.940 | 75.738 | 11.543 | n/a | 4.535 | 4.753 | 4.826 | 4.882 | 4.645 | 4.895 |
| 3 | 9 | 79.501 | 80.210 | 12.305 | n/a | 5.595 | 5.816 | 5.478 | 5.465 | 5.184 | 5.819 |
| 3 | 10 | 83.556 | 84.187 | 12.991 | n/a | 6.651 | 6.865 | 6.159 | 6.061 | 5.752 | 6.767 |
| 3 | 11 | 87.163 | 87.724 | 13.607 | n/a | 7.655 | 7.853 | 6.857 | 6.667 | 6.346 | 7.685 |
| 3 | 12 | 90.370 | 90.869 | 14.160 | n/a | 8.568 | 8.745 | 7.547 | 7.269 | 6.948 | 8.531 |
| 3 | 13 | 93.222 | 93.665 | 14.592 | n/a | 9.371 | 9.524 | 8.209 | 7.856 | 7.546 | 9.286 |
| 3 | 14 | 95.758 | 96.152 | 14.977 | n/a | 10.061 | 10.191 | 8.822 | 8.412 | 8.117 | 9.940 |

Table 2.21—Time series of EBS Pacific cod age 3+ biomass and spawning biomass for the years 19772005 as estimated in last year's and this year's assessments, with $95 \%$ confidence intervals ("SB 95\% CI") for spawning biomass as estimated in this year's assessment. Note that last year's model used 1978 as the initial year, so age 3+ and spawning biomass for 1977 are not available from that assessment. Biomass values are in 1000s of $t$.

| Year | Age 3+ Biomass |  | Spawning Biomass |  | SB 95\% CI (This Year) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Last Year | This Year | Last Year | This Year | Lower | Upper |
| 1977 | n/a | 269 | n/a | 91 | 64 | 119 |
| 1978 | 494 | 282 | 85 | 109 | 77 | 140 |
| 1979 | 560 | 276 | 100 | 114 | 81 | 147 |
| 1980 | 630 | 401 | 110 | 118 | 85 | 151 |
| 1981 | 677 | 452 | 113 | 137 | 99 | 175 |
| 1982 | 713 | 492 | 123 | 169 | 124 | 215 |
| 1983 | 742 | 588 | 136 | 197 | 147 | 248 |
| 1984 | 747 | 624 | 147 | 223 | 167 | 279 |
| 1985 | 759 | 643 | 159 | 249 | 188 | 310 |
| 1986 | 783 | 651 | 167 | 264 | 201 | 327 |
| 1987 | 799 | 732 | 171 | 267 | 204 | 329 |
| 1988 | 808 | 759 | 168 | 277 | 214 | 341 |
| 1989 | 820 | 750 | 167 | 294 | 230 | 359 |
| 1990 | 828 | 785 | 163 | 298 | 234 | 363 |
| 1991 | 817 | 750 | 152 | 289 | 226 | 353 |
| 1992 | 807 | 776 | 144 | 281 | 218 | 344 |
| 1993 | 789 | 777 | 141 | 280 | 217 | 343 |
| 1994 | 782 | 768 | 145 | 294 | 230 | 358 |
| 1995 | 773 | 764 | 149 | 301 | 237 | 365 |
| 1996 | 741 | 729 | 146 | 289 | 227 | 351 |
| 1997 | 710 | 683 | 140 | 273 | 213 | 332 |
| 1998 | 667 | 683 | 129 | 252 | 195 | 310 |
| 1999 | 624 | 643 | 118 | 243 | 186 | 299 |
| 2000 | 576 | 591 | 108 | 231 | 175 | 287 |
| 2001 | 553 | 564 | 104 | 219 | 165 | 273 |
| 2002 | 548 | 575 | 100 | 212 | 160 | 264 |
| 2003 | 530 | 584 | 95 | 210 | 158 | 262 |
| 2004 | 501 | 554 | 93 | 214 | 160 | 267 |
| 2005 | n/a | 505 | n/a | 207 | 152 | 262 |

Table 2.22—Time series of GOA Pacific cod age 0 recruitment (1000s of fish) as estimated in last year's and this year's assessments, 1977-2004. Because last year's assessment used 1 as the initial age in the model, age 0 recruitments for last year’s assessment were inferred here by multiplying last year’s estimates of age 1 recruits by $\exp (0.37)$, where 0.37 is the value of the natural mortality rate used in last year's assessment. The columns labeled "L95\%CI" and "U95\%CI" under this year's assessment represent the lower and upper bounds of the $95 \%$ confidence interval for each cohort. Bold font indicates a value from this year's assessment in excess of the 1977-2004 average of 401,982,571 fish.

| Year | Last Year's Assess. <br> Recruits | This Year's Assessment |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Recruits | L95\%CI | U95\%CI |
| 1977 | 693465 | 932492 | 1342981 | 647471 |
| 1978 | 298233 | 298002 | 583083 | 152303 |
| 1979 | 480648 | 309685 | 578388 | 165814 |
| 1980 | 406813 | 700002 | 1063491 | 460749 |
| 1981 | 367725 | 305532 | 550934 | 169439 |
| 1982 | 289547 | 257677 | 448290 | 148113 |
| 1983 | 415500 | 261980 | 476044 | 144175 |
| 1984 | 561721 | 849034 | 1269123 | 567997 |
| 1985 | 369172 | 386918 | 669788 | 223512 |
| 1986 | 403918 | 244942 | 448497 | 133772 |
| 1987 | 518289 | 663144 | 968392 | 454113 |
| 1988 | 419843 | 302767 | 539090 | 170042 |
| 1989 | 555930 | 754025 | 1108154 | 513064 |
| 1990 | 437216 | 483401 | 768915 | 303904 |
| 1991 | 364829 | 306221 | 533728 | 175691 |
| 1992 | 334427 | 389203 | 620570 | 244097 |
| 1993 | 347456 | 354761 | 580778 | 216701 |
| 1994 | 389441 | 304335 | 506306 | 182932 |
| 1995 | 386545 | 651067 | 931426 | 455096 |
| 1996 | 244667 | 226004 | 384489 | 132846 |
| 1997 | 257697 | 267584 | 417453 | 171519 |
| 1998 | 332979 | 353197 | 525372 | 237447 |
| 1999 | 387993 | 484700 | 711179 | 330344 |
| 2000 | 367725 | 419653 | 635794 | 276990 |
| 2001 | 123057 | 152910 | 261956 | 89257 |
| 2002 | 104237 | 154540 | 274572 | 86981 |
| 2003 | n/a | 142874 | 262240 | 77841 |
| 2004 | n/a | 298862 | 570400 | 156589 |

Table 2.23-Time series of EBS Pacific cod catch divided by age 3+ biomass as estimated in last year's and this year's assessments, 1977-2005. Note that last year's model used 1978 as the initial year, so an estimate of the ratio for 1977 is not available from that assessment. Also, note that the last entry in each column is based on partial catches for the respective year, because the year was/is still in progress at the time of the assessment.

| Year | Last Year | This Year |
| :---: | :---: | :---: |
| 1977 | $\mathrm{n} / \mathrm{a}$ | 0.01 |
| 1978 | 0.02 | 0.04 |
| 1979 | 0.03 | 0.05 |
| 1980 | 0.06 | 0.09 |
| 1981 | 0.05 | 0.08 |
| 1982 | 0.04 | 0.06 |
| 1983 | 0.05 | 0.06 |
| 1984 | 0.03 | 0.04 |
| 1985 | 0.02 | 0.02 |
| 1986 | 0.03 | 0.04 |
| 1987 | 0.04 | 0.05 |
| 1988 | 0.04 | 0.04 |
| 1989 | 0.05 | 0.06 |
| 1990 | 0.09 | 0.09 |
| 1991 | 0.09 | 0.10 |
| 1992 | 0.10 | 0.10 |
| 1993 | 0.07 | 0.07 |
| 1994 | 0.06 | 0.06 |
| 1995 | 0.09 | 0.09 |
| 1996 | 0.09 | 0.09 |
| 1997 | 0.11 | 0.11 |
| 1998 | 0.11 | 0.11 |
| 1999 | 0.13 | 0.13 |
| 2000 | 0.12 | 0.11 |
| 2001 | 0.09 | 0.09 |
| 2002 | 0.10 | 0.09 |
| 2003 | 0.10 | 0.09 |
| 2004 | 0.11 | 0.10 |
| 2005 | $\mathrm{n} / \mathrm{a}$ | 0.08 |

Table 2.24—Definitions of labels and terms used in the Pacific cod projection tables.

| Symbol | Definition |
| :--- | :--- |
| SPR | Equilibrium spawning per recruit, expressed as a percentage of the maximum level |
| L90\%CI | Lower bound of the 90\% confidence interval |
| Median | Point that divides projection outputs into two groups of equal size ( $50 \%$ higher, $50 \%$ lower) |
| Mean | Average value of the projection outputs |
| U90\%CI | Upper bound of the 90\% confidence interval |
| Std. Dev. | Standard deviation of the projection outputs |

Table 2.25-Equilibrium reference points and projections for GOA Pacific cod catch (t), spawning biomass (t), and fishing mortality under the assumption that $F=\max F_{A B C}$ in 2006-2018 (Scenario 1), with random variability in future recruitment. See Table 2.24 for label definitions.

| Equilibrium Reference Points |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| SPR | Catch | Spawning Bio. | Fishing Mort. |  |  |
| 100\% | 0 | 329000 | 0 |  |  |
| 40\% | 94700 | 132000 | 0.56 |  |  |
| 35\% | 102000 | 115000 | 0.69 |  |  |
| Catch Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 108743 | 108744 | 108745 | 108746 | 1 |
| 2007 | 71368 | 71373 | 71374 | 71382 | 5 |
| 2008 | 48524 | 48589 | 48601 | 48715 | 61 |
| 2009 | 51197 | 52519 | 52769 | 55106 | 1250 |
| 2010 | 60559 | 69856 | 71293 | 88187 | 8171 |
| 2011 | 65810 | 88635 | 89084 | 113474 | 14894 |
| 2012 | 68588 | 98207 | 97887 | 127946 | 18230 |
| 2013 | 69396 | 100140 | 99968 | 129513 | 19031 |
| 2014 | 67831 | 99730 | 99174 | 130496 | 19026 |
| 2015 | 67124 | 98367 | 98395 | 128964 | 18826 |
| 2016 | 69005 | 98529 | 98847 | 131983 | 18748 |
| 2017 | 69881 | 100570 | 100320 | 131571 | 18946 |
| 2018 | 72371 | 101696 | 102055 | 134784 | 19010 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 165201 | 165201 | 165201 | 165201 | 0 |
| 2007 | 122263 | 122263 | 122263 | 122263 | 0 |
| 2008 | 102197 | 102267 | 102280 | 102400 | 65 |
| 2009 | 102599 | 104184 | 104473 | 107213 | 1466 |
| 2010 | 107445 | 116049 | 117318 | 131264 | 7509 |
| 2011 | 110605 | 127895 | 130358 | 156295 | 14428 |
| 2012 | 112390 | 135767 | 138007 | 172436 | 18532 |
| 2013 | 113476 | 137943 | 140435 | 174186 | 19812 |
| 2014 | 112371 | 137924 | 140313 | 175964 | 20002 |
| 2015 | 112057 | 137030 | 140056 | 174438 | 19959 |
| 2016 | 113820 | 137926 | 140788 | 178741 | 20187 |
| 2017 | 114505 | 140028 | 142496 | 179080 | 20595 |
| 2018 | 116384 | 141547 | 144216 | 182619 | 20875 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.558 | 0.558 | 0.558 | 0.558 | 0.000 |
| 2007 | 0.517 | 0.517 | 0.517 | 0.517 | 0.000 |
| 2008 | 0.427 | 0.427 | 0.427 | 0.428 | 0.000 |
| 2009 | 0.429 | 0.436 | 0.437 | 0.449 | 0.007 |
| 2010 | 0.450 | 0.489 | 0.493 | 0.557 | 0.030 |
| 2011 | 0.465 | 0.542 | 0.530 | 0.558 | 0.032 |
| 2012 | 0.473 | 0.558 | 0.540 | 0.558 | 0.029 |
| 2013 | 0.477 | 0.558 | 0.541 | 0.558 | 0.029 |
| 2014 | 0.472 | 0.558 | 0.541 | 0.558 | 0.030 |
| 2015 | 0.471 | 0.558 | 0.541 | 0.558 | 0.030 |
| 2016 | 0.479 | 0.558 | 0.542 | 0.558 | 0.028 |
| 2017 | 0.482 | 0.558 | 0.543 | 0.558 | 0.027 |
| 2018 | 0.490 | 0.558 | 0.545 | 0.558 | 0.025 |

Table 2.26—Equilibrium reference points and projections for GOA Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=1 / 2 \max F_{A B C}$ in 2006-2018 (Scenarios 2 and 3), with random variability in future recruitment. See Table 2.24 for label definitions.

Equilibrium Reference Points

| SPR | Catch | Spawning Bio. | Fishing Mort. |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 100\% | 0 | 329000 | 0 |  |  |
| 40\% | 94700 | 132000 | 0.56 |  |  |
| 35\% | 102000 | 115000 | 0.69 |  |  |
| Catch Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 58684 | 58685 | 58685 | 58686 | 0 |
| 2007 | 47581 | 47584 | 47584 | 47589 | 2 |
| 2008 | 39394 | 39437 | 39444 | 39518 | 39 |
| 2009 | 39848 | 40687 | 40832 | 42323 | 746 |
| 2010 | 45337 | 48457 | 48833 | 53521 | 2549 |
| 2011 | 48626 | 56526 | 57565 | 69772 | 6701 |
| 2012 | 50554 | 63610 | 64633 | 82169 | 9791 |
| 2013 | 52823 | 68166 | 69003 | 87456 | 11084 |
| 2014 | 54008 | 70647 | 71451 | 90817 | 11517 |
| 2015 | 55109 | 71689 | 72817 | 93151 | 11713 |
| 2016 | 56115 | 72477 | 73658 | 94370 | 11753 |
| 2017 | 57013 | 73616 | 74382 | 94915 | 11768 |
| 2018 | 57739 | 73997 | 75004 | 95290 | 11749 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 169803 | 169803 | 169803 | 169803 | 0 |
| 2007 | 144876 | 144876 | 144876 | 144876 | 0 |
| 2008 | 130035 | 130106 | 130119 | 130242 | 66 |
| 2009 | 129221 | 130831 | 131127 | 133910 | 1494 |
| 2010 | 133463 | 142539 | 143884 | 158707 | 7939 |
| 2011 | 137845 | 158132 | 160810 | 190719 | 16640 |
| 2012 | 143396 | 173884 | 176034 | 216046 | 22995 |
| 2013 | 149305 | 185550 | 187423 | 231705 | 26340 |
| 2014 | 153360 | 193091 | 195407 | 243285 | 28099 |
| 2015 | 157490 | 198964 | 200885 | 250165 | 28953 |
| 2016 | 160196 | 202707 | 204827 | 256500 | 29289 |
| 2017 | 163794 | 206375 | 207992 | 258593 | 29428 |
| 2018 | 167420 | 208520 | 210281 | 261956 | 29317 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2007 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2008 | 0.269 | 0.269 | 0.269 | 0.269 | 0.000 |
| 2009 | 0.268 | 0.271 | 0.272 | 0.278 | 0.003 |
| 2010 | 0.277 | 0.279 | 0.279 | 0.279 | 0.001 |
| 2011 | 0.279 | 0.279 | 0.279 | 0.279 | 0.001 |
| 2012 | 0.279 | 0.279 | 0.279 | 0.279 | 0.001 |
| 2013 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2014 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2015 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2016 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2017 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |
| 2018 | 0.279 | 0.279 | 0.279 | 0.279 | 0.000 |

Table 2.27—Equilibrium reference points and projections for GOA Pacific cod catch (t), spawning biomass (t), and fishing mortality under the assumption that $F=$ the 2001-2005 average in 2006-2018 (Scenario 4), with random variability in future recruitment. See Table 2.24 for label definitions.

| Equilibrium Reference Points |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| SPR | Catch | Spawning Bio. | Fishing Mort. |  |  |
| 100\% | 0 | 329000 | 0 |  |  |
| 40\% | 94700 | 132000 | 0.56 |  |  |
| 35\% | 102000 | 115000 | 0.69 |  |  |
| Catch Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 40960 | 40961 | 40961 | 40961 | 0 |
| 2007 | 34829 | 34831 | 34831 | 34835 | 2 |
| 2008 | 30775 | 30790 | 30793 | 30819 | 14 |
| 2009 | 31290 | 31539 | 31583 | 32015 | 231 |
| 2010 | 34371 | 36310 | 36605 | 39867 | 1732 |
| 2011 | 36887 | 42428 | 43179 | 51775 | 4703 |
| 2012 | 38618 | 48089 | 48838 | 61540 | 7061 |
| 2013 | 40677 | 52137 | 52713 | 66799 | 8184 |
| 2014 | 42096 | 54579 | 55199 | 70027 | 8654 |
| 2015 | 43372 | 55928 | 56756 | 72206 | 8892 |
| 2016 | 44353 | 56820 | 57676 | 72969 | 8948 |
| 2017 | 44905 | 57642 | 58237 | 73833 | 8926 |
| 2018 | 45686 | 57818 | 58522 | 74061 | 8853 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 171301 | 171301 | 171301 | 171301 | 0 |
| 2007 | 153289 | 153289 | 153289 | 153289 | 0 |
| 2008 | 142332 | 142405 | 142418 | 142544 | 68 |
| 2009 | 142964 | 144620 | 144924 | 147789 | 1534 |
| 2010 | 148236 | 157551 | 158935 | 174173 | 8137 |
| 2011 | 154361 | 175324 | 177935 | 208737 | 17074 |
| 2012 | 161835 | 193696 | 196033 | 238230 | 24030 |
| 2013 | 169820 | 209032 | 210909 | 260303 | 28175 |
| 2014 | 176266 | 219790 | 222455 | 275476 | 30666 |
| 2015 | 182343 | 228694 | 230966 | 286536 | 32031 |
| 2016 | 187110 | 235035 | 237019 | 292857 | 32582 |
| 2017 | 192447 | 239872 | 241291 | 297039 | 32687 |
| 2018 | 196441 | 241747 | 243829 | 300419 | 32390 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2007 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2008 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2009 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2010 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2011 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2012 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2013 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2014 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2015 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2016 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2017 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |
| 2018 | 0.190 | 0.190 | 0.190 | 0.190 | 0.000 |

Table 2.28—Equilibrium reference points and projections for GOA Pacific cod catch (t), spawning biomass (t), and fishing mortality under the assumption that $F=0$ in 2006-2018 (Scenario 5), with random variability in future recruitment. See Table 2.24 for label definitions.

| Equilibrium Reference Points <br> SPR | Catch | Spawning Bio. | Fishing Mort. |
| :--- | ---: | ---: | ---: | ---: | ---: |


| Spawning Biomass Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 174543 | 174543 | 174543 | 174543 | 0 |
| 2007 | 173164 | 173164 | 173164 | 173164 | 0 |
| 2008 | 174248 | 174321 | 174335 | 174460 | 68 |
| 2009 | 182792 | 184450 | 184755 | 187625 | 1536 |
| 2010 | 194103 | 203495 | 204897 | 220288 | 8215 |
| 2011 | 206113 | 227781 | 230610 | 262910 | 17732 |
| 2012 | 219463 | 254573 | 257162 | 304385 | 26228 |
| 2013 | 233504 | 279652 | 282005 | 338110 | 32565 |
| 2014 | 247769 | 299874 | 303842 | 368639 | 37210 |
| 2015 | 260745 | 318668 | 321510 | 391664 | 40232 |
| 2016 | 270858 | 331744 | 334402 | 404547 | 41671 |
| 2017 | 278755 | 341014 | 342556 | 415669 | 41887 |
| 2018 | 283928 | 344470 | 345966 | 416588 | 41107 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2014 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2015 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 2.29—Equilibrium reference points and projections for GOA Pacific cod catch (t), spawning biomass ( t ), and fishing mortality under the assumption that $F=F_{\text {OFL }}$ in 2006-2018 (Scenario 6), with random variability in future recruitment. See Table 2.24 for label definitions.

| Equilibrium Reference Points |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| SPR | Catch | Spawning Bio. | Fishing Mort. |  |  |
| 100\% | 0 | 329000 | 0 |  |  |
| 40\% | 94700 | 132000 | 0.56 |  |  |
| 35\% | 102000 | 115000 | 0.69 |  |  |
| Catch Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 130051 | 130052 | 130053 | 130054 | 1 |
| 2007 | 74777 | 74783 | 74784 | 74793 | 5 |
| 2008 | 49901 | 49974 | 49987 | 50115 | 68 |
| 2009 | 54261 | 55763 | 56047 | 58703 | 1420 |
| 2010 | 65168 | 75718 | 77538 | 96615 | 9799 |
| 2011 | 70241 | 95705 | 98244 | 130574 | 19061 |
| 2012 | 72135 | 106419 | 106936 | 145123 | 22475 |
| 2013 | 71687 | 105653 | 106440 | 142026 | 22817 |
| 2014 | 68906 | 101727 | 103094 | 141105 | 22354 |
| 2015 | 67808 | 98999 | 101344 | 137232 | 22041 |
| 2016 | 69339 | 100432 | 102287 | 140745 | 22088 |
| 2017 | 70421 | 104311 | 104853 | 139948 | 22469 |
| 2018 | 72583 | 106912 | 107123 | 144996 | 22630 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 163047 | 163047 | 163047 | 163047 | 0 |
| 2007 | 113345 | 113345 | 113345 | 113345 | 0 |
| 2008 | 93517 | 93587 | 93599 | 93719 | 65 |
| 2009 | 94941 | 96518 | 96805 | 99530 | 1458 |
| 2010 | 100118 | 108625 | 109866 | 123642 | 7390 |
| 2011 | 102780 | 119679 | 121853 | 147374 | 13705 |
| 2012 | 103684 | 125581 | 127110 | 157879 | 16604 |
| 2013 | 104090 | 125143 | 126803 | 156041 | 16819 |
| 2014 | 101840 | 122500 | 124801 | 155428 | 16363 |
| 2015 | 101596 | 121588 | 123884 | 152355 | 16088 |
| 2016 | 102207 | 122660 | 124680 | 155384 | 16369 |
| 2017 | 103682 | 124258 | 126485 | 155951 | 16793 |
| 2018 | 105004 | 125822 | 127954 | 159346 | 17020 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.692 | 0.692 | 0.692 | 0.692 | 0.000 |
| 2007 | 0.591 | 0.591 | 0.591 | 0.591 | 0.000 |
| 2008 | 0.481 | 0.482 | 0.482 | 0.482 | 0.000 |
| 2009 | 0.489 | 0.498 | 0.499 | 0.514 | 0.008 |
| 2010 | 0.518 | 0.565 | 0.571 | 0.648 | 0.039 |
| 2011 | 0.532 | 0.626 | 0.625 | 0.692 | 0.053 |
| 2012 | 0.537 | 0.659 | 0.641 | 0.692 | 0.053 |
| 2013 | 0.540 | 0.656 | 0.639 | 0.692 | 0.055 |
| 2014 | 0.527 | 0.642 | 0.633 | 0.692 | 0.057 |
| 2015 | 0.526 | 0.637 | 0.630 | 0.692 | 0.057 |
| 2016 | 0.529 | 0.642 | 0.633 | 0.692 | 0.055 |
| 2017 | 0.537 | 0.651 | 0.638 | 0.692 | 0.055 |
| 2018 | 0.545 | 0.660 | 0.643 | 0.692 | 0.053 |

Table 2.30-Equilibrium reference points and projections for GOA Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in each year 2006-2007 and $F$ $=F_{\text {OFL }}$ thereafter (Scenario 7), with random variability in future recruitment. See Table 2.24 for label definitions.

| Equilibrium Reference Points |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| SPR | Catch | Spawning Bio. | Fishing Mort. |  |  |
| 100\% | 0 | 329000 | 0 |  |  |
| 40\% | 94700 | 132000 | 0.56 |  |  |
| 35\% | 102000 | 115000 | 0.69 |  |  |
| Catch Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 108743 | 108744 | 108745 | 108746 | 1 |
| 2007 | 71368 | 71373 | 71374 | 71382 | 5 |
| 2008 | 58243 | 58322 | 58336 | 58473 | 73 |
| 2009 | 57638 | 59176 | 59467 | 62187 | 1455 |
| 2010 | 66340 | 76953 | 78768 | 97966 | 9793 |
| 2011 | 70539 | 96054 | 98484 | 130483 | 18958 |
| 2012 | 72801 | 107221 | 107471 | 145475 | 22424 |
| 2013 | 72944 | 107689 | 108060 | 143465 | 22903 |
| 2014 | 70621 | 104665 | 105547 | 144076 | 22662 |
| 2015 | 69417 | 101819 | 103593 | 139796 | 22354 |
| 2016 | 70509 | 101853 | 103740 | 142315 | 22259 |
| 2017 | 71085 | 104990 | 105582 | 140792 | 22534 |
| 2018 | 72940 | 107336 | 107514 | 145592 | 22661 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 165201 | 165201 | 165201 | 165201 | 0 |
| 2007 | 122263 | 122263 | 122263 | 122263 | 0 |
| 2008 | 101351 | 101421 | 101433 | 101553 | 64 |
| 2009 | 98366 | 99939 | 100226 | 102945 | 1455 |
| 2010 | 101481 | 109972 | 111213 | 124962 | 7380 |
| 2011 | 103289 | 120191 | 122378 | 147927 | 13711 |
| 2012 | 104221 | 126215 | 127793 | 158676 | 16715 |
| 2013 | 104971 | 126584 | 128188 | 158426 | 17160 |
| 2014 | 103071 | 124105 | 126630 | 158537 | 16911 |
| 2015 | 102584 | 122946 | 125492 | 154677 | 16590 |
| 2016 | 102844 | 123514 | 125733 | 157289 | 16691 |
| 2017 | 104061 | 124803 | 127069 | 157131 | 16974 |
| 2018 | 105230 | 126180 | 128306 | 159965 | 17128 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2006 | 0.558 | 0.558 | 0.558 | 0.558 | 0.000 |
| 2007 | 0.517 | 0.517 | 0.517 | 0.517 | 0.000 |
| 2008 | 0.524 | 0.525 | 0.525 | 0.526 | 0.000 |
| 2009 | 0.508 | 0.517 | 0.518 | 0.533 | 0.008 |
| 2010 | 0.525 | 0.572 | 0.579 | 0.655 | 0.039 |
| 2011 | 0.535 | 0.629 | 0.628 | 0.692 | 0.052 |
| 2012 | 0.540 | 0.662 | 0.644 | 0.692 | 0.052 |
| 2013 | 0.545 | 0.664 | 0.644 | 0.692 | 0.053 |
| 2014 | 0.534 | 0.650 | 0.639 | 0.692 | 0.055 |
| 2015 | 0.531 | 0.644 | 0.635 | 0.692 | 0.055 |
| 2016 | 0.533 | 0.647 | 0.636 | 0.692 | 0.054 |
| 2017 | 0.539 | 0.654 | 0.640 | 0.692 | 0.054 |
| 2018 | 0.546 | 0.662 | 0.644 | 0.692 | 0.053 |

Table 2.31a—Bycatch of nontarget and "other" species taken in the GOA Pacific cod trawl fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the GOA Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year.

|  | Bycatch in GOA Pacific cod trawl fishery |  |  |  |  |  | Proportion of total GOA catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 201 | 109 | 127 | 124 | 69 | 75 | 0.22 | 0.20 | 0.23 | 0.13 | 0.12 | 0.08 |
| Skates | 476 | 411 | 385 | 219 | 272 | 120 | 0.15 | 0.09 | 0.19 | 0.07 | 0.15 | 0.02 |
| Shark | 11 | 7 | 4 | 1 | 1 | 0 | 0.09 | 0.00 | 0.12 | 0.02 | 0.01 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Dogfish | 30 | 624 | 14 | 21 | 61 | 3 | 0.05 | 0.72 | 0.04 | 0.05 | 0.12 | 0.02 |
| Sleepershk | 17 | 6 | 5 | 11 | 0 | 26 | 0.12 | 0.07 | 0.01 | 0.02 | 0.00 | 0.12 |
| Octopus | 25 | 1 | 4 | 0 | 3 | 7 | 0.11 | 0.01 | 0.03 | 0.00 | 0.03 | 0.02 |
| Squid | 1 | 1 | 1 | 0 | 1 | 0 | 0.01 | 0.01 | 0.03 | 0.01 | 0.01 | 0.00 |
| Smelts | 0 | 1 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 |  | 0 |  | 0.00 | 0.00 | 0.00 |  | 1.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.56 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 |  | 0 | 0 | 0.00 |  | 0.00 |  | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 1.00 | 1.00 | 0.97 | 0.12 | 1.00 |
| Grenadier | 0 | 1 | 17 | 114 | 376 | 0 | 0.00 | 0.00 | 0.00 | 0.01 | 0.04 | 0.00 |
| Otherfish | 58 | 211 | 110 | 43 | 68 | 42 | 0.10 | 0.03 | 0.13 | 0.04 | 0.10 | 0.02 |
| Crabs | 1 | 12 | 1 | 0 | 0 | 0 | 0.08 | 0.47 | 0.06 | 0.03 | 0.06 | 0.04 |
| Starfish | 63 | 59 | 62 | 22 | 27 | 22 | 0.06 | 0.05 | 0.04 | 0.02 | 0.06 | 0.04 |
| Jellyfish | 7 | 5 | 1 | 1 | 13 | 1 | 0.18 | 0.03 | 0.01 | 0.02 | 0.05 | 0.00 |
| Invertunid | 2 | 28 | 0 | 5 | 1 | 0 | 0.22 | 0.65 | 0.10 | 0.31 | 0.13 | 0.00 |
| seapen/whip | 0 | 0 | 3 | 0 | 0 | 0 | 0.00 | 0.01 | 0.99 | 0.00 | 0.00 | 0.00 |
| Sponge | 0 | 1 | 1 | 1 | 1 | 0 | 0.04 | 0.24 | 0.10 | 0.12 | 0.26 | 0.09 |
| Anemone | 3 | 3 | 11 | 1 | 3 | 6 | 0.17 | 0.20 | 0.65 | 0.07 | 0.21 | 0.27 |
| Tunicate | 1 | 0 | 0 | 0 | 0 | 0 | 0.43 | 0.13 | 0.38 | 0.05 | 0.04 | 0.03 |
| Benthinv | 3 | 22 | 11 | 1 | 1 | 0 | 0.11 | 0.72 | 0.42 | 0.07 | 0.06 | 0.09 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 3 | 23 | 2 | 2 | 1 | 2 | 0.13 | 0.72 | 0.24 | 0.31 | 0.12 | 0.26 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 | 0.01 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.08 | 0.02 | 0.01 | 0.03 | 0.01 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.07 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 2.31b—Bycatch of nontarget and "other" species taken in the GOA Pacific cod trawl fishery, 20032005. The first part of the table ("Bycatch") shows the amount (t) of each species group taken as bycatch in the GOA Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.


Table 2.32a-Bycatch of nontarget and "other" species taken in the GOA Pacific cod longline fishery, 1997-2002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the GOA Pacific cod longline fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year.

|  | Bycatch in GOA Pacific cod longline fishery |  |  |  |  |  | Proportion of total GOA catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 63 | 181 | 207 | 203 | 197 | 291 | 0.07 | 0.33 | 0.38 | 0.22 | 0.33 | 0.31 |
| Skates | 478 | 461 | 789 | 1823 | 617 | 5005 | 0.15 | 0.10 | 0.39 | 0.56 | 0.34 | 0.77 |
| Shark | 2 | 4 | 8 | 2 | 1 | 5 | 0.02 | 0.00 | 0.25 | 0.03 | 0.01 | 0.19 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Dogfish | 28 | 104 | 146 | 8 | 111 | 7 | 0.04 | 0.12 | 0.47 | 0.02 | 0.23 | 0.06 |
| Sleepershk | 42 | 14 | 501 | 366 | 66 | 40 | 0.31 | 0.19 | 0.90 | 0.60 | 0.26 | 0.18 |
| Octopus | 1 | 25 | 17 | 16 | 6 | 7 | 0.00 | 0.22 | 0.10 | 0.09 | 0.07 | 0.02 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 |  | 0 |  | 0.00 | 0.00 | 0.00 |  | 0.00 |  |
| Sticheidae | 0 | 0 | 4 | 0 | 0 | 0 | 0.00 | 0.00 | 1.00 | 0.00 | 0.01 | 0.00 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 |  | 0 | 0 | 0.00 |  | 0.00 |  | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 191 | 0 | 423 | 0 | 0 | 92 | 0.02 | 0.00 | 0.04 | 0.00 | 0.00 | 0.01 |
| Otherfish | 15 | 50 | 36 | 39 | 2 | 128 | 0.03 | 0.01 | 0.04 | 0.04 | 0.00 | 0.06 |
| Crabs | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Starfish | 304 | 162 | 765 | 199 | 347 | 207 | 0.31 | 0.13 | 0.51 | 0.22 | 0.74 | 0.40 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 5 | 0 | 4 | 0.05 | 0.00 | 0.17 | 0.34 | 0.05 | 0.32 |
| seapen/whip | 0 | 3 | 0 | 1 | 0 | 0 | 0.00 | 0.99 | 0.00 | 0.87 | 0.00 | 0.07 |
| Sponge | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.01 | 0.01 | 0.01 |
| Anemone | 0 | 8 | 5 | 5 | 0 | 1 | 0.02 | 0.52 | 0.27 | 0.33 | 0.02 | 0.06 |
| Tunicate | 0 | 0 | 0 | 1 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.17 | 0.00 | 0.00 |
| Benthinv | 0 | 1 | 1 | 1 | 5 | 0 | 0.00 | 0.03 | 0.03 | 0.07 | 0.40 | 0.07 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.05 | 0.00 | 0.00 | 0.02 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 1 | 1 | 1 | 1 | 0 | 0.13 | 0.12 | 0.16 | 0.21 | 0.43 | 0.40 |

Table 2.32b—Bycatch of nontarget and "other" species taken in the GOA Pacific cod hook-and-line (including jigs) fishery, 2003-2005. The first part of the table ("Bycatch") shows the amount (t) of each species group taken as bycatch in the GOA Pacific cod hook-and-line fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 19972002.

| Species group | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata |  |  |  |  |  |  |
| Birds | 0 | 0 |  | 0.01 | 0.03 |  |
| Bivalves | 0 | 0 | 0 | 0.11 | 0.00 | 0.02 |
| Brittle star unidentified |  | 0 |  |  | 0.30 |  |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans |  |  | 0 |  |  | 0.00 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 | 0 |  | 0.00 | 0.00 |  |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 1 | 1 | 0.05 | 0.06 | 0.16 |
| Grenadier |  | 0 |  |  | 0.00 |  |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified |  |  |  |  |  |  |
| Invertebrate unidentified | 0 | 2 |  | 0.00 | 0.27 |  |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 39 | 129 | 49 | 0.33 | 0.20 | 0.09 |
| Misc crabs | 0 | 0 | 0 | 0.00 | 0.02 | 0.01 |
| Misc crustaceans |  |  |  |  |  |  |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 11 | 6 | 2 | 0.03 | 0.02 | 0.01 |
| Misc inverts (worms etc) |  |  |  |  |  |  |
| Octopus | 2 | 1 | 0 | 0.05 | 0.01 | 0.00 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 90 | 7 | 7 | 0.17 | 0.14 | 0.15 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified |  |  |  |  |  |  |
| Scypho jellies |  |  |  |  |  |  |
| Sea anemone unidentified | 1 | 1 | 0 | 0.06 | 0.09 | 0.02 |
| Sea pens whips | 0 |  | 0 | 0.40 |  | 0.05 |
| Sea star | 110 | 246 | 170 | 0.20 | 0.23 | 0.17 |
| Shark | 59 | 13 | 10 | 0.17 | 0.11 | 0.04 |
| Skate | 464 | 472 | 108 | 0.12 | 0.21 | 0.06 |
| Snails | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Sponge unidentified |  | 0 | 1 |  | 0.07 | 0.34 |
| Squid | 10 | 0 | 0 | 0.13 | 0.00 | 0.00 |
| Stichaeidae |  |  |  |  |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers |  | 0 |  |  | 0.00 |  |

Table 2.33a-Bycatch of nontarget and "other" species taken in the GOA Pacific cod pot fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( $t$ ) of each species group taken as bycatch in the GOA Pacific cod pot fishery, broken down by year. The second part of the table
("Proportion of...") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year.

|  | Bycatch in GOA Pacific cod pot fishery |  |  |  |  |  | Proportion of total GOA catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 106 | 61 | 106 | 357 | 29 | 79 | 0.12 | 0.11 | 0.19 | 0.38 | 0.05 | 0.09 |
| Skates | 1 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 1 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| Dogfish | 0 | 0 | 0 | 0 | 1 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Octopus | 168 | 74 | 142 | 137 | 63 | 252 | 0.72 | 0.66 | 0.85 | 0.78 | 0.71 | 0.84 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 |  | 0 |  | 0.00 | 0.00 | 0.00 |  | 0.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 |  | 0 | 0 | 0.00 |  | 0.00 |  | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 1 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 30 | 4 | 92 | 19 | 52 | 43 | 0.05 | 0.00 | 0.11 | 0.02 | 0.07 | 0.02 |
| Crabs | 6 | 10 | 9 | 10 | 2 | 1 | 0.41 | 0.42 | 0.81 | 0.84 | 0.36 | 0.19 |
| Starfish | 468 | 210 | 633 | 566 | 35 | 66 | 0.47 | 0.17 | 0.42 | 0.63 | 0.08 | 0.13 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.03 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sponge | 0 | 0 | 5 | 0 | 0 | 0 | 0.03 | 0.00 | 0.39 | 0.04 | 0.01 | 0.01 |
| Anemone | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.03 | 0.41 | 0.02 | 0.00 | 0.00 |
| Benthinv | 10 | 2 | 10 | 4 | 1 | 2 | 0.40 | 0.08 | 0.40 | 0.34 | 0.08 | 0.28 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 1 | 0 | 1 | 1 | 1 | 1 | 0.06 | 0.00 | 0.09 | 0.14 | 0.16 | 0.09 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.02 | 10 |

Table 2.33b—Bycatch of nontarget and "other" species taken in the GOA Pacific cod pot fishery, 20032005. The first part of the table ("Bycatch") shows the amount (t) of each species group taken as bycatch in the GOA Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total GOA catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the GOA during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

| Species group | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 |  |  | 0.01 |  |  |
| Birds | 0 | 0 | 0 | 0.02 | 0.09 | 0.08 |
| Bivalves | 0 | 0 | 0 | 0.14 | 0.00 | 0.01 |
| Brittle star unidentified | 0 | 0 | 0 | 0.03 | 0.65 | 0.53 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 0 | 0 |  | 0.00 | 0.01 |  |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 |  | 7 | 0.13 |  | 0.34 |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 1 | 0 | 0.10 | 0.04 | 0.04 |
| Grenadier |  |  |  |  |  |  |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 0 | 0 | 0 | 0.05 | 0.08 | 0.45 |
| Invertebrate unidentified | 0 |  |  | 0.02 |  |  |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 14 | 262 | 157 | 0.11 | 0.41 | 0.28 |
| Misc crabs | 1 | 0 | 2 | 0.44 | 0.23 | 0.54 |
| Misc crustaceans |  |  |  |  |  |  |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 43 | 20 | 80 | 0.10 | 0.07 | 0.26 |
| Misc inverts (worms etc) |  |  |  |  |  |  |
| Octopus | 42 | 135 | 88 | 0.88 | 0.86 | 0.96 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 195 | 7 | 8 | 0.38 | 0.15 | 0.18 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified |  |  |  |  |  |  |
| Scypho jellies | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 |
| Sea anemone unidentified |  | 0 | 0 |  | 0.01 | 0.01 |
| Sea pens whips | 0 |  |  | 0.01 |  |  |
| Sea star | 341 | 756 | 748 | 0.61 | 0.71 | 0.73 |
| Shark |  |  |  |  |  |  |
| Skate | 1 | 0 | 1 | 0.00 | 0.00 | 0.00 |
| Snails | 5 | 0 | 5 | 0.56 | 0.34 | 0.68 |
| Sponge unidentified | 0 | 0 |  | 0.00 | 0.00 |  |
| Squid |  | 0 | 0 |  | 0.00 | 0.00 |
| Stichaeidae |  |  |  |  |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 0 | 0 | 0 | 0.03 | 0.09 | 0.12 |

Table 2.34—Summary of major results for the stock assessment of Pacific cod in the GOA region.

| Tier |  | 3 a |
| :--- | ---: | ---: |
| Reference mortality rates |  |  |
|  | $M$ | 0.38 |
|  | $F_{40 \%}$ | 0.56 |
| Equilibrium spawning biomass | $F_{35 \%}$ | 0.69 |
|  | $B_{35 \%}$ | $115,000 \mathrm{t}$ |
|  | $B_{40 \%}$ | $132,000 \mathrm{t}$ |
|  | $B_{100 \%}$ | $329,000 \mathrm{t}$ |
| Projected biomass for 2006 |  |  |
| Spawning (at max FABC) | $165,000 \mathrm{t}$ |  |
| Age 3+ | $453,000 \mathrm{t}$ |  |
| ABC for 2006 |  |  |
| FABC (maximum permissible) | 0.56 |  |
| FABC (recommended) | 0.28 |  |
| ABC (maximum permissible) | $109,000 \mathrm{t}$ |  |
| ABC (recommended) | $58,700 \mathrm{t}$ |  |
| Overfishing level for 2006 |  |  |
| Fishing Mortality | 0.69 |  |
| Catch | $130,000 \mathrm{t}$ |  |



Figure 2.1—Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in 2004, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.2-Mean GOA Pacific cod length at age by sex, with sex-specific $95 \%$ confidence intervals. Values were computed from age and length data collected during the 2003 bottom trawl survey. Because these data were collected during summer surveys, ages are shown at mid-year. A few fish aged 9.5-12.5 were also observed, but these are not represented here due to extremely small sample sizes.

2003 GOA Shelf Bottom Trawl Survey Data ( $n=553$ )


Figure 2.3-Mean GOA Pacific cod weight at length by sex, with sex-specific $95 \%$ confidence intervals. Values were computed from length and weight data collected during the 2003 bottom trawl survey. The full range of lengths in the sample extended from 14 cm to 98 cm . However, lengths shorter than 35 cm or longer than 75 cm are not represented here due to extremely small sample sizes. The sample size for 71 cm males was also extremely small ( 1 fish), so the confidence interval for male mean weight at 71 cm extends off the graph in both directions.


Figure 2.4—Mean GOA Pacific cod commercial fishery weight at length for six example years. Values were computed from aggregated length and weight data across all gear types and months. Because sample sizes in certain year/length combinations are very small, only those year/length combinations with at least 5 data points are shown.


Figure 2.5-Condition factor of GOA Pacific cod for six example years, with $95 \%$ confidence intervals. Condition factor is defined as the average ratio of weight to the cube of length. Values in this figure have been normalized by expressing condition factor as the ratio of the year-specific estimate to the estimate for the entire time series.


Figure 2.6-Time series of GOA Pacific cod female spawning biomass for 1978-2005 as estimated by Models 1, 2, and 3. The three points for each year have been displaced from one another slightly to improve readability. Error bars for Models 2 and 3 represent 95\% confidence intervals (SS1 does not compute confidence intervals).


Figure 2.7-Selectivity at length (cm, evaluated at midpoints of length bins) as estimated by Model 3.




Figure 2.8a—Observed and estimated size compositions from the 2003 January-May fisheries.




Figure 2.8b—Observed and estimated size compositions from the 2004 January-May fisheries.




Figure 2.8c—Observed and estimated size compositions from the 2005 January-May fisheries.


Figure 2.9—Observed and estimated size compositions from the 2001-2005 GOA bottom trawl surveys.


Figure 2.10—Selectivity at age (years) in 2005 as estimated by Model 3. Because selectivity is defined in the model as a function of length rather than age and because a range of lengths are associated with any given age, the above curves do not reach peak values of 1 .


Figure 2.11—Observed and estimated age compositions from the 2003 GOA bottom trawl survey.


Figure 2.12—Biomass time trends (age 3+ biomass, female spawning biomass, survey biomass) of GOA Pacific cod as estimated by Model 3.


Figure 2.13-Time series of GOA Pacific cod recruitment at age 0 , with $95 \%$ confidence intervals, as estimated by Model 3.


Figure 2.14—Age 0 recruitment versus female spawning biomass for GOA Pacific cod during the years $1977-2004$ as estimated by Model 3, with Ricker stock-recruitment curve fit by assuming $F_{M S Y}=F_{35 \%}$ and $B_{M S Y}=B_{35 \%}$.


Figure 2.15-Trajectory of GOA Pacific cod fishing mortality and female spawning biomass as estimated by Model 3, 1977-present.
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